

StrathE2E2 version 3.3.1: Implementation for the North Sea.

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Introduction

This document describes the configuration of StrathE2E2 for the North Sea and its parameterization to enable stationary state fitting for two time periods; 1970-1999 and 2003-2013. These represent contrasting periods of environmental conditions and fishing intensity.

Volumetric and seabed habitat data define the physical configuration of the system, and we can regard these as being fixed in time. Similarly, we regard the physiological parameters of the ecology model as being fixed in time. Some of them are set from external data while the remainder are fitted as detailed here. Changes in the model performance between the different time periods are thus due only to the hydrodynamic, hydro-chemical and fishery driving data.

North Sea model domain, physical structure, and time-independent parameters

Model domain

The perimeter of the North Sea model domain is bounded to the west and east by the UK and continental European coastlines respectively, and by open-sea boundaries to the north between Scotland and Denmark, and to the south at the English Channel. The open boundary to the north separates the shelf west of Scotland from the northern North Sea along a transect across the gap between the Orkney and Shetland Islands and then tracks approximately along the 200m isobaths at the shelf edge (Figure 1). The internal boundary between the inshore/shallow and offshore/deep zones is defined as the 30m isobath, which roughly divides the region into the shallow southern North Sea (International Council for the Exploration of the Sea (ICES) area IVc, northern boundary at latitude 53° 30'N), and the deeper northern North Sea (ICES areas IVa and IVb).

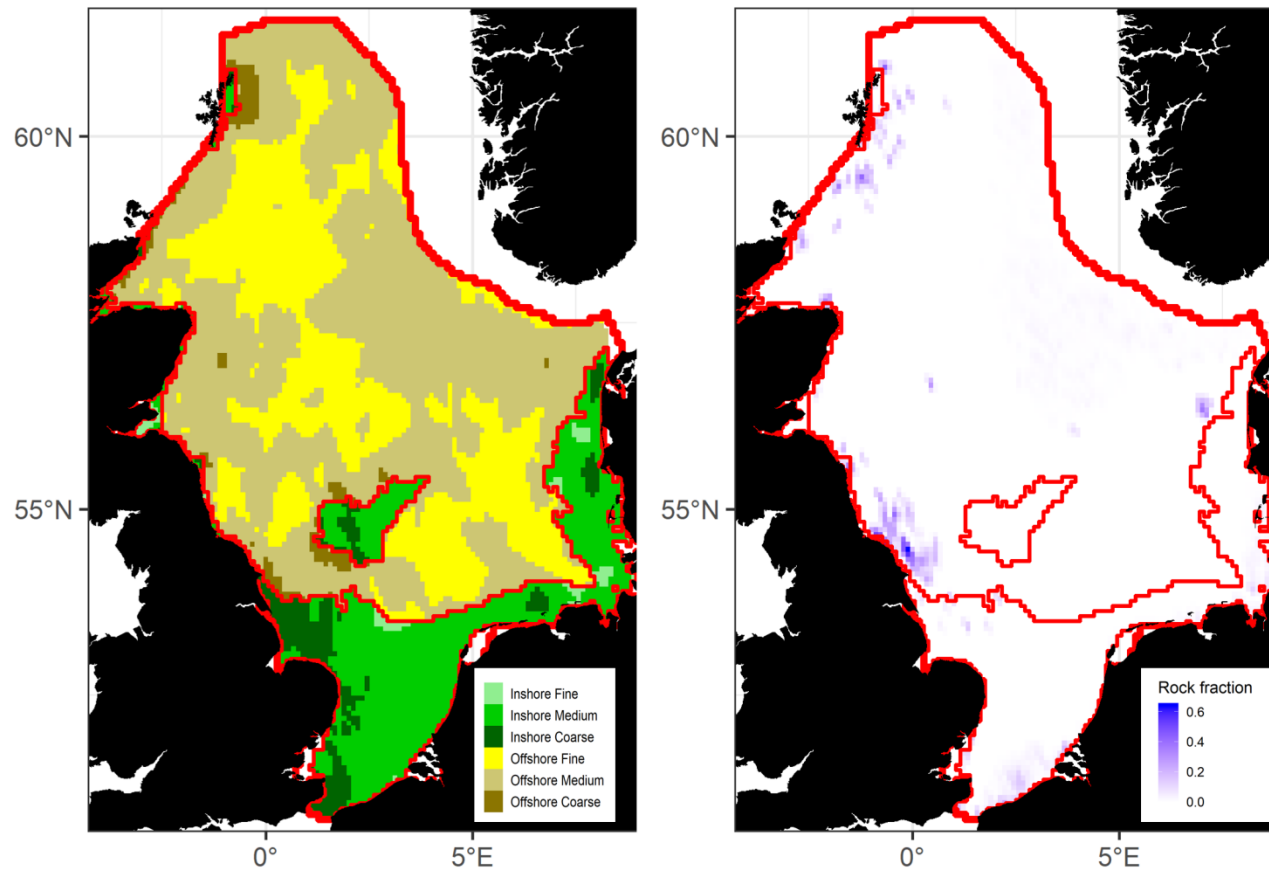


FIGURE 1 Maps of the StrathE2E2 model region. Within the North Sea the model resolves sub-area of seabed sediment habitat divided into inshore (shallower than 30m) and offshore (separated by the thin red line in each panel). Within each zone, three sediment classes are represented – fine (muddy), medium (sandy) and coarse (gravel) (left panel). Within each of the six sediment habitats a proportion of the seabed area may present as exposed bedrock (right panel) which has different geochemical properties and in the inshore zone supports the kelp forests which are included in the model food web. Sedimentary data are from Wilson *et al.* (2018). The sea surface area of the model domain was estimated to be 485,605 km².

Fixed physical configuration parameters

Background to the fixed area-proportions, volumetric and sediment property parameters of the model are shown in Tables 1 and 2.

TABLE 1. Description of the fixed (time-invariant) physical configuration parameters for the North Sea demonstration model.

| Data | Description |
|---|--|
| Water column inshore/shallow and offshore/deep zone area proportions and layer thicknesses; seabed habitat area proportions and sediment properties. | Area proportions of depth zones and seabed habitats derived from 1/8 degree resolution atlas of seabed sediment properties (Wilson <i>et al.</i> 2018). The atlas provides gridded data sets of bathymetry, median grain size, mud, sand and gravel content, porosity, permeability, organic nitrogen and carbon content, and natural disturbance rates due to wave and current bed shear stress. |
| Parameters for relationship between median grain size, sediment porosity and permeability. Permeability is used as the basis for estimating hydraulic conductivity which is a parameter in the representation of sediment processes in the model. | <p>Porosity (proportion by volume of interstitial water) and permeability of each sediment habitat were derived from median grain sizes using empirically-based relationships.</p> $\log_{10}(\text{porosity}) = p_3 + p_4 \left(\frac{1}{1 + e^{\left(\frac{-\log_{10}(D_{50}) - p_1}{p_2} \right)}} \right)$ <p>D_{50} = median grain size (mm); parameters $p_1 = -1.227$, $p_2 = -0.270$, $p_3 = -0.436$, $p_4 = 0.366$ (Heath <i>et al.</i> 2015)</p> $\text{permeability} = 10^{p_5} \cdot D_{50}^{* p_6}$ <p>where $D_{50}^* = 0.11 \leq D_{50} \leq 0.50$</p> <p>$p_5 = -9.213$, $p_6 = 4.615$ (Heath <i>et al.</i> 2015)</p> <p>These relationships are coded into the StrathE2E2 R-package with the parameters in the csv setup file for the North Sea model. The parameters are probably a reasonable starting point for any future model of a new region. Derivation of the parameters is described in the following text.</p> |
| Parameters for in-built relationship between sediment mud content, and slowly degrading (refractory) organic nitrogen content of seabed sediments (see description in this document). | <p>Values for each sediment type derived from parameterised relationships between total organic nitrogen content of sediments (<i>TON%</i>, percent by weight), mud content (<i>mud%</i>, percent by weight) and median grain size (D_{50}, mm).</p> $\text{mud}\% = 10^{p_7} \cdot D_{50}^{p_8}$ |

| | |
|--|--|
| | <p>$p_7 = 0.657, p_8 = -0.800$</p> <p>$TON\% = 10^{p_9} \cdot mud\%^{p_{10}}$</p> <p>$p_9 = -1.965, p_{10} = 0.590$</p> <p>Proportion of <i>TON</i> estimated to be refractory = 0.9</p> <p>These relationships are coded into the StrathE2E2 R-package with the parameters in the csv setup file for the North Sea model. The relationships and parameters are probably a reasonable starting point for any future model of a new region, though there are clear regional variations. Derivation of the parameters is described in the following text.</p> |
|--|--|

TABLE 2 Area-proportions of the inshore and offshore zones and the thicknesses of the water column layers. Sea surface area of the North Sea in the model domain was estimated to be 458,605 km².

| Property | Inshore/shallow | Offshore/deep |
|-----------------------------|------------------------|----------------------|
| Sea-surface area proportion | 0.2496 | 0.7504 |
| Upper layer thickness (m) | 24.16 | 30 |
| Lower layer thickness (m) | NA | 50.04 |

Area-proportions of seabed-habitats

Derivation of the area-proportions of seabed habitat in the inshore and offshore zones of the model domain relied on the atlas of seabed sediment properties from Wilson *et al.* (2018). The atlas provides a range of seabed data for 1/8 degree cells over the NW European shelf, including the percentage of seabed area defined as rock (within 5cm of the seafloor), the percentage of mud, sand and gravel fractions in the sediments, the whole-sediment median grain size, and the natural disturbance rate by currents and waves. Within each zone, we ranked the spatial cells by median grain size, and assembled cumulative area-proportion curves (Figure 3). Cells were then assigned to fine, medium and coarse sediment habitats according to the 15th and 50th centiles of these curves. The actual area of each habitat was then the sum of the areas of each set of assigned cells, less the proportion of area in these cells defined as rock (Table 3).

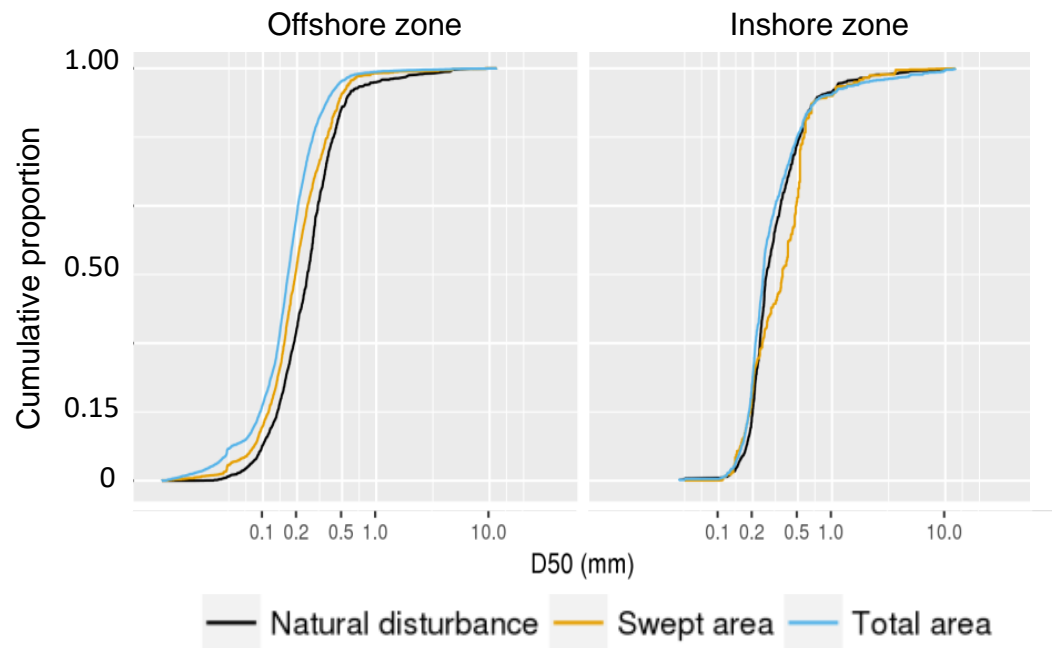


FIGURE 2 Cumulative proportion of area by whole-sediment median grain size (D50), for the inshore and offshore zones of the North Sea. Also shown is the cumulative proportion of natural disturbance due to currents and waves, and the cumulative proportion of seabed area swept by fishing gears (see later description), according to grain size.

TABLE 3 Area proportions of the 8 seabed habitat classes defined in the model by depth, rock or sediment type. Grey shaded cells indicate habitats in water deeper than 30m. The sea surface area of the model domain was estimated to be 485,605 km².

| Sediment type and depth zone | Inshore/shallow rock | Inshore/shallow muddy sediments | Inshore/shallow sandy sediments | Inshore/shallow gravels | Offshore/deep rock | Offshore/deep muddy sediments | Offshore/deep sandy sediments | Offshore/deep gravels |
|---|-----------------------------|--|--|--------------------------------|---------------------------|--------------------------------------|--------------------------------------|------------------------------|
| Area proportions | 0.0030 | 0.0110 | 0.1878 | 0.0478 | 0.0057 | 0.2665 | 0.4595 | 0.0187 |
| Median grain size (mm) | NA | 0.130 | 0.273 | 1.816 | NA | 0.114 | 0.231 | 0.928 |
| Porosity | NA | 0.441 | 0.391 | 0.367 | NA | 0.456 | 0.399 | 0.370 |
| Permeability (m ²) | NA | 5.00 x 10 ⁻¹⁴ | 1.53 x 10 ⁻¹² | 2.50 x 10 ⁻¹¹ | NA | 2.68 x 10 ⁻¹⁴ | 7.05 x 10 ⁻¹³ | 2.50 x 10 ⁻¹¹ |
| Refractory organic nitrogen content (%g.g ⁻¹) | NA | 0.081 | 0.057 | 0.023 | NA | 0.064 | 0.046 | 0.024 |

Parameters linking median grain size to porosity, permeability and organic nitrogen content

Sediment porosity

Various authors have presented data on sediment porosity and grain size: Ruardij & Van Raaphorst (1995) and Lohse *et al.* (1993) for muds and sands from the southern North Sea; Serpetti (2012) and Serpetti *et al.* (2012) for coarse, mixed and fine grained sediments at 8 sites off the northeast coast of Scotland, repeated at monthly interval over an annual cycle. Wiesner *et al.* (1990), list data on grain size and water content (by weight) for a wide range of North Sea sediments. Water content can be converted to porosity assuming a solid material density of 2650 kg.m⁻³ and a fluid density of 1027 kg.m⁻³. Combining these data sets, log-transformed porosity showed a sigmoidal relationship with log₁₀(median grain size) (D, mm), to which we fitted a relationship of the logistic form using Nelder Mead optimization in the 'optim' package of R (Table 4, Figure 3):

$$\log_{10}(\text{porosity}) = p_3 + p_4 \left(\frac{1}{1 + e^{\left(\frac{-\log_{10}(D) - p_1}{p_2} \right)}} \right) \quad \text{eqn 1}$$

TABLE 4 Fitted values and their standard error, of the four parameters for the function relating sediment porosity to median grain size.

| Parameter | Fitted value | Standard error |
|----------------|--------------|----------------|
| p ₁ | -1.227 | 0.063 |
| p ₂ | -0.270 | 0.046 |
| p ₃ | -0.436 | 0.023 |
| p ₄ | 0.366 | 0.050 |

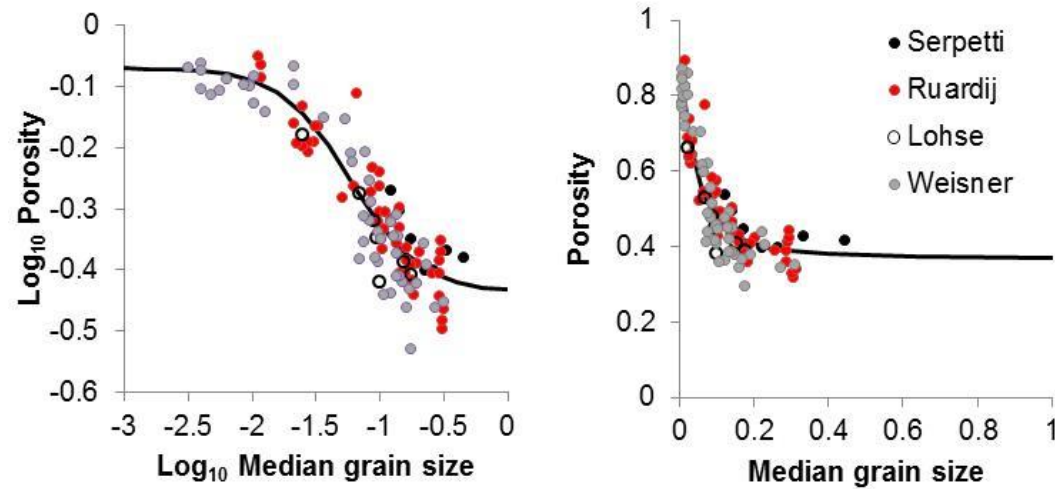


FIGURE 3 Assembled data on sediment porosity and median grain size, and the fitted relationship (solid line). Left panel, log-transformed data, right panel un-transformed data. black symbols: annual averaged data from Serpetti (2012) and Serpetti *et al.* (2012); red: Ruardij & van Raaphorst (1995); open: Lohse *et al.* (1993); grey: Weisner *et al.* (1990).

Hydraulic conductivity

Hydraulic conductivity (H , $\text{m}\cdot\text{s}^{-1}$) represents the ease with which fluids flow through the particle grain matrix. The related term ‘permeability’ (m^2) is a measure of the connectedness of the fluid filled void spaces between the particle grains. Permeability is a function only of the sediment matrix, whilst conductivity is a function of both the sediment and the permeating fluid, in particular the fluid viscosity and density.

Hydraulic conductivity is related to permeability by:

$$H = \text{Permeability} \cdot \text{fluid density} \cdot \frac{g}{\text{dynamic viscosity}} \quad \text{eqn 2}$$

where: seawater density = $1027 \text{ kg}\cdot\text{m}^{-3}$ at salinity 35 and temperature 10°C ; seawater dynamic viscosity = $1.48 \times 10^{-3} \text{ kg}\cdot\text{m}^{-1}\cdot\text{s}^{-1}$ at salinity 35 and temperature 10°C ; g = acceleration due to gravity = $9.8 \text{ m}\cdot\text{s}^{-2}$

$$\text{Hence, } H = \text{Permeability} \cdot 6.8004 \times 10^6 \text{ (m}\cdot\text{s}^{-1} \text{ at salinity 35 and temperature } 10^\circ\text{C)} \quad \text{eqn 3}$$

One of the few available datasets on whole sediment permeability in relation to median grain size is that of Serpetti (2012) and Serpetti *et al.* (2012). These data cover muddy-sand, sand and mixed sediments in the median grain size range 0.11 to 0.45 mm median grain size, sampled approximately monthly over an annual cycle at 7 sites off the east coast of Scotland. Permeability and median grain size were measured on cores from the upper 5cm and upper 10cm of the seabed at each site. A power function of median grain size (D , mm) was found to explain the differences in annual average permeability (m^{-2}) between sites ($r^2 = 0.999$ for 10cm cores, $r^2 = 0.966$ for 5 cm cores) (Figure 4):

$$\text{Permeability} = 10^{-8.675} \cdot D^{4.958} \text{ (5 cm cores)} \quad \text{eqn 4}$$

$$\text{Permeability} = 10^{-9.213} \cdot D^{4.615} \text{ (10 cm cores)} \quad \text{eqn 5}$$

The relationship for 10cm cores was used in the model configuration to parameterise permeability given median grain size.

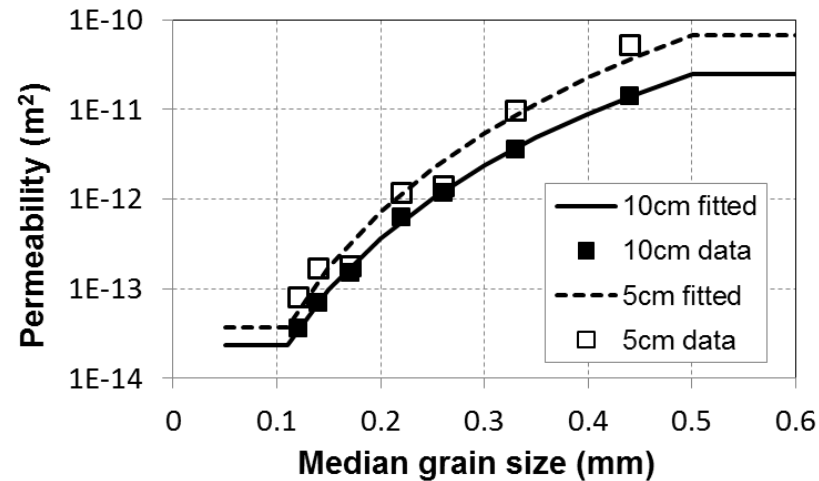


FIGURE 4 Annual average permeability (m^{-2}) of sediments from 7 sites off the north east coast of Scotland - data from Serpetti (2012) and Serpetti *et al.* (2012). Open symbols, permeability over the upper 5cm of sediment, filled symbols over the upper 10cm. We did not extrapolate the fitted permeability relationships to estimate values outside the median grain size range of the observations, but instead assumed that permeability was independent of grain size below 0.11mm and above 0.5mm. This constraint was based on unpublished field observations (M.Pace, pers. comm).

Refractory (non-dynamic) organic nitrogen content of sediments

The magnitude of the static organic nitrogen detritus pool in each sediment type is a required input to the model. The code includes an option to impute values from empirical relationships between total organic nitrogen (TON) and mud content, and between mud content and median grain size.

Comparison of sediment pigment content and total organic nitrogen (TON) content of sediments off northeast Scotland (Serpetti, 2012; Serpetti *et al.*, 2012) suggests that 90% may be refractory (assuming nitrogen:pigment ratios of labile material). This is borne out by the observation that although phytoplankton pigment contents of shelf sediments show strong seasonality, total organic nitrogen content is almost constant. Total organic nitrogen content is typically expressed as a dry-weight specific percentage ratio (TON%; percentage by weight, %g.g⁻¹). Using such data, the organic nitrogen mass (ONM, mMN) in each sediment habitat can be calculated depending on porosity and sediment layer area and thickness.

$$ONM = \frac{1}{14} (TON\% \cdot 10^4 \cdot \rho_{sediment} \cdot area \cdot thickness \cdot (1 - porosity)) \quad \text{eqn 6}$$

where $\rho_{sediment}$ is the density of particle grains in the sediment (quartz density = 2650 kg.m⁻³), and the sediment layer area and thickness have units of m² and m respectively.

Empirical evidence shows that sediment TON% is strongly related to the mud content (% grain sizes <0.063mm). However, relating TON% to whole-sediment median grain size (which is the sediment-defining measure for imputing porosity and permeability) is more problematic since mixed and coarse grained sediments may have highly variable mud content, more or less independent of median grain size.

To derive parameters linking TON% and mud% content, we combined two sets of observations from the North Sea (Serpetti, 2012; Serpetti *et al.*, 2012; Stephens & Diesing, 2015; data for the latter being downloaded from the Cefas Data Hub (<http://data.cefas.co.uk/#/Search/1/sediment>). Together these sets provided 356 pairs of TON% and mud% values. The data from Serpetti (2012) and Serpetti *et al.* (2012) also included values for median grain size, but not the Cefas data. So, we also assembled data from the Cefas Data Hub on particle size distributions, measured according to the same protocol as followed by Serpetti (2012) and Serpetti *et al.* (2012), and from these computed percentage mud content and whole-sediment median grain size. Using the combined data sets we were able to compute an approximate relationship between median grain size and mud content (Table 5, Figure 4, 5).

$$mud\% = \text{Min}\{100, 10^{tp1} \cdot D^{tp2}\} \quad \text{eqn 7}$$

$$TON\% = 10^{tp3} \cdot (mud\%)^{tp4} \quad \text{eqn 8}$$

TABLE 5 Fitted parameters for the relationships between mud content and whole-sediment median grain size, and between Total Organic Nitrogen (TON; % dry weight) and mud content.

| Relationship | Regression statistics | Parameter | Estimated value | Standard error |
|---------------------------|--|------------|-----------------|----------------|
| mud% vs median grain size | <i>R</i> -squared: 0.6528 | <i>tp1</i> | 0.65653 | 0.02263 |
| | <i>F</i> -statistic: 577.2 on 1 and 307 DF | <i>tp2</i> | -0.80043 | 0.03332 |
| TON% vs mud% | <i>R</i> -squared: 0.6576 | <i>tp3</i> | -1.96546 | 0.03247 |
| | <i>F</i> -statistic: 674.1 on 1 and 351 DF | <i>tp4</i> | 0.58994 | 0.02272 |

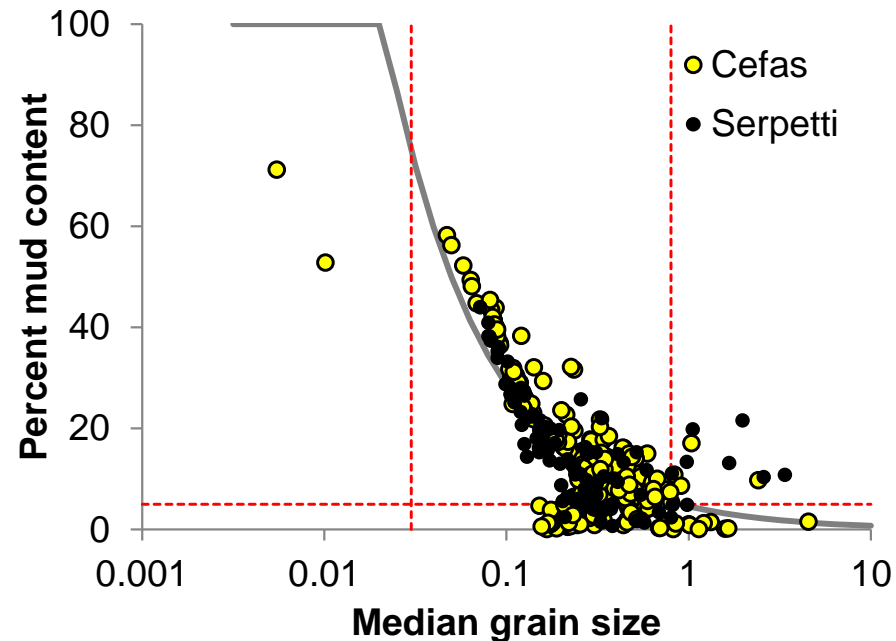


FIGURE 4 Data on mud content and corresponding whole-sediment median grain size from Serpetti (2012), Serpetti *et al.* (2012) and the Cefas Data Hub. As is to be expected given that many of the samples will have been from mixed sediment types, there is considerable scatter in these data, so we confined the dataset to points within the ranges defined by the red-dashed lines for the purpose of fitted the parameters for the curve shown by the grey line (Table 5).

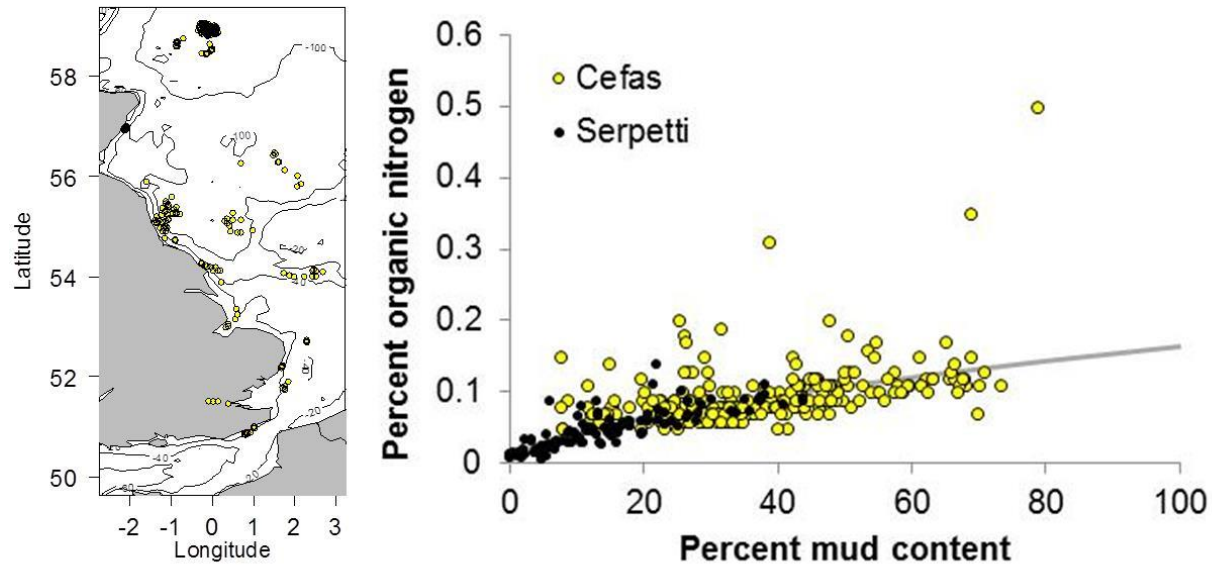


FIGURE 5 Left panel, sampling locations of sediments for Total Organic Nitrogen content (% dry weight) and mud content in Serpetti (2012), Serpetti *et al.* (2012) and the Cefas Data Hub. Right panel, scatter plot of the two data sets with the fitted relationship shown by the grey line (Table 5).

Fixed biological configuration parameters which were not subject to fitting

TABLE 6 Background to the fixed biological configuration parameters for the North Sea model which were not subject to fitting.

| Data | Description |
|--|---|
| Assimilation efficiencies for each living guild in the model (Heath, 2012) | Fixed parameters defining the proportion of ingested mass of food that contributes to new body tissue, after subtracting defecation and the metabolic costs of digestion and synthesis. |
| Biomass loss rates due to | Proportion of biomass lost to ammonia per day due to non-feeding related metabolism at a given reference |

| | |
|--|--|
| temperature-dependent metabolism for each living resource guild | temperature. Ranges for individual guilds broadly related to typical body mass of representative species. Temperature dependency following a Q_{10} function. |
| Q_{10} values for temperature dependent processes, and the Q_{10} reference temperature | Separate Q_{10} values for autotrophic uptake of nutrient, heterotrophic feeding, and heterotrophic metabolism based on literature data. |
| Light intensity required to saturate autotrophic nutrient uptake | Light saturation intensity for nutrient uptake cannot be treated as a fitted value since it is confounded with other uptake parameters. Value estimated from survey of laboratory experiments. |
| Annual weight specific fecundities of planktivorous and demersal fish guilds and the two benthos guilds in the model (suspension/deposit feeders and carnivore/scavenge feeders) | Guild-level values derived by surveying the literature |
| Harvestable biomass density threshold for each resource guild. | The living resource guilds in the model represent a mixture of harvestable and non-harvestable species, especially the invertebrate guilds. The density threshold parameter sets a limit for the guild biomass below which the harvestable species are assumed to be exhausted. Values set from analysis of trawl, plankton and benthos survey species biomass compositions. |
| Minimum inedible biomass of carnivorous zooplankton | The carnivorous zooplankton guild is a key component of the food web, predated on by all the fish and top-predators. However it represents an extremely diverse range of fauna many of which are not edible in significant quantities by the guild predators, e.g. scyphomedusae. A minimum edible threshold is set to ensure that the guild as a whole cannot be extirpated by predation. The value is a rough estimate of scyphomedusae biomass. |

Biological event timing parameters

TABLE 7 Background to the biological event timing parameters for the North Sea model (not subject to fitting).

| Data | Description |
|---|--|
| Spawning start and end dates for fish and benthos | For the fish guilds the dates were obtained from literature survey (Heath, 2012), and for the benthos guilds by reference to Continuous Plankton Recorder (CPR) data for the North Sea (Kirby <i>et al.</i> , 2008). The annual weight-specific fecundity is assumed to be shed uniformly between the start and end dates of spawning. |

| | |
|--|--|
| Recruitment start and end dates for fish and benthos | Obtained from literature survey (Heath, 2012). The annual cohort of larvae/juveniles of each fish and benthos guild is assumed to recruit to the settled stage at a uniform daily rate between the start and end dates. |
| Extra-domain stock biomass of migratory, and the proportion invading the domain each year. Start and end dates for the annual invasion, and start and end dates for the emigration. (see description below). | The main migratory fish species undertaking a seasonal transit of the North Sea is the Atlantic mackerel. Data on the North East Atlantic stock biomass, the proportion entering the North Sea and the timing of the migration, were derived from stock assessment literature (ICES, 2013a) and data on the spatial distribution of landings (Nøttestad <i>et al.</i> , 2016). |

Migratory fish in the North Sea model are assumed to be Atlantic mackerel. The fishery for Atlantic mackerel is one of the most valuable in the northeast Atlantic. Spawning takes place off southwest Ireland in April, and post spawning fish migrating rapidly northwards along the continental shelf edge over several thousand km to feed in the Norwegian Sea or more recently off Iceland (Holst *et al.*, 2016; Nøttestad *et al.*, 2016). The return migration in autumn and winter is slower and a proportion of the stock travels south in shelf waters of the northern North Sea and west of Scotland where a proportion of the harvest is taken (ICES 2013a).

For the purposes of the model, we assume that there is no feedback between fishing and environmental conditions in the North Sea and the biomass and migrations patterns of the whole northeast Atlantic mackerel stock. Implementing such a feedback would be an interesting but separate research project. However, in this version of StrathE2E2 the timing of immigrations and emigrations, and the mass influx across the ocean boundary during the annual immigration phase are treated as period-specific external driving data.

Data on the 'global' stock of northeast Atlantic mackerel (wet biomass) are available from stock assessments (ICES, 2013a), and converted to molar nitrogen mass using appropriate conversion ratios (Greenstreet, 1996). The proportion of the migrating stock entering the North Sea, and the timing of the inward and outward migrations are estimated from monthly resolved data on the spatial distribution of fishery catches. A residual proportion of the peak abundance in the North Sea remaining as residents (if any) is estimated from summer trawl survey data. The model setup code calculates the parameters which are needed in the ecology model. These are the only fixed (i.e. non-fitted) ecology model parameters which are period-specific.

TABLE 8 Migratory fish data and parameters for the periods 1970-1999 and 2003-2013. The data are processed in the model setup to calculate the immigration flux parameters needed in the ecology model.

| Migratory fish data and timing parameters | 1970-1999 | 2003-2013 |
|--|------------------|------------------|
| Migratory fish oceanic biomass (tonnes wet weight) | 3190000 | 3800000 |
| Migratory fish carbon to wet weight (g.g ⁻¹) | 0.184 | 0.184 |

| | | |
|---|--------|--------|
| Model domain sea surface area (for purposes of calculating immigration flux density (km ²)) | 485605 | 485605 |
| Proportion of oceanic population entering the model domain each year | 0.33 | 0.66 |
| Immigration start day | 210 | 210 |
| Immigration end day | 330 | 330 |
| Proportion of peak population in the model domain which remains and does not emigrate | 0.1 | 0.1 |
| Emigration start day | 15 | 15 |
| Emigration end day | 45 | 45 |

Time-varying physical and chemical driving data for the ecology model

Monthly resolution time-varying physical and chemical driving parameters for the model were derived from a variety of sources:

- Temperature, vertical mixing coefficients, volume fluxes, and boundary nutrient, detritus and chlorophyll concentrations from outputs of a NEMO-ERSEM, coupled hydro-geochemical model hindcast from 1980-2015 (Butenschön *et al.*, 2016)
- Bed shear stress due to tidal currents from a simulated climatological year with an FVCOM hydrodynamic model of the North Sea and waters west of the British Isles (Scottish Shelf Model; De Dominicis *et al.*, 2017)
- Remote sensing data products on Suspended Particulate Matter (SPM, ftp://cems-oc.isac.cnr.it/Core/OCEANCOLOUR_GLO_OPTICS_L4_REP_OBSERVATIONS_009_081/dataset-oc-glo-opt-multi-l4-spm_4km_monthly-rep-v02)
- Wave height and period from the ERA-Interim reanalysis (Dee *et al.*, 2011)
- Nitrate data from the NODC World Ocean Data Climatology 2013 (WOA13 V2; Garcia *et al.*, 2014)
- Ammonia data from the ICES Hydro-chemical Data Centre (<http://www.ices.dk/marine-data/data-portals/Pages/ocean.aspx>)
- Atmospheric deposition of nitrate and ammonia from the Co-operative Programme for Monitoring and Evaluation of the Long-range Transmission of Air pollutants in Europe (European Monitoring and Evaluation Programme; EMEP; Simpson *et al.*, 2003, Tarrasón, 2003)
- River nitrate and ammonia concentrations and freshwater volume outflows from a statistical reconstruction of European discharge data 1960-2005 (Heath, 2007a)

Details of how these data were processed are given in Table 9.

TABLE 9 Description of the time-varying (monthly resolution) physical, and chemical driving data for the North Sea model

| Data | Description |
|--|---|
| Natural disturbance rate of each sediment habitat. | Monthly averaged area-proportions of each seabed sedimentary habitat type where the bed shear stress exceed the critical value for particle motion, were taken from the 1/8 degree resolution atlas of seabed sediment properties (Wilson <i>et al.</i> , 2018). The atlas of critical shear stress exceedance was based on a climatological year of high resolution hydrodynamic model outputs (FVCOM Scottish Shelf Model; De Dominicis <i>et al.</i> , 2017), and wave climatology from the ERA-Interim reanalysis (Dee <i>et al.</i> , 2011). Climatological annual cycle of data used for both 1970-1999 and 2003-2013 simulation periods. |
| Vertical mixing coefficients between the upper and lower layers of the deep zone. | Extracted as monthly averaged values from NEMO model output (Butenschön <i>et al.</i> , 2016). Period-specific climatological annual cycle of data used for 1970-1999 and 2003-2013 simulation periods. |
| Volume fluxes into the model domain across open sea boundaries, and from the upper layer of the offshore/deep zone into the inshore/shallow zone, expressed as proportions of the receiving layer volume per day | Monthly averaged daily inflow and outflow volume fluxes derived by integrating daily mean velocities directed perpendicular to the model domain boundary at grid points in each depth layer along transects through outputs from the NEMO hydrodynamic model (Butenschön <i>et al.</i> , 2016). Monthly averaged daily inflow volume fluxes then divided by the volume of the receiving layer in the model domain to estimate a daily flushing rate. Period-specific climatological annual cycles of data used for 1970-1999 and 2003-2013 simulation periods. |
| Monthly averaged temperatures for each water column layer. | Derived by monthly averaging values at grid points within the inshore and vertical layers of the offshore zones from the NEMO hydrodynamic model (Butenschön <i>et al.</i> , 2016). Period-specific climatological annual cycles of data used for 1970-1999 and 2003-2013 simulation periods. |
| Monthly averaged suspended particulate matter (SPM) concentrations ($\text{mg}\cdot\text{m}^{-3}$) in the shallow zone and the deep zone upper layer | Monthly averaged of 4km, 8-day estimates of non-algal surface SPM ($\text{g}\cdot\text{m}^{-3}$) from September 1997 to August 2017 from the Globcolour Project. These data are derived from satellite observations using the algorithm of Gohin (2011). Data were downloaded from the ftp server ftp://cems-oc.isac.cnr.it/Core/OCEANCOLOUR GLO OPTICS L4 REP OBSERVATIONS 009 081/dataset-oc-glo-opt-multi-l4-spm 4km monthly-rep-v02. Climatological annual cycle of data used for both 1970-1999 and 2003-2013 simulation periods. |
| Monthly average light attenuation coefficient for the inshore and offshore surface layers | Parameterised from a linear relationship between light attenuation coefficient and suspended particulate matter concentration (SPM) (Devlin <i>et al.</i> , 2008) |
| Monthly averaged daily integrated irradiance at the sea surface ($\text{E}\cdot\text{m}^{-2}\cdot\text{d}^{-1}$) | Derived from regional meteorology data. Climatological annual cycle of data used for both 1970-1999 and 2003-2013 simulation periods. |
| Monthly averaged daily atmospheric deposition rates of wet and dry, | Derived from 50 x 50 km^2 gridded data for the years 1980, 1985, 1990, 1995, and 2000 - 2015, available from the Co-operative Programme for Monitoring and Evaluation of the Long-range |

| | |
|--|---|
| <p>oxidised and reduced nitrogen onto the sea surface in the shallow and deep zones ($\text{mMN}\cdot\text{m}^{-2}\cdot\text{d}^{-1}$)</p> | <p>Transmission of Air pollutants in Europe (EMEP) Unified 50 x 50 km^2 grid model. Pre-2000 data available only as annual averages, from revision 1.7 of the model; Simpson <i>et al.</i>, 2003, Tarrasón, 2003; see https://www.emep.int/mscw/mscw_ydata.html#Foot2. Data previously downloaded from www.emep.int/Model_data/yearly_data.html in 2007 (Heath 2007a) and no longer available online. Data from 2000 onwards available as monthly averages from https://thredds.met.no/thredds/catalog/data/EMEP/2019_Reporting/catalog.html. Climatological annual cycles of monthly oxidised and reduced nitrogen deposition rates extracted for 2003-2013. Monthly rates relative to the annual mean were then used to generate a climatological seasonal cycle representative of the 1970-1999 period based on the annual mean deposition rates for 1980,1985,1990 and 1995.</p> |
| <p>Monthly averaged , freshwater river infow rates (expressed as a daily proportion o fthe receiving layer volume), and volume weighted concentrations of oxidised and reduced dissolved inorganic nitrogen in the inflowing river waters ($\text{mMN}\cdot\text{m}^{-3}$)</p> | <p>Derived from 1960-2005 monthly averaged nutrient flux and freshwater discharge fluxes into 1 longitude x 0.5 latitude cells around the entire northwest European coastline, originating from a synthesis of national monitoring data and statistical modelling based on rainfall data (Heath, 2007a). Period-specific climatological annual cycles of data used for 1970-1999 and 2003-2013 simulation periods.</p> |
| <p>Mean concentrations of nitrate, ammonia, phytoplankton and suspended detritus ($\text{mMN}\cdot\text{m}^{-3}$), in adjacent ocean waters inflowing to the offshore/deep zone upper layer, adjacent ocean waters inflowing to the offshore/deep zone lower layer, and adjacent shelf waters inflowing to the inshore/shallow zone</p> | <p>Observational data on the boundary variables are of extremely variable resolution. However, simulated outputs of all four variables were available at high space-time resolution from a hind-cast run of the NEMO-ERSEM model (Butenschön <i>et al.</i>, 2016).</p> <p>Comparisons of the available observational data on nitrate, ammonia and chlorophyll with corresponding values in in the NEMO-ERSEM outputs (e.g. Ciavatta <i>et al.</i>, 2016 for chlorophyll) show sufficiently large space-time varying biases as to rule out driving the StrathE2E2 model with boundary data extracted directly from the NEMO-ERSEM outputs. To do so would mean that we were driving the model with data which were inconsistent with the other observational measurements on the state of the ecosystem against which the StrathE2E2 parameters were to be optimized.</p> <p>We therefore used a bias correction methodology (Maraun 2016) to generate monthly resolution 3-dimensional climatologies of nitrate, ammonia and phytoplankton concentrations from the NEMO-ERSEM outputs, and from these we extracted the required boundary concentrations for StrathE2E. We used the same climatological boundary data for both the 1970-1999 and 2003-2013 StrathE2E2 simulation periods on the grounds that the magnitudes of the bias corrections were equivalent to or larger than inter-period differences in the ERSEM outputs.</p> |

Nitrate

A depth-resolved, 1 degree by 1-degree gridded monthly climatology of nitrate was available from the World Ocean Atlas 2013 Version 2 (WOA13 V2; Garcia *et al.*, 2014). These data alone were sufficient to have provided credible climatological boundary conditions for the offshore zone and layers of StrathE2E2, but not for the inshore zone. We therefore used the WOA13 climatology to bias-correct the monthly climatologies of nitrate predicted by NEMO-ERSEM over the period corresponding the years in which the majority of the WOA data were collected (1980-1999).

We first calculated monthly 3-dimensional climatologies of NEMO-ERSEM nitrate data for the period 1980-1999. Then, for grid cells corresponding to the mid-points of WOA13 cells we calculated the relative bias or “change factor” ($\text{Nitrate}_{\text{WOA}}/\text{Nitrate}_{\text{ERSEM}}$), and interpolated these over the entire 3-dimensional grid using nearest neighbour. Finally, the interpolated monthly change factors grids were multiplied into the monthly ERSEM climatologies to generated a bias-corrected climatological annual cycle of 3-dimensional nitrate concentrations.

Ammonia

Observational ammonia data for the period 1980-2013 were downloaded from the ICES data portal (<http://www.ices.dk/marine-data/data-portals/Pages/ocean.aspx>). Each observation was resolved by time, longitude, latitude and depth of collection. However, the data set was extremely sparse with large spatio-temporal gaps.

We first created an aggregated observational data set at a spatial resolution of 1 by 1 degrees per month, and for the surface and deep layers (surface defined as the top 30 m). This was an approximate climatology of ammonia, but with many gaps. We then followed the same procedure as for nitrate to create a bias-corrected climatological annual cycle of 3-dimensional ammonia concentrations.

Phytoplankton

There are no space-time gridded data produces of in-situ phytoplankton biomass. In addition, satellite chlorophyll products are almost universally based on algorithms fitted to global observational data and can therefore exhibit regional spatial and temporal bias. We therefore used the surface (top 5 m) chlorophyll climatology of Clarke *et al.* (2006) which was generated by a statistical methodology to blend satellite chlorophyll with in-situ water-bottle derived chlorophyll measurements from the North Atlantic. This surface climatology was first interpolated to the

| | |
|--|--|
| | <p>NEMO-ERSEM horizontal grid. We then calculated a comparable monthly climatology of surface chlorophyll from NEMO-ERSEM outputs. As with nitrate, we then calculated a change-factor ($\text{Chl}_{\text{Clarke}}/\text{Chl}_{\text{ERSEM}}$) for each surface grid cell and in each month and applied this across all NEMO-ERSEM chlorophyll predictions at each location in time and space. The change factors were applied uniformly across all depths on the assumption that NEMO-ERSEM reflected the true vertical distribution of chlorophyll. This resulted in a set of monthly 3-dimensional chlorophyll data sets where the surface chlorophyll climatology matched the observational data set of Clarke <i>et al.</i> (2006). Chlorophyll concentrations were then converted to nitrogen units assuming carbon:chlorophyll (weight ratio) of 20 and Redfield molar ratios of carbon:nitrogen.</p> <p>Suspended detritus Observational data on organic detritus are extremely sparse. Conditions for detritus nitrogen were therefore taken directly from NEMO-ERSEM without any bias correction.</p> <p>Boundary data extraction For each bias-corrected climatological data set we calculated monthly averages of values at grid cells located at vertical slices along each of the open inshore and offshore zone boundaries and depth layers of the StrathE2E2 model domain.</p> |
|--|--|

Inputs to the North Sea fishing fleet model

Background

The key configuration data for the fishing fleet model are the definitions of the gears in terms of their power with respect to each of the harvestable resource guilds, discarding rates, processing-at-sea rates, and their seabed abrasion rates. These can be regarded as static parameters for each gear.

An additional class of static parameters is the scaling coefficients between effort (activity x power) and the harvest ratio generated on each model resource guild. These parameters have to be derived by fitting.

Finally, there are parameters which we can consider as driving data since they would be expected to vary with time. These are the activity rates of each gear, and their spatial distributions across the habitat types.

Static gear-definition parameters in the fishing fleet model

TABLE 10 Description of static parameters for the fishing fleet model. These parameters would be expected to remain constant over time, so any changes invoked would imply a change in the design or operation of a gear type.

| Data | Description |
|---|--|
| Definition of gear types | Data for 2003-2013 derived from EU Scientific, Technical and Economic Committee for Fisheries (STECF) reports (data filename: '2014_STECF 14-20 - Fishing Effort Regimes data tables.zip'; http://stecf.jrc.ec.europa.eu/data-reports). Full description in Heath <i>et al.</i> , 2015. |
| Definition of the proportion of retained catch of each model resource guild which is processed (gutted) at sea by each gear type, and the proportion of live weight discarded as offal as a result of processing. | Proportion of live weight discarded by processing, from Coull <i>et al.</i> (1989). Proportion of retained catch processed at sea approximately estimated from the fish market sampling data. |
| Scaling parameters relating effort to harvest ratios applied to each model resource guild | Derived by fitting harvest ratios, as described later in this document.. |
| Seabed abrasion rates of each gear type. | Data on the area of seabed disturbed per unit time of towing by different fishing gears obtained from published studies (Eigaard <i>et al.</i> , 2015). |
| Sediment penetration depth for seabed-contact fishing gears | Single penetration depth (5 cm) assumed across all seabed-contact gears, and independent of sediment type, based on data from Eigaard <i>et al.</i> (2015). |
| Damage mortality rates on benthos species caused by seabed-contact towed gears | Proportion of fauna killed per trawl pass assuming 5 cm penetration depth, obtained from literature meta-analysis (Hiddink <i>et al.</i> , 2017) |
| Parameters for an empirically-based relationship between demersal fish biomass in the model, and the proportion of annual catch weight made up of non-quota limited species (see this document for details). | Relationship established from analysis of research vessel trawl survey catch per unit effort data, and species landings data: $\varphi_{NQ} = p_{11} \cdot e^{-p_{12} \cdot B}$ where φ_{NQ} is the proportion of annual commercial catch weight which comprises non-quota species, B is the demersal fish community biomass on 1 January, p_{11} and p_{12} are fitted parameters. |
| Parameters for empirically based relationships between demersal fish biomass in the model, and the proportion of annual catch which is undersize for landing or marketing. Separate relationships for quota-limited and non-quota species groups (see this document for details). | Relationship established from analysis of research vessel trawl survey catch per unit effort data, and species landings data: $\varphi_{U,x} = p_{13,x} \cdot e^{-p_{14,x} \cdot B}$ where $\varphi_{U,x}$ (x = quota-limited/non-quota) is the proportion of catch weight which is undersize for legal landing (quota-limited group) or for marketing (non-quota group), B is the demersal fish community biomass on 1 January, $p_{13,x}$ and $p_{14,x}$ are fitted parameters. |

Potentially time-varying parameters of the fishing fleet model

TABLE 11 Description of potentially time-varying driving data for the fishing fleet model.

| Data | Description |
|--|--|
| Catching power and discard rates of each resource guild by each gear | <p>Data on power and discard rates for 2003-2013 derived from EU Scientific, Technical and Economic Committee for Fisheries (STECF) reports (data filename: '2014_STECF 14-20 - Fishing Effort Regimes data tables.zip'; http://stecf.jrc.ec.europa.eu/data-reports). Full description in Heath <i>et al.</i>, 2015.</p> <p>Technological creep was assumed to results in a 2% per year increase in power for all gear/guild combinations (Engelhard, 2008; Palomares & Pauly, 2019).</p> <p>Data for parameterising catching power of gears for small cetaceans derived from a variety of sources including the ICES Working Group on the Bycatch of Protected Species (ICES, 2015a,b)</p> |
| Regional activity rates, of each gear type | <p>Data for 2003-2013 derived from EU Scientific, Technical and Economic Committee for Fisheries (STECF) reports (data filename: '2014_STECF 14-20 - Fishing Effort Regimes data tables.zip'; http://stecf.jrc.ec.europa.eu/data-reports). Full description in Heath <i>et al.</i> (2015).</p> |
| Spatial proportional distribution of activity by each gear | <p>Proportion of domain-wide annual average activity rate over each seabed habitat type, derived by overlaying spatial distributions of activity from the EU Scientific, Technical and Economic Committee for Fisheries (STECF) reports, onto spatial distributions of seabed sediment types derived from the atlas of sediment properties (Wilson <i>et al.</i>, 2018).</p> |

Data processing to derive fish and invertebrate related parameters for the fleet model

Processing of Scientific, Technical and Economic Committee for Fisheries (STECF) and Norwegian landings data analysis

Data on the landings, discards, activity and economic performance of the fleet sectors of all EU member states are available for STECF (<https://stecf.jrc.ec.europa.eu/dd/effort>). From 2000 onwards, the landings and effort data are resolved by at least 1 longitude x ½ latitude cells

(approximately 30 x 30 nautical miles). Discard data are available only at a more aggregated spatial resolution. The dataset for NW European water includes records on 101 different species covering mostly finfish. Invertebrates are under-represented in the records. We aggregated the species data into the coarse 'functional' categories defined in the StrathE2E2 model.

The data contain 32 different fishing gear designations. Some of which are local variants appropriate to particular countries or regions. We aggregated the STECF gear types up into 11 coarser groups for use in StrathE2E2. The aggregation rules are shown in Table 12

TABLE 12 Correspondence between raw STECF gear codes and gear categories in the analyses presented here.

| STECF Code | Gear description | StrathE2E2 model gear type |
|---|---|-----------------------------------|
| PELAGIC TRAWLS | Pelagic trawls | Pelagic trawls & seines |
| PEL_TRAWL | Pelagic Trawl | |
| PEL_SEINE | Pelagic seine nets | |
| TR3 | Bottom trawls and seines of mesh size equal to or larger than 16 mm and less than 32 mm – mostly targeting sprat. | Sandeel & sprat trawls |
| OTTER | Bottom trawls (for sandeel) | |
| LL (landings in a statistical rectangle >50% by weight mackerel) | Drifting longlines (for pelagic fish) | Longline mackerel |
| BT2 | Beam trawls of mesh equal to or larger than 80 mm and less than 120 mm. | Beam trawl demersal |
| BT1 | Beam trawls of mesh equal to or larger than 120 mm | |
| DEM_SEINE | Danish and Scottish seiners | Demersal seine |
| TR1 | demersal trawls/seines with larger mesh sizes > 100MM | Demersal otter trawl (mainly TR1) |
| TR2 (landings in a statistical rectangle <30% by weight Norway lobster) | Demersal trawls and seines with mesh 70-99mm | |
| BOTTOM | Bottom trawls | |

| | | |
|---|--|------------------------------|
| TRAWLS | | |
| 3A | Bottom trawler mesh size \geq 32 mm) | |
| LL (landings in a statistical rectangle <50% by weight mackerel) | Set longlines (for demersal fish) | Longline & gillnets demersal |
| GN1 | Gill nets, entangling nets. | |
| GILL | Drift and fixed Nets except Trammel Nets | |
| 3B | Gillnet \geq 60 mm | |
| TRAMMEL | Trammel nets | |
| GT1 | Trammel nets | |
| BEAM | Beam trawl targeting shrimp (in the North Sea) | |
| TR2 (landings in a statistical rectangle >30% by weight Norway lobster) | Demersal trawls and seines with mesh 70-99mm | Nephrops trawl |
| POTS | Pots and traps | Creels and pots |
| DREDGE | Dredges (targeting scallops in the North Sea) | Mollusc dredges |

There were some specific considerations involved in the gear aggregations, depending on the species targeted in particular areas.

Beam trawls: The STECF beam trawl categories BT1 and BT2 more or less exclusively target demersal finfish species in the North Sea. It is therefore reasonable to combine BT1 and BT2 into a single model gear. However, care must be taken with the BEAM class. In the North Sea, landings from BEAM gears are more or less exclusively common shrimp (98.7% of the total). Elsewhere, landings from BEAM are almost exclusively demersal fish. We therefore created a 'Beam trawl shrimp' gear in the North Sea.

TR2 trawls: The TR2 category covers gears used in a variety of fisheries in the North Sea:

- A fishery for Nephrops, which has a significant bycatch of demersal fish
- Mixed fishery in the southern North Sea, with whiting and other finfish species as the main components
- Danish and Swedish fishery targeting demersal finfish in the Skagerrak.

Since the targeted Nephrops fishery operates exclusively in muddy areas and there are particular concerns about the seabed impact of this fishery we sought to disaggregate TR2 to identify the Nephrops trawl component. Country level landings data help us with the disaggregation. If the TR2 landings by an individual country from an individual ICES statistical rectangle comprised more than 30% Nephrops then we assigned that rectangle's TR2 landings and activity to the Nephrops gear. If it was less than 30% we assigned it to the demersal otter trawl category.

Longlines (LL): Longlines are extensively used in drifting, near-surface set mode for catching mainly mackerel and tuna, and in a near-seabed set mode for catching demersal fish such as cod and ling. Apart from the very different species-targeting of these two modes of operation, there are consequences for by-catch of non-target species. In particular, various seabird species are vulnerable to

Estimating Norwegian fishing activity from landings data

A shortcoming of the STECF data is that it only includes data from EU Member States. So, while effort and landings from EU activity in Norwegian and Faroese territorial waters are included, the equivalent data for Norwegian and Faeroese vessels are not. This is a problem for analysis of spatial and national shares of total yields and effort in the North Sea where Norway has a significant share of the total catch. The Faeroe Islands also have an access agreement with the EU, but their activities in the North Sea are relatively minor.

We made a request to the Norwegian Directorate of Fisheries, Statistics Department, who provided all the Norwegian annual landings data from the North Sea and west of Scotland regions, resolved by species and 1 longitude x ½ latitude cells for the years 2003-2016. We processed these data to conform with the STECF data, but still we lacked a breakdown of the landings by gear, or any record of the activity rates of Norwegian vessels to compare with the EU activity data. However, given that each of the STECF gears largely targets particular species (e.g. TR3 targets sandeels/sprats, etc), and assuming that the pattern of targeting and the selectivity of the gear types is the same in the EU and Norwegian fishery, we developed a scheme to impute the Norwegian gear activity and the distribution of Norwegian landings across gear types for a given year and geographic area (Figure 6). We carried out this procedure for the inshore and offshore zones separately.

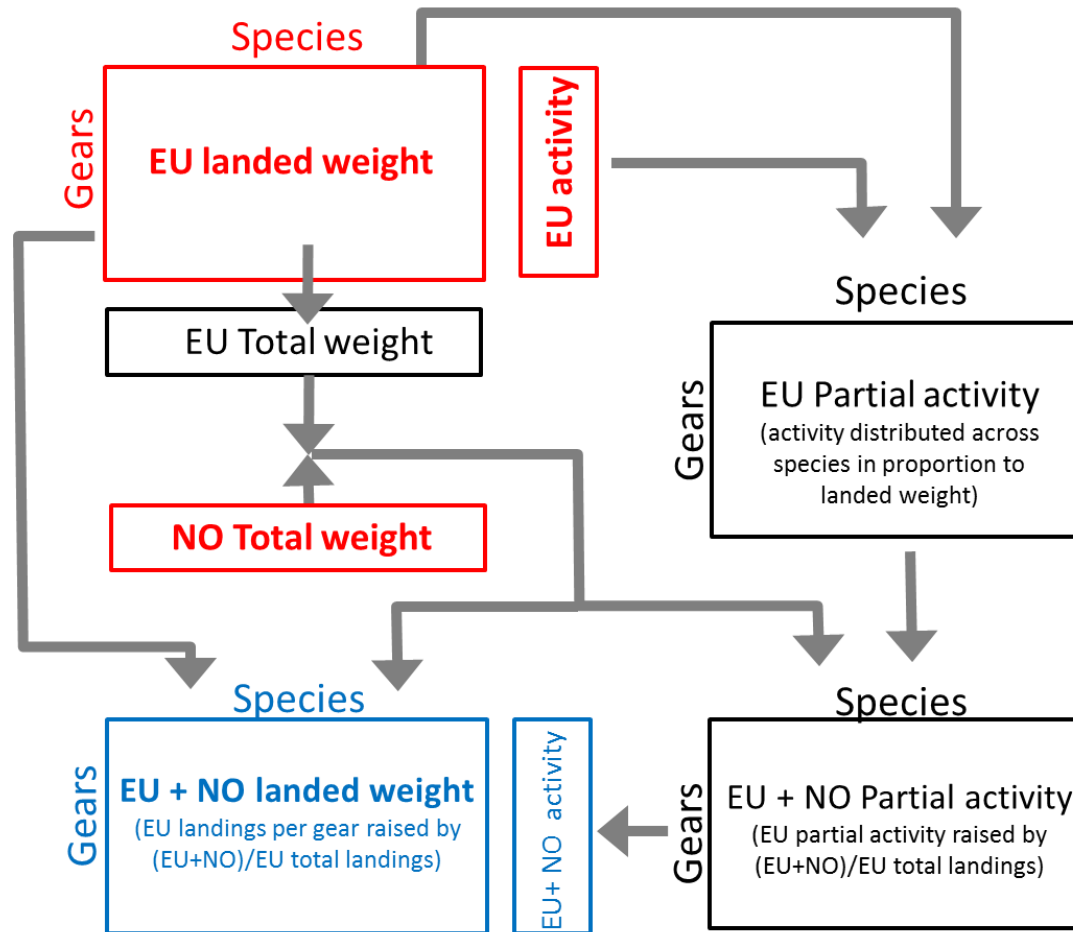


FIGURE 6 Workflow for imputing Norwegian effort per gear type, and the distribution of Norwegian landings across gear types in a given year and geographic area, given the STECF data and the Norwegian landings data obtained from the Directorate of Fisheries. Red cells indicate the input data that we have from STECF and the Norwegian Fisheries Directorate, blue cells indicate the data we wish to impute, grey cells represent intermediate data generated during the processing. EU = European Union fleet data, NO = Norwegian fleet data. Estimates of Norwegian effort and landings per gear alone, is simply the imputed total (EU+NO) minus the known EU component.

Impacts of different fishing gears on the seabed

Eigaard *et al.* (2015) evaluated seabed areas impacted per hour of trawling by a range of fishing gear fleets, including those listed by STECF. We mapped our gear classes on to those of Eigaard *et al.* and produced estimates of ploughed area per unit time, assuming a gear penetration depth of 5cm (Table 13).

TABLE 13. Seabed abrasion rates for fishing fleets in the North Sea (Eigaard *et al.*, 2015).

| Gear category | Seabed abrasion rate $m^2.s^{-1}$ |
|------------------------------|---|
| Pelagic trawls and seines | 0 |
| Sandeel and sprat trawls | 8.8 |
| Longline Mackerel | 0 |
| Beam trawl demersal | 54.1 |
| Demersal Seine | 22.4 |
| Demersal otter trawl | 17.1 |
| Longline & gillnets demersal | 0 |
| Beam trawl shrimp | 13.5 |
| Nephrops trawl | 78.9 |
| Creels & pots | 0 |
| Mollusc dredges | 22.4 |
| Norwegian whalers | 0 |

Regional activity density of each gear category

For each of the StrathE2E2 gear types assembled from the STECF data (Table 12), we summed the inshore and offshore annual activity rates (seconds per year) of the StrathE2E2 gear groups, and the corresponding Norwegian activity, and divided by the area of the whole model domain to obtain an annual activity density. The inherent assumption in this process was that the STECF gears contributing to each StrathE2E2 gear group have equivalent power. Finally, we then averaged the annual values for each gear group over the duration of the STECF data period (2003-2013).

TABLE 14 2003-2013 annual average activity (hours per year within the StrathE2E2 model domain) by EU and Norwegian vessels for each of the gear aggregations in the fleet model.

| StrathE2E2 gear group | Activity (s.m⁻².d⁻¹) |
|---|---|
| Pelagic Trawl & Seine | 2.17E-06 |
| Sandeel & sprat trawls (Otter30-70mm & TR3) | 4.23E-06 |
| Longline mackerel | 1.68E-06 |
| Beam Trawl demersal (BT1 & BT2) | 1.15E-05 |
| Demersal Seine | 1.72E-08 |
| Demersal Otter Trawl (TR1) | 2.16E-05 |
| Gill Nets & Longline demersal | 7.92E-06 |
| Beam Trawl shrimp | 1.27E-05 |
| Nephrops Trawl (TR2) | 1.72E-05 |
| Creels | 2.40E-05 |
| Mollusc Dredge | 3.11E-06 |

Spatial distribution of fishing activity

The 2003-2013 average geographical distribution of EU gear activity, resolved by 1 degree longitude x 0.5 degree latitude statistical rectangles was derived from the STEFC database. Then, we overlaid the spatial distribution of the 8 seabed habitat classes in the model, and derived the proportion of total EU activity occurring within each habitat (Table 15). Finally, we assumed that Norwegian activity in each gear class was distributed in proportion to EU activity.

There are no data on the spatial distributions of gear activity during the 1970-1999 model fitting period, so we assumed that the proportional distribution of activity during 2003-2013 was also representative of this earlier period.

TABLE 15 2003-2013 average proportion of North Sea domain-wide activity by each gear category occurring over each seabed habitat, derived from STECF activity data and seabed sediment class data. The bottom row shows the area-proportions of each seabed habitat for comparison.

| Gear category | Shallow rock | Shallow mud | Shallow sand | Shallow gravel | Deep rock | Deep mud | Deep sand | Deep gravel |
|--------------------------------|---------------------|--------------------|---------------------|-----------------------|------------------|-----------------|------------------|--------------------|
| Pelagic trawls & seines | 0.0023 | 0.0223 | 0.1857 | 0.0566 | 0.0041 | 0.3618 | 0.3324 | 0.0349 |
| Sandeel & sprat trawls | 0.0016 | 0.0419 | 0.1885 | 0.0993 | 0.0033 | 0.2203 | 0.4222 | 0.0229 |
| Longline Mackerel | 0.0026 | 0.0192 | 0.1334 | 0.0046 | 0.0022 | 0.0481 | 0.7862 | 0.0037 |
| Beam trawl demersal | 0.0000 | 0.0103 | 0.5558 | 0.0516 | 0.0000 | 0.1579 | 0.2183 | 0.0061 |
| Demersal seine | 0.0108 | 0.0417 | 0.2481 | 0.0393 | 0.0157 | 0.2316 | 0.3695 | 0.0433 |
| Demersal otter trawl | 0.0038 | 0.0055 | 0.0791 | 0.0868 | 0.0096 | 0.3050 | 0.4735 | 0.0367 |
| Longline & gillnets demersal | 0.0176 | 0.0115 | 0.3819 | 0.3067 | 0.0097 | 0.0575 | 0.2055 | 0.0096 |
| Beam trawl shrimp | 0.0000 | 0.0414 | 0.8679 | 0.0612 | 0.0000 | 0.0190 | 0.0102 | 0.0003 |
| Nephrops trawl | 0.0000 | 0.1137 | 0.0000 | 0.0000 | 0.0000 | 0.8863 | 0.0000 | 0.0000 |
| Creels & pots | 0.0148 | 0.0484 | 0.1900 | 0.1670 | 0.0238 | 0.0503 | 0.4214 | 0.0845 |
| Mollusc dredges | 0.0000 | 0.0114 | 0.2565 | 0.0929 | 0.0000 | 0.0508 | 0.4915 | 0.0969 |
| Habitat area proportion | 0.0030 | 0.0110 | 0.1878 | 0.0478 | 0.0057 | 0.2665 | 0.4595 | 0.0187 |

Allocation of STECF fish and benthos landings and discards to gear classes

Full details of the procedures for a) allocating landed and discarded species to StrathE2E2 model guilds, b) allocating landings to gears, and c) allocating discards to gears and statistical rectangles, are described by Heath *et al.* (2015). Brief summaries of the essential results are presented here.

TABLE 16 2003-2013 average annual live weights landed (tonnes) of each StrathE2E2 fish and invertebrate resource category, by each gear in the fleet model, derived from the combined STECF and Norwegian data

| Gear category | Planktivorous fish | Demersal fish | Migratory fish | Suspension/ deposit feeding benthos | Carnivore/ scavenge feeding benthos | Carnivorous zooplankton |
|------------------------------|--------------------|---------------|----------------|-------------------------------------|-------------------------------------|-------------------------|
| Pelagic trawls & seines | 349610 | 237 | 185311 | 0 | 1 | 0 |
| Sandeel & sprat trawls | 358202 | 571 | 13462 | 1 | 485 | 0 |
| Longline Mackerel | 2 | 11 | 1038 | 0 | 8 | 0 |
| Beam trawl demersal | 1 | 58239 | 51 | 3 | 687 | 0 |
| DemersalSeine | 0 | 25 | 16 | 0 | 0 | 6 |
| Demersal otter trawl | 1338 | 121323 | 3912 | 7 | 1875 | 1850 |
| Longline & gillnets demersal | 173 | 9177 | 28 | 2 | 50 | 0 |
| Beam trawl shrimp | 152 | 269 | 14 | 0 | 27124 | 0 |
| Nephrops trawl | 6 | 12005 | 133 | 1 | 16706 | 237 |
| Creels & pots | 1 | 66 | 80 | 8 | 7277 | 0 |
| Mollusc dredges | 32 | 11 | 2 | 4314 | 3 | 0 |
| Totals | 709516 | 201934 | 204046 | 4336 | 54216 | 2093 |

TABLE 17 2003-2013 average annual live weights discarded (tonnes) of each fish and invertebrate StrathE2E2 resource category, by each gear in the fleet model, derived from STECF

| Gear category | Planktivorous fish | Demersal fish | Migratory fish | Suspension/ deposit feeding benthos | Carnivore/ scavenge feeding benthos | Carnivorous zooplankton |
|------------------------------|---------------------------|----------------------|-----------------------|--|--|--------------------------------|
| Pelagic trawls & seines | 1031 | 30 | 10219 | 0 | 0 | 0 |
| Sandeel & sprat trawls | 42 | 664 | 118 | 0 | 101 | 0 |
| Longline Mackerel | 0 | <1 | 0 | 0 | 0 | 0 |
| Beam trawl demersal | 0 | 69626 | 104 | 0 | 121 | 0 |
| DemersalSeine | 0 | 0 | 0 | 0 | 0 | 0 |
| Demersal otter trawl | 192 | 109270 | 13080 | <1 | 852 | 0 |
| Longline & gillnets demersal | 0 | 260 | 9 | 0 | 49 | 0 |
| Beam trawl shrimp | 111 | 8770 | 8 | 0 | 36853 | 0 |
| Nephrops trawl | 4 | 26014 | 104 | 0 | 1070 | 0 |
| Creels & pots | 0 | 1 | 0 | 0 | 0 | 0 |
| Mollusc dredges | 0 | 10 | 0 | 24 | 0 | 0 |
| Totals | 1380 | 214644 | 23641 | 25 | 39048 | 0 |

Landings data required to be converted from live wet weight to nitrogen mass for use in the model. The wet weight to nitrogen conversion factors assumed as given in Table 18.

TABLE 18 Nitrogen content (mMN) per gram wet weight for living guilds which were assumed for converting landed live weights to nitrogen mass. Where necessary the values were estimated from quoted carbon mass data assuming Redfield molar ratios.

| Guild | Nitrogen mass per unit wet weight (mMN.gWW⁻¹) | Source |
|------------------------------------|---|---|
| Macrophytes | 2.070 | Black (1950), Sjøtun <i>et al.</i> (1996) |
| Carnivorous zooplankton | 1.258 | Greenstreet (1996) |
| Planktivorous fish | 2.038 | Greenstreet (1996) |
| Demersal fish | 1.340 | Greenstreet (1996) |
| Migratory fish | 2.314 | Greenstreet (1996) |
| Suspension/deposit feeding benthos | 0.503 | Greenstreet (1996), Ricciardi & Bourget (1998) |
| Carnivore/scavenge feeding benthos | 1.006 | Greenstreet (1996), Ricciardi & Bourget (1998) |
| Birds | 2.518 | Taylor & Konarzewski (1989), and assuming dry weight/wet weight = 0.440, carbon weight/dry weight = 0.455, and Redfield molar C:N ratio = 6.625. |
| Pinnipeds | 2.518 | Lavigne <i>et al.</i> (1986), and assuming dry weight/wet weight = 0.440, carbon weight/dry weight = 0.455, and Redfield molar C:N ratio = 6.625. |
| Cetaceans | 2.518 | Lavigne <i>et al.</i> (1986), and assuming dry weight/wet weight = 0.440, carbon weight/dry weight = 0.455, and Redfield molar C:N ratio = 6.625. |

TABLE 19 Discard rates (proportion of catch discarded) by each gear with respect to each resource guild. Values calculated for the period 2003-2013.

| Gear category | Planktivorous fish | Demersal fish | Migratory fish | Suspension/deposit feeding benthos | Carnivore/scavenge feeding benthos | Carnivorous zooplankton |
|-------------------------|---------------------------|----------------------|-----------------------|---|---|--------------------------------|
| Pelagic trawls & seines | 0.003 | 0.112 | 0.052 | 0 | 0 | 0 |
| Sandeel & sprat trawls | 0 | 0.538 | 0.009 | 0 | 0.173 | 0 |
| Longline Mackerel | 0 | 0.005 | 0 | 0 | 0 | 0 |
| Beam trawl demersal | 0 | 0.545 | 0.671 | 0.014 | 0.150 | 0 |
| DemersalSeine | 0 | 0 | 0 | 0 | 0.000 | 0 |
| Demersal otter | 0.126 | 0.474 | 0.770 | 0.022 | 0.313 | 0 |

| | | | | | | |
|------------------------------|-------|-------|-------|-------|-------|---|
| trawl | | | | | | |
| Longline & gillnets demersal | 0 | 0.027 | 0.231 | 0 | 0.498 | 0 |
| Beam trawl shrimp | 0.422 | 0.970 | 0.366 | 0 | 0.576 | 0 |
| Nephrops trawl | 0.380 | 0.684 | 0.439 | 0 | 0.060 | 0 |
| Creels & pots | 0.056 | 0.008 | 0.001 | 0 | 0 | 0 |
| Mollusc dredges | 0 | 0.477 | 0 | 0.006 | 0 | 0 |

TABLE 20 Catching power (mMN s^{-1}) of each gear with respect to each fish and invertebrate resource guild in the ecology model for the period 2003-2013. Values calculated as the ratio of catch rate : activity rate for each gear.

| Gear category | Planktivorous fish | Demersal fish | Migratory fish | Suspension/ deposit feeding benthos | Carnivore/ scavenge feeding benthos | Carnivorous zooplankton |
|------------------------------|---------------------------|----------------------|-----------------------|--|--|--------------------------------|
| Pelagic trawls & seines | 1932.12 | 0.89 | 1077.41 | 0.00 | 0.00 | 0.00 |
| Sandeel & sprat trawls | 1013.87 | 2.12 | 38.43 | 0.00 | 0.77 | 0.00 |
| Longline Mackerel | 0.01 | 0.05 | 7.38 | 0.00 | 0.03 | 0.00 |
| Beam trawl demersal | 0.00 | 80.90 | 0.16 | 0.00 | 0.39 | 0.00 |
| Demersal Seine | 0.00 | 10.64 | 10.94 | 0.00 | 0.02 | 5.09 |
| Demersal otter trawl | 0.85 | 77.63 | 9.42 | 0.00 | 0.70 | 1.50 |
| Longline & gillnets demersal | 0.26 | 8.66 | 0.06 | 0.00 | 0.07 | 0.00 |
| Beam trawl shrimp | 0.25 | 5.18 | 0.02 | 0.00 | 27.91 | 0.00 |
| Nephrops trawl | 0.01 | 16.07 | 0.16 | 0.00 | 5.72 | 0.14 |
| Creels & pots | 0.00 | 0.02 | 0.04 | 0.00 | 1.68 | 0.00 |
| Mollusc dredges | 0.12 | 0.05 | 0.01 | 5.80 | 0.01 | 0.00 |

TABLE 21 Catching power (mMN s⁻¹) of each gear with respect to each fish and invertebrate resource guild in the ecology model for the period 1970-1999. Values calculated from the 2003-2013 data (Table 20) by assuming a 2% per year increase in power (Engelhard, 2008; Palomares & Pauly, 2019) for each gear/guild combination. The arithmetic mean power during 1970-1999 was thus estimated to be 0.65359 * the mean power during 2003-2013.

| Gear category | Planktivorous fish | Demersal fish | Migratory fish | Suspension/ deposit feeding benthos | Carnivore/ scavenge feeding benthos | Carnivorous zooplankton |
|------------------------------|--------------------|---------------|----------------|-------------------------------------|-------------------------------------|-------------------------|
| Pelagic trawls & seines | 1262.81 | 0.58 | 704.19 | 0.00 | 0.00 | 0.00 |
| Sandeel & sprat trawls | 662.65 | 1.39 | 25.12 | 0.00 | 0.50 | 0.00 |
| Longline Mackerel | 0.01 | 0.03 | 4.82 | 0.00 | 0.02 | 0.00 |
| Beam trawl demersal | 0.00 | 52.87 | 0.11 | 0.00 | 0.25 | 0.00 |
| Demersal Seine | 0.00 | 6.95 | 7.15 | 0.00 | 0.01 | 3.33 |
| Demersal otter trawl | 0.55 | 50.74 | 6.16 | 0.00 | 0.46 | 0.98 |
| Longline & gillnets demersal | 0.17 | 5.66 | 0.04 | 0.00 | 0.05 | 0.00 |
| Beam trawl shrimp | 0.16 | 3.38 | 0.01 | 0.00 | 18.24 | 0.00 |
| Nephrops trawl | 0.00 | 10.50 | 0.11 | 0.00 | 3.74 | 0.09 |
| Creels & pots | 0.00 | 0.01 | 0.03 | 0.00 | 1.10 | 0.00 |
| Mollusc dredges | 0.08 | 0.03 | 0.00 | 3.79 | 0.00 | 0.00 |

Processing of catch at sea and production of offal

The proportion of the catch of each resource guild which is processed at sea aboard each gear group was estimated roughly from market sampling data (proportion of landing as whole vs gutted fish) and expert knowledge (Table 22). The proportion of live weight discarded as offal as a result of processing was estimated to be 10%.

TABLE 22 Processing-at-sea proportions for each gear with respect to each fish and invertebrate resource guild in the ecology model.

| Gear category | Planktivorous fish | Demersal fish | Migratory fish | Suspension/ deposit feeding benthos | Carnivore/ scavenge feeding benthos | Carnivorous zooplankton |
|---------------|--------------------|---------------|----------------|-------------------------------------|-------------------------------------|-------------------------|
|---------------|--------------------|---------------|----------------|-------------------------------------|-------------------------------------|-------------------------|

| | | | | | | |
|------------------------------|---|-----|---|---|-----|---|
| Pelagic trawls & seines | 0 | 0 | 0 | 0 | 0 | 0 |
| Sandeel & sprat trawls | 0 | 0 | 0 | 0 | 0 | 0 |
| Longline Mackerel | 0 | 0 | 0 | 0 | 0 | 0 |
| Beam trawl demersal | 0 | 0.5 | 0 | 0 | 0 | 0 |
| Demersal Seine | 0 | 0.5 | 0 | 0 | 0 | 0 |
| Demersal otter trawl | 0 | 0.5 | 0 | 0 | 0 | 0 |
| Longline & gillnets demersal | 0 | 0 | 0 | 0 | 0 | 0 |
| Beam trawl shrimp | 0 | 0 | 0 | 0 | 0 | 0 |
| Nephrops trawl | 0 | 0 | 0 | 0 | 0.8 | 0 |
| Creels & pots | 0 | 0 | 0 | 0 | 0 | 0 |
| Mollusc dredges | 0 | 0 | 0 | 0 | 0 | 0 |

Scaling parameters relating effort to harvest ratio for fish and invertebrate guilds in the model

The parameters linking effort (activity x power) of any gear in the fleet model to harvest ratios in the ecology model are key terms in the coupled system. To estimate these independently we need estimates of the integrated harvest ratios for each of the living resource guilds in the ecology model to compare with the integrated effort across all the gears in the fleet model.

Finfish harvest ratios

The approach to estimating fish harvest ratios was a refinement of that outlined by Heath (2012). Annual species-specific total stock biomass data from the analytical assessments conducted by the International Council for the Exploration of the Sea (ICES) were assembled for the period 1960-onwards where possible. For the North Sea these data are available for only the major commercial species but these constitute a high proportion of the community biomass of the guilds of fish in the model (planktivorous: herring, sandeel, Norway pout; demersal: cod, haddock, whiting, plaice, saithe, sole; migratory: mackerel). Assessments have commenced at different times for the various species, with plaice, cod, haddock and herring being the longest-running series. Data-gaps in the early years were filled by extrapolating back in time from the first assessed year using independent trawl survey data as an index of population biomass. In addition, the age at first inclusion in the

assessment varied between 3 and 9 months depending on species, so a compensatory correction was applied to bring the biomasses of all species into line.

Annual catches of each of the assessed species were constructed from the sum of landings and discards as provided in the ICES stock assessment reports. Where discard data were absent, values were filled in from a statistical reconstruction of discard histories for the North Sea (Heath & Cook, 2015), or prior to 1980 by extrapolating the discard rate (proportion of catch discarded).

The whole-guild biomass of each guild was estimated by up-scaling the combined biomass of the assessed species, using a ratio of all-species to assessed species biomass derived from independent trawl surveys. Finally, the annual harvest ratios for each fish guild were determined from whole-guild biomasses and catches across all the species representing each guild;.

The estimated harvest ratios for both planktivorous and demersal fish increased from the 1960s to the 1970s and remained high during the 1990s. During the 2000s harvest ratios decreased towards low levels by 2010 (Figure 7). This pattern is entirely consistent with the changes in fishing mortality reported in the ICES Greater North Sea Eco-region review (ICES, 2016). ICES determined that fishing mortality rates during the period 1970-1999 were around 2-times F_{MSY} (the fishing mortality associated with maximum steady state catch) for demersal fish, and 1.0-1.3-times F_{MSY} for pelagic fish. The ratio F/F_{MSY} does not necessarily correspond to HR/HR_{MSY} , but we can be certain that average harvest ratios during this period (1970-1999) were in excess of HR_{MSY} , especially for demersal fish.

Benthic invertebrate harvest ratios

There are no stock assessments for invertebrates of comparable detail to those for finfish. This is partly due to the inability to reliably determine the age of individuals, so that there is a lack of data to support age-based population dynamics approaches. The most detailed assessments are for Norway lobster (*Nephrops norvegicus*) where television surveys are used to provide fishery-independent data on stock biomass. Some degree of assessment is available for Atlantic scallop and brown shrimp, but very little for other benthic crustaceans and molluscs, or for squids. Based on the time-series of Norway lobster stock and landings we estimated the harvest ratios for carnivorous/scavenge feeding benthic invertebrates (Figure 7), and assumed that these apply also to other invertebrates.

The ICES North Sea Eco-region review (ICES, 2016) determined that fishing mortality rates for invertebrates have risen steadily from around 0.3-times F_{MSY} to 1.25-times F_{MSY} between 1970 and 2010. This synopsis largely reflects the trend in landings over the period.

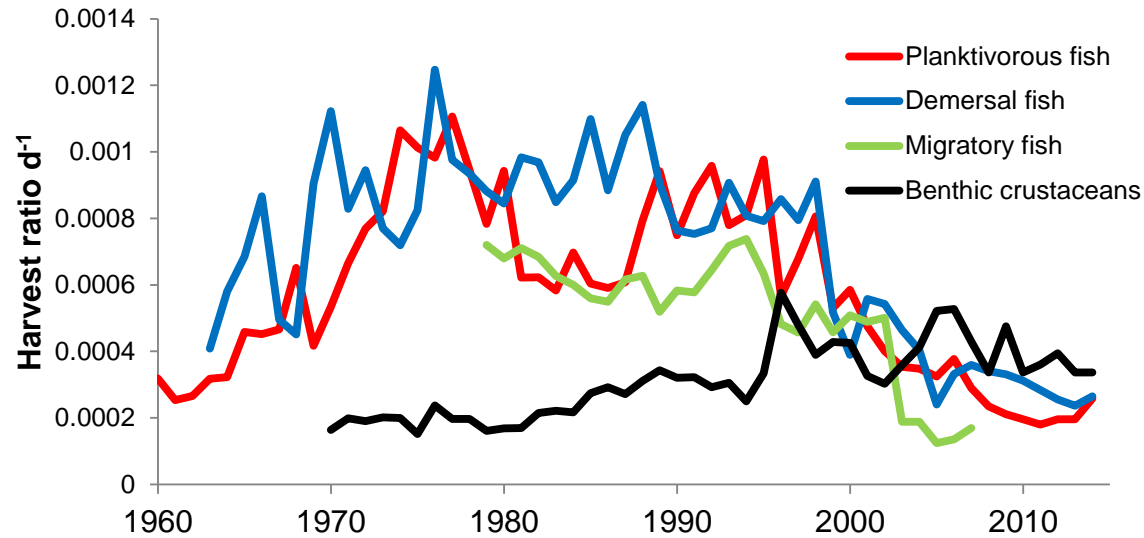


FIGURE 7 Time series of harvest ratio (proportion of biomass removed per day) for guilds of fish and benthic crustaceans in the North Sea. Data compiled from analyses of stock assessments, trawl survey data and with reference to ICES (2016).

Parameterisation of selectivity and harvest ratios for top predators.

In addition to the power parameters and the effort-harvest ratio scaling parameters defining the selectivity of each gear for fish and invertebrate guilds in the model, as outlined above, we also require equivalent parameters defining the unintended by-catch of the top-predator guilds (birds, pinnipeds, cetaceans) by these gears. There is no one simple source of data, equivalent to STECF, from which these parameters can be calculated, so we drew on data from a variety of sources:

First, we required data on the biomass of each top-predator guild in the North Sea, which we sourced from:

- Atlas of bird and cetacean species spatial abundances developed by statistical modelling of observer line-survey data on seabirds-at-sea and cetacean abundances (pers.comm, Dr James Waggitt & Dr Peter Evans, Bangor University; Waggitt *et al.*, 2019)
- Periodic assessments of grey and common seal population numbers in UK and European waters (Sea Mammal Research Unit (SMRU) & Marine Scotland, 2017)

Data on by-catch rates were sourced as follows:

- Records of strandings of cetaceans around the UK including pathology data on the likely cause of death (Deaville & Jepson, 2011). These data were used to identify cetaceans which had died as a result of entanglement in ropes – which we assumed to be predominantly creel lines.
- Synthesis of data on numbers of cetaceans entangled in fishing gears from national returns to ICES (e.g. ICES, 2015a,b, 2018), and summaries by ASCOBANS (Agreement on the Conservation of Small Cetaceans of the Baltic, North East Atlantic, Irish and North Seas, www.ascobans.org).
- Literature on cetacean by-catch (Bjorge *et al.*, 1994, 2013; Brown *et al.*, 2013; Dawson *et al.*, 2013; Evans & Hinter, 2013; Hammond *et al.*, 2002; Kaschner, 2003; Larsen & Eigaard, 2014; Morizur *et al.*, 1999; Northridge and Hammond, 1999; Northridge *et al.*, 2005, 2010; Pierce *et al.*, 2010; Read *et al.*, 2006; Ryan *et al.*, 2016; Vinther, 1995, 1999; Vinther and Larsen, 2004)
- Literature on seal by-catch (Cosgrove *et al.*, 2016)
- Literature on seabird by-catch (Anderson *et al.*, 2011; Genovart *et al.*, 2017; ICES, 2013b; Tasker *et al.*, 2000; Wiedenfeld *et al.*, 2012; Zydulis *et al.*, 2013)

Data on the activity rates of gears generating by-catch were as detailed in the preceding sections, based on the STECF database. The bird, pinniped and cetacean by-catch data to accompany these activity rates were assembled almost entirely from partial estimates for national fleet segments of particular gears, from a range of regions (not just the North Sea). For each ICES/FAO statistical region in the NE Atlantic where we were able to locate such data, we extracted the population biomass, the activity rate of the particular national fleet, and the corresponding species by-catch. From these data, we calculated a partial harvest ratio, and an activity density (activity per unit area) for the fleet segment. Finally, we combined all the individual records for a given species (partial harvest ratio and activity density) as a scatter-plot and fitted a linear regression forced through the origin (0,0). The slope of this regression represents the scaling coefficient between activity density and harvest ratio, which we assume to be fixed over time and regions. Finally, the individual species coefficients, harvest ratios and by-catch rates were aggregated up to guild-level, weighted by the estimates of species biomass in the North Sea (Tables 23-25). In effect, use of these activity-harvest ratio scaling parameters in the fishing fleet model assumes that each gear has a notional power of 1.0 for each vulnerable guild of top-predators.

TABLE 23 Fishing gears for which there are quantitative data on by-catch weights of particular species of top-predators, together with North Sea guild-aggregated scaling parameters linking regionally averaged activity density ($\text{sec.m}^{-2}.\text{d}^{-1}$) and regional harvest ratio (d^{-1}).

| Gear | Vulnerable seabird species | Vulnerable pinniped species | Vulnerable cetacean species | Seabird guild scaling parameter | Pinniped guild scaling parameter | Cetacean guild scaling parameter |
|-------------------|--------------------------------------|-----------------------------|---|---------------------------------|----------------------------------|----------------------------------|
| Demersal gillnets | Guillemot, razorbill, fulmar, gannet | Grey seal | Common dolphin, striped dolphin, harbour porpoise | 0.011 | 0.750 | 1.812 |
| Pelagic | Gannet | | Common dolphin, | 0.175 | | 0.351 |

| | | | | | | |
|-------------------|--------|--|--|-------|--|-------|
| trawl and seine | | | bottlenose dolphin, striped dolphin, pilot whale | | | |
| Pelagic longlines | Fulmar | | | 0.085 | | |
| Creels & pots | | | Fin whale, Minke whale | | | 0.033 |

TABLE 24 2003-2013 annual average partial harvest ratio (d^{-1}), of the three top-predator guilds in the North Sea by each of the relevant fishing gear groups in the model.

| Gear | Seabird harvest ratio | Pinniped harvest ratio | Cetacean harvest ratio |
|-----------------------|-----------------------|------------------------|------------------------|
| Demersal gillnets | 7.07×10^{-8} | 4.97×10^{-6} | 1.20×10^{-5} |
| Pelagic trawl & seine | 4.02×10^{-7} | | 8.40×10^{-8} |
| Pelagic longlines | 1.21×10^{-7} | | |
| Creels & pots | | | 8.33×10^{-7} |

TABLE 25 2003-2013 annual by-catch rates ($mMN.m^{-2}.y^{-1}$) of the three top-predator guilds in the North Sea by all fishing gears combined.

| Seabird by-catch | Pinniped by-catch | Cetacean by-catch |
|-----------------------|-----------------------|-----------------------|
| 6.25×10^{-7} | 2.74×10^{-5} | 2.75×10^{-4} |

Clearly, the methodology outlined above is an approximation since, for example, it disregards the relative spatial distributions of animals and fishing gear activity within the North Sea and how this many vary seasonally. Also, the national by-catch data are almost certainly partial. However, we do not regard the results as a definitive study, merely a pragmatic approach which is sufficient for the purposes of parameterising the coarse-scale StrathE2E2 model.

Directed whaling catch in the North Sea

In addition to the unintended by-catch of cetaceans by fishing gears, there is a small-scale targeted catch of Minke whales in the Norwegian sector of the region under objection to the International Whaling Commission zero catch limits. We obtained spatially resolved (1° longitude \times $\frac{1}{2}^\circ$ latitude) annual catch weights during 2003-2013 from the Norwegian Directorate of Fisheries and, using the Minke whale population biomass data as outlined above in the description of by-catch estimation, we estimated the annual harvest ratio. In order to represent this whaling activity in the model, we designated an additional fishing gear “Norwegian whalers”, which has no interaction with any other aspect of

the ecosystem except the cetacean guild. This meant that we could assign this fleet a notional activity density during 2003-2013 and a power of 1.0, and derive the scaling parameter linking activity density of the whalers to the harvest ratio (Table 26).

TABLE 26 2003-2013 annual average Minke whale by-catch rate in the North Sea, harvest ratio, and the derived parameter linking activity density of “Norwegian whalers” to harvest ratio assuming a notional activity density of 1000 hours per year.

| | |
|---|--|
| 2003-2013 annual average Minke whale catch | 67 tonnes = $8.38 \times 10^{-5} \text{ mMN.m}^{-2}.\text{y}^{-1}$ |
| Harvest ratio (d^{-1}) | 4.17×10^{-6} |
| Scaling parameter linking activity density ($\text{s.m}^{-2}.\text{d}^{-1}$) to harvest ratio (d^{-1}) | 210.31 |

We did not have access to annual Minke whale catch data prior to 2003, so we roughly estimated the catches in earlier years by assuming that North Sea catches have varied in proportion to the total Norwegian catch in the NE Atlantic (1986-onwards data compiled from the IWC www.iwc.int/table_objection). On this basis, the 1986-1999 average catch in the North Sea was estimated to be 32 tonnes per year.

Parameterisation of quota-limited/non-quota species composition of demersal fish catches

For the North Sea, the empirical evidence for density dependent relationships describing catch and discard composition comes from analysis of catch per unit effort data in research vessel trawl surveys carried out in quarter 1 of each year since 1980, and the corresponding species composition of annual commercial landings and discards (Heath & Cook, 2015). The analysis shows that at the scale of the whole North Sea the proportion of non-quota demersal fish species in the commercial catch has been indirectly related to the community biomass (Figure 8). There may be a number of explanations for this, but most likely is that depletion of the community biomass reflects the selective targeting of the valuable quota-limited species by the fisheries. In the model, we represent this relationship by a negative exponential function.

$$p_{(non-quota)} = a_{pnq} \cdot \exp(-b_{pnq} \cdot N_{dem.fish}) \quad \text{eqn 9}$$

where b_{pnq} is a scaling parameter, and ($N_{dem.fish}$) is the survey-based demersal fish biomass per unit swept area (mMN.m^{-2}), as measured on 1st January. The survey data were converted to nitrogen units by applying species-specific wet-weight to length relationships to the individual species number density-at-length data, and summing over all demersal species (Heath & Cook, 2015). Then, we assumed a nominal catching efficiency for the survey gear of 20% (Fraser *et al.*, 2007), and a wet-weight to nitrogen conversion of $1.34 \text{ mMN.g-WW}^{-1}$

Capture efficiency of the survey trawl is only approximately known, so to facilitate incorporation of this relationship in the model we included a proportionality constant (ϕ) to relate survey catch per unit swept area to nitrogen mass per unit sea surface area ($M_{dem.fish}$) as simulated in the model:

$$p_{(non-quota)} = a_{pnq} \cdot \exp(-b_{pnq} \cdot \phi \cdot M_{dem.fish}) \quad \text{eqn 10}$$

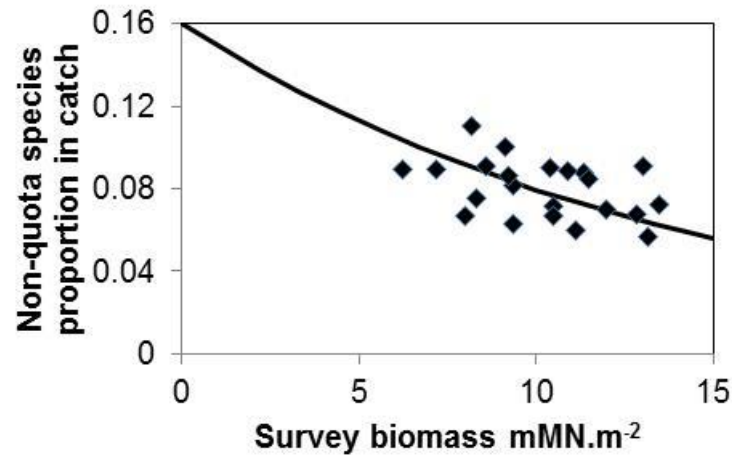


FIGURE 8 Proportion by weight of non-quota species in commercial catches from the North Sea, 1989-2010, in relation to the total biomass density of the demersal fish community as estimated in the corresponding year by the ICES IBTS quarter 1 surveys. Fitted equation: $p_{(non-quota)} = a_{pnq} \cdot \exp(-b_{pnq} \cdot N_{dem.fish})$, $a_{pnq} = 0.16$, $b_{pnq} = 0.07$; $p < 0.05$.

Parameterisation of the proportion of demersal fish catch which is smaller than the legal or *de-facto* marketable landing size.

The prototype version of StrathE2E2 included an empirically parameterised relationship between the proportion of demersal fish in commercial catches which were discarded on account of being undersize, and the biomass of demersal fish in the sea. The relationship expressed an exponentially declining discard rate with increasing biomass:

$$p_{(discarded)} = a_{disc} \cdot \exp(-b_{disc} \cdot M_{dem.fish}) \quad \text{eqn 11}$$

The general form of this relationship is retained in the new version of the model, but separate parameters are needed for the quota-limited and non-quota fractions of the demersal fish catch.

Technically, there is no minimum legal landing size for non-quota species. However, there is a de-facto minimum marketable size, below which there is no incentive to land the fish. Combined analysis of the 1980-2010 North Sea survey, landings and discards data (Heath & Cook, 2015) has shown that the average proportion by weight of non-quota species in the commercial catches which is below the marketable size, is approximately double the corresponding proportion of quota-limited species smaller than the minimum legal landing size. In addition, for both groups, these proportions have varied in inverse relation to demersal fish community biomass (Figure 9). The explanation for these density dependent relationships lies in the observed decrease in mean body size of demersal fish with declining community biomass. This is typically summarised for ecosystem assessment purposes by the Large Fish Indicator (LFI) which, in the North Sea, is defined as the proportion by weight of fish in the community which are larger than 40cm in length (Greenstreet *et al.*, 2011; Modica *et al.*, 2014).

The data from the North Sea showed that the exponents, or slopes, of the negative exponential relationships defining the proportion by weight of fish smaller than the minimum landing size (corresponding roughly to the historic discard rates) were not significantly different between the quota-limited and non-quota groups. However, the intercept term is substantially higher for the non-quota group.

$$p_{(undersize)Q} = a_{undersizeQ} \cdot \exp(-b_{undersizeQ} \cdot N_{dem.fish}) \text{ (for quota limited catch)} \quad \text{eqn 12}$$

$$p_{(undersize)NQ} = a_{undersizeNQ} \cdot \exp(-b_{undersizeNQ} \cdot N_{dem.fish}) \text{ (for non-quota limited catch)} \quad \text{eqn 13}$$

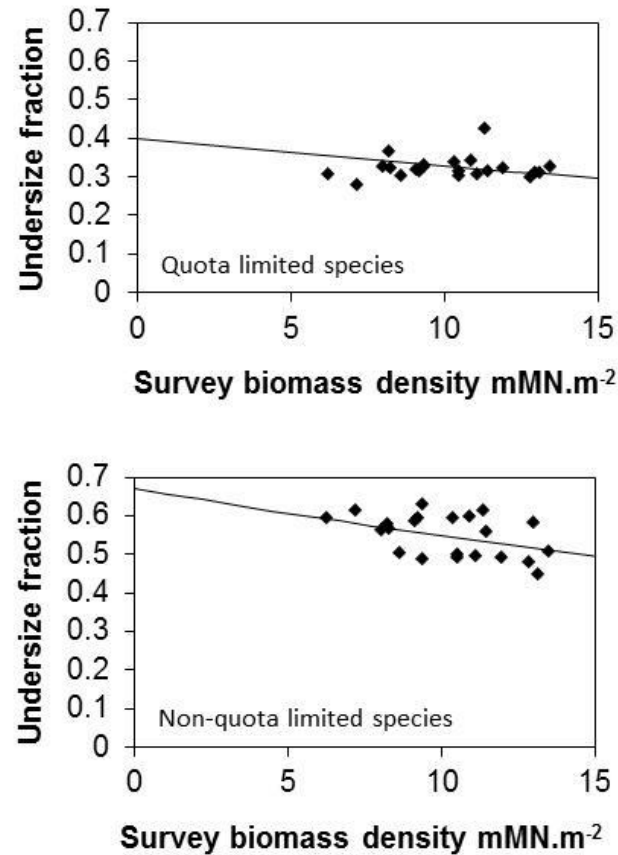


FIGURE 9 Proportions by weight of under minimum or marketable landing size demersal fish in the commercial catch from the North Sea, 1989-2010, in relation to the total biomass density of the demersal fish community as estimated in the corresponding year by the ICES IBTS quarter 1 surveys. Upper panel, quota-limited species, lower panel, non-quota species. Fitted equations of the form: $p_{(undersize)} = a_{undersize} \cdot \exp(-b_{undersize} \cdot N_{dem.fish})$. Separate fits for the quota and non-quota species showed that the coefficients $b_{undersize}$ were not significantly different ($p < 0.05$). Refitting assuming a common value for this parameter resulted in, for quota-limited species, $a_{(undersize)Q} = 0.40$, $b_{(undersize)Q} = 0.02$; and for non-quota species, $a_{(undersize)NQ} = 0.67$, $b_{(undersize)NQ} = 0.02$.

As for the imputation of non-quota fraction in the catch, we include the same scaling coefficient linking the survey catch per unit swept area and the biomass density in the model

$$N_{dem.fish} = \varphi \cdot M_{dem.fish} \quad \text{eqn 14}$$

Observational data for model fitting

Observational data on conditions in the North Sea during the periods 1970-1999 and 2003-2013 were assembled from a range of literature and data analyses (Table 27, 28). The data assembly by Mackinson & Daskalov (2007) was a key source of information for 1970-1999. In each case the information was such that an equivalent measure could be derived for comparison from the final year of a run to stationary state of the model.

TABLE 27 Observational data on the conditions in the North Sea relevant to the period 1970-1999, or a general value where no period-specific data were available. The standard deviation of the observed data was in some cases based on an actual analysis of multi-year data (e.g. in the case of fishery landings). In other case the standard deviation was a rough estimate based on very few data, or just a scaled value relative to the mean to assign a weighting to a particular measure in the likelihood calculation.

| Description | Sources | Mean value | s.d. of value | Units | Notes |
|--|---|------------|---------------|--------------------------------------|-----------------|
| Annual total primary production | Skogen & Moll (2005) | 1522 | 150.94 | mMN.m ⁻² .y ⁻¹ | Period-specific |
| Annual new production from drawdown of depth integrated nitrate plus summer river and atmospheric nitrate inputs | Heath & Beare (2008) | 624.4 | 66.4 | mMN.m ⁻² .y ⁻¹ | Period-specific |
| Annual within forest net production of kelp | Burrows <i>et al.</i> (2018) | 600 | 100 | gC.m ⁻² .y ⁻¹ | General value |
| Annual omnivorous zooplankton gross production | Heath (2005); Mackinson & Daskalov (2007) | 339.6 | 25.16 | mMN.m ⁻² .y ⁻¹ | Period-specific |
| Annual carnivorous zooplankton gross production | Heath (2005); Mackinson & Daskalov (2007) | 44.35 | 2.516 | mMN.m ⁻² .y ⁻¹ | Period-specific |
| Annual planktivorous fish gross production | Heath (2005) | 29.97 | 3.509 | mMN.m ⁻² .y ⁻¹ | Period-specific |

| | | | | | |
|--|---|-----------|----------|-----------------------------------|-----------------|
| Annual demersal fish gross production | Heath (2005) | 11.5 | 2.277 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | Period-specific |
| Annual suspension/deposit feeding benthos gross production | Eleftheriou & Basford (1989); Greenstreet <i>et al.</i> (2007); Heip & Craeymeersch (1995); Heip <i>et al.</i> (1984, 1989,1992); Kiinitzer <i>et al.</i> (1992); Mackinson & Daskalov (2007) | 1248 | 449 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | Period-specific |
| Annual carnivore/scavenge feeding benthos gross production | Eleftheriou & Basford (1989); Greenstreet <i>et al.</i> (2007); Heip & Craeymeersch (1995); Heip <i>et al.</i> (1984, 1989,1992); Kiinitzer <i>et al.</i> (1992); Mackinson & Daskalov (2007) | 21.1 | 7.6 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | Period-specific |
| Annual net production of birds | Mackinson & Daskalov (2007) | 8.452E-04 | 2.00E-04 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | Period-specific |
| Annual net production of pinnipeds | Mackinson & Daskalov (2007) | 7.245E-04 | 3.50e-04 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | Period-specific |
| Annual net production of cetaceans | Mackinson & Daskalov (2007) | 1.691E-03 | 8.00E-04 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | Period-specific |
| Annual monthly max concentration of benthos suspension/deposit feeder larvae | Lindley & Kirby (2007); Analysis of Continuous Plankton Recorder data | 1.185 | 0.4421 | mMN.m^{-3} | Period-specific |
| Annual monthly max concentration of benthos carnivore/scavenge feeder larvae | Lindley & Kirby (2007); Analysis of Continuous Plankton Recorder data | 0.334 | 0.1013 | mMN.m^{-3} | Period-specific |
| Annual consumption of planktivorous fish by fish | Heath (2005); Mackinson & Daskalov (2007) | 23.48 | 9.057 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | Period-specific |
| Annual consumption of demersal fish by fish | Heath (2005); Mackinson & Daskalov (2007) | 2.138 | 0.503 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | Period-specific |

| | | | | | |
|---|--|---------|--------|-----------------------------------|-----------------|
| Annual consumption of omnivorous zooplankton by fish and fish larvae | Heath (2005, 2007b); Mackinson & Daskalov (2007) | 92.28 | 13.019 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | Period-specific |
| Annual consumption of omnivorous zooplankton by carnivorous zooplankton | Heath (2005); Mackinson & Daskalov (2007) | 60.38 | 25.157 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | Period-specific |
| Annual consumption of benthos by fish | Heath (2005); Mackinson & Daskalov (2007) | 12.58 | 6.289 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | Period-specific |
| Annual food consumption by birds | Bryant & Doyle (1992); Mackinson & Daskalov (2007) | 0.6538 | 0.325 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | Period-specific |
| Proportion planktivorous fish in diet of birds | Bryant & Doyle (1992); Mackinson & Daskalov (2007) | 0.6 | 0.2 | dimensionless | Period-specific |
| Proportion demersal fish in diet of birds | Bryant & Doyle (1992); Mackinson & Daskalov (2007) | 0.1 | 0.05 | dimensionless | Period-specific |
| Proportion migratory fish in diet of birds | Bryant & Doyle (1992); Mackinson & Daskalov (2007) | 0.05 | 0.015 | dimensionless | Period-specific |
| Proportion discards in diet of birds | Bryant & Doyle (1992); Mackinson & Daskalov (2007) | 0.05 | 0.02 | dimensionless | Period-specific |
| Annual food consumption by pinnipeds | Mackinson & Daskalov (2007) | 0.2161 | 0.105 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | Period-specific |
| Proportion pelagic fish in diet of pinnipeds | Mackinson & Daskalov (2007) | 0.2910 | 0.0728 | dimensionless | Period-specific |
| Proportion demersal fish in diet of pinnipeds | Mackinson & Daskalov (2007) | 0.6969 | 0.1742 | dimensionless | Period-specific |
| Proportion migratory fish in diet of pinnipeds | Mackinson & Daskalov (2007) | 0.01208 | 0.0030 | dimensionless | Period-specific |
| Annual food consumption by cetaceans | Mackinson & Daskalov (2007) | 0.9691 | 0.48 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | Period-specific |
| Proportion pelagic fish in diet of cetaceans | Mackinson & Daskalov (2007); Olsen & Holst (2001) | 0.6632 | 0.1658 | dimensionless | Period-specific |

| | | | | | |
|--|---|---------|----------|--------------------------------------|-----------------|
| Proportion demersal fish in diet of cetaceans | Mackinson & Daskalov (2007); Olsen & Holst (2001) | 0.0995 | 0.04 | dimensionless | Period-specific |
| Proportion migratory fish in diet of cetaceans | Mackinson & Daskalov (2007); Olsen & Holst (2001) | 0.08014 | 0.035 | dimensionless | Period-specific |
| Annual planktivorous fish landings (live weight) | Analysis of EuroSTAT ICES landings data (Lassen <i>et al.</i> , 2012) | 5.555 | 0.2 | mMN.m ⁻² .y ⁻¹ | Period-specific |
| Annual demersal fish landings (live weight) | Analysis of EuroSTAT ICES landings data (Lassen <i>et al.</i> , 2012) | 1.735 | 0.08 | mMN.m ⁻² .y ⁻¹ | Period-specific |
| Annual migratory fish landings (live weight) | Analysis of EuroSTAT ICES landings data (Lassen <i>et al.</i> , 2012) | 0.775 | 0.308 | mMN.m ⁻² .y ⁻¹ | Period-specific |
| Annual suspension/deposit feeding benthos landings (live weight) | Analysis of EuroSTAT ICES landings data (Lassen <i>et al.</i> , 2012) | 0.0953 | 0.0382 | mMN.m ⁻² .y ⁻¹ | Period-specific |
| Annual carnivore/scavenge feeding benthos landings (live weight) | Analysis of EuroSTAT ICES landings data (Lassen <i>et al.</i> , 2012) | 0.0829 | 0.0169 | mMN.m ⁻² .y ⁻¹ | Period-specific |
| Annual carnivorous zooplankton landings (live weight) | Analysis of EuroSTAT ICES landings data (Lassen <i>et al.</i> , 2012) | 0.00147 | 9.32E-04 | mMN.m ⁻² .y ⁻¹ | Period-specific |
| Annual kelp landings (live weight) | Analysis of EuroSTAT ICES landings data (Lassen <i>et al.</i> , 2012) | 0 | 0 | mMN.m ⁻² .y ⁻¹ | Period-specific |
| Annual carbon gross PB ratio of kelp | Brady-Campbell <i>et al.</i> (1984) | 2 | 0.5 | y ⁻¹ | General value |
| Annual gross PB ratio larvae of suspension/deposit feeding benthos | Mackinson & Daskalov (2007) | 10 | 5 | y ⁻¹ | General value |
| Annual gross PB ratio larvae of carnivore/scavenge feeding benthos | Mackinson & Daskalov (2007) | 10 | 5 | y ⁻¹ | General value |
| Annual gross PB ratio suspension/deposit feeding benthos | Mackinson & Daskalov (2007) | 10 | 3 | y ⁻¹ | General value |

| | | | | | |
|---|---|-------|-------|--------------------------------------|-----------------|
| Annual gross PB ratio carnivore/scavenge feeding benthos | Mackinson & Daskalov (2007) | 1.2 | 1 | y ⁻¹ | General value |
| Annual gross PB ratio omnivorous zooplankton | Mackinson & Daskalov (2007) | 20 | 10 | y ⁻¹ | General value |
| Annual gross PB ratio carnivorous zooplankton | Mackinson & Daskalov (2007) | 5 | 1.315 | y ⁻¹ | General value |
| Annual gross PB ratio larvae of planktivorous fish | Mackinson & Daskalov (2007) | 4 | 2 | y ⁻¹ | General value |
| Annual gross PB ratio larvae of demersal fish | Mackinson & Daskalov (2007) | 4 | 2 | y ⁻¹ | General value |
| Annual gross PB ratio planktivorous fish | Mackinson & Daskalov (2007) | 1.72 | 0.86 | y ⁻¹ | General value |
| Annual gross PB ratio demersal fish | Mackinson & Daskalov (2007) | 0.88 | 0.44 | y ⁻¹ | General value |
| Annual gross PB ratio migratory fish | Mackinson & Daskalov (2007) | 1.3 | 0.6 | y ⁻¹ | General value |
| Annual net PB ratio birds | Mackinson & Daskalov (2007) | 0.28 | 0.14 | y ⁻¹ | General value |
| Annual net PB ratio pinnipeds | Mackinson and Daskalov 2007 | 0.09 | 0.045 | y ⁻¹ | General value |
| Annual net PB ratio cetaceans | Mackinson & Daskalov (2007) | 0.02 | 0.01 | y ⁻¹ | General value |
| Annual average proportion of kelp C uptake which is exuded | Abdullah & Fredriksen (2004) | 0.3 | 0.1 | dimensionless | General value |
| Annual average molar NC ratio of kelp | Broch & Slagstad (2012); Sjtun <i>et al.</i> (1996) | 0.12 | 0.2 | dimensionless | General value |
| Annual denitrification | Brion <i>et al.</i> (2004) | 129 | 42 | mMN.m ⁻² .y ⁻¹ | Period-specific |
| Proportion of demersal fish catch discarded | Heath & Cook (2015) | 0.37 | 0.075 | dimensionless | Period-specific |
| Annual average ammonia concentration in porewater of sand grain size 0.25mm | Serpetti (2012); Serpetti <i>et al.</i> (2016) | 19.24 | 9 | mMN.m ⁻³ | General value |
| Annual average ammonia concentration in porewater of mud grain size 0.12mm | Serpetti (2012); Serpetti <i>et al.</i> (2016) | 63.45 | 22 | mMN.m ⁻³ | General value |

| | | | | | |
|---|--|---------|---------|------------------------------------|-----------------|
| Annual average nitrate concentration in porewater of sand grain size 0.25mm | Serpetti (2012); Serpetti <i>et al.</i> (2016) | 4.15 | 2 | mMN.m ⁻³ | General value |
| Annual average nitrate concentration in porewater of mud grain size 0.12mm | Serpetti (2012); Serpetti <i>et al.</i> (2016) | 2.34 | 1 | mMN.m ⁻³ | General value |
| Annual average organic N content of sand grain size 0.25mm (0.19-0.43mm) | Serpetti (2012); Serpetti <i>et al.</i> (2016) | 0.05152 | 0.02441 | %N (gN.(g dry sed) ⁻¹) | General value |
| Annual average organic N content of mud grain size 0.12mm (0.03-0.07mm) | Serpetti (2012); Serpetti <i>et al.</i> (2016) | 0.07357 | 0.0343 | %N (gN.(g dry sed) ⁻¹) | General value |
| Average winter (Nov-Feb) nitrate concentration shallow layer | Analysis of ICES hydro-chemical data | 9.998 | 2.135 | mMN.m ⁻³ | Period-specific |
| Average summer (May-Aug) nitrate concentration shallow layer | Analysis of ICES hydro-chemical data | 2.161 | 1.089 | mMN.m ⁻³ | Period-specific |
| Average winter (Nov-Feb) nitrate concentration deep layer | Analysis of ICES hydro-chemical data | 6.995 | 0.836 | mMN.m ⁻³ | Period-specific |
| Average summer (May-Aug) nitrate concentration deep layer | Analysis of ICES hydro-chemical data | 2.837 | 0.917 | mMN.m ⁻³ | Period-specific |
| Average winter (Nov-Feb) ammonia concentration shallow layer | Analysis of ICES hydro-chemical data | 2.367 | 0.774 | mMN.m ⁻³ | Period-specific |
| Average summer (May-Aug) ammonia concentration shallow layer | Analysis of ICES hydro-chemical data | 1.737 | 0.669 | mMN.m ⁻³ | Period-specific |
| Average winter (Nov-Feb) ammonia concentration deep layer | Analysis of ICES hydro-chemical data | 0.853 | 0.32 | mMN.m ⁻³ | Period-specific |
| Average summer (May-Aug) ammonia concentration deep layer | Analysis of ICES hydro-chemical data | 1.338 | 0.708 | mMN.m ⁻³ | Period-specific |
| Inshore offshore ratio of annual mean carnivorous zooplankton | Analysis of Continuous Plankton Recorder data | 0.9041 | 0.2 | dimensionless | Period-specific |

| | | | | | |
|---|--|----------|----------|--------------------------------------|-----------------|
| depth averaged concentration | | | | | |
| Inshore offshore ratio of annual mean omnivorous zooplankton depth averaged concentration | Analysis of Continuous Plankton Recorder data | 1.676 | 0.2715 | dimensionless | Period-specific |
| Inshore offshore ratio of annual mean phytoplankton surface layer concentration | Analysis of ICES hydro-chemical data | 3.744 | 1 | dimensionless | Period-specific |
| Inshore offshore ratio of annual mean nitrate surface layer concentration | Analysis of ICES hydro-chemical data | 3.000 | 1 | dimensionless | Period-specific |
| Inshore offshore ratio of annual mean ammonia surface layer concentration | Analysis of ICES hydro-chemical data | 2.4294 | 0.9 | dimensionless | Period-specific |
| Inshore offshore ratio of annual mean planktivorous fish density (m ⁻²) | Analysis of ICES International Bottom Trawl Survey data and Stock Assessment Working Group reports | 0.67 | 0.49 | dimensionless | Period-specific |
| Inshore offshore ratio of annual mean demersal fish density (m ⁻²) | Analysis of ICES International Bottom Trawl Survey data and Stock Assessment Working Group reports | 0.39 | 0.31 | dimensionless | Period-specific |
| Annual bycatch of birds | Insufficient data available | NA | NA | mMN.m ⁻² .y ⁻¹ | NA |
| Annual bycatch of pinnipeds | Insufficient data available | NA | NA | mMN.m ⁻² .y ⁻¹ | NA |
| Annual bycatch of cetaceans | Insufficient data available | NA | NA | mMN.m ⁻² .y ⁻¹ | NA |
| Proportion of kelp annual nitrogen uptake exported as beach-cast | Zemke-White <i>et al.</i> (2005) | 0.15 | 0.05 | dimensionless | General value |
| Cetacean (Minke whale) catch | Analysis of data from Norwegian Directorate of Fisheries and International Whaling Commission | 4.05E-05 | 2.00E-05 | mMN.m ⁻² .y ⁻¹ | Period-specific |

TABLE 28 Observational data on the conditions in the North Sea relevant to the period 2003-2013, or a general value where no period-specific data were available. The standard deviation of the observed data was in some cases based on an actual analysis of multi-year data (e.g. in the case of fishery landings). In other case the standard deviation was a rough estimate based on very few data, or just a scaled value relative to the mean to assign a weighting to a particular measure in the likelihood calculation. For many fields, no period specific data were available for 2003-2013 or have not yet been processed for use in the model.

| Description | Sources | Mean value | s.d. of value | Units | Notes |
|--|---|------------|---------------|--------------------------------------|---------------------------|
| Annual total primary production | Skogen & Moll (2005) | 1522 | 150.94 | mMN.m ⁻² .y ⁻¹ | General value |
| Annual new production from drawdown of depth integrated nitrat, plus summer river and atmospheric nitrate inputs | Heath & Beare (2008) | 672.8 | 73.0 | mMN.m ⁻² .y ⁻¹ | Period-specific (partial) |
| Annual within forest net production of kelp | Burrows <i>et al.</i> (2018) | 600 | 100 | gC.m ⁻² .y ⁻¹ | General value |
| Annual omnivorous zooplankton gross production | Heath (2005); Mackinson & Daskalov (2007) | 339.6 | 25.157 | mMN.m ⁻² .y ⁻¹ | Period-specific (partial) |
| Annual carnivorous zooplankton gross production | NA | NA | NA | mMN.m ⁻² .y ⁻¹ | Insufficient data |
| Annual planktivorous fish gross production | NA | NA | NA | mMN.m ⁻² .y ⁻¹ | Insufficient data |
| Annual demersal fish gross production | NA | NA | NA | mMN.m ⁻² .y ⁻¹ | Insufficient data |
| Annual suspension/deposit feeding benthos gross production | NA | NA | NA | mMN.m ⁻² .y ⁻¹ | Insufficient data |
| Annual carnivore/scavenge feeding benthos gross production | NA | NA | NA | mMN.m ⁻² .y ⁻¹ | Insufficient data |
| Annual net production of birds | NA | NA | NA | mMN.m ⁻² .y ⁻¹ | Insufficient data |
| Annual net production of pinnipeds | NA | NA | NA | mMN.m ⁻² .y ⁻¹ | Insufficient data |
| Annual net production of cetaceans | NA | NA | NA | mMN.m ⁻² .y ⁻¹ | Insufficient data |
| Annual monthly max concentration of benthos | NA | NA | NA | mMN.m ⁻³ | Insufficient data |

| | | | | | |
|--|----|----|----|---------------------|-------------------|
| suspension/deposit feeder larvae | | | | | |
| Annual monthly max concentration of benthos carnivore/scavenge feeder larvae | NA | NA | NA | $mMN.m^{-3}$ | Insufficient data |
| Annual consumption of planktivorous fish by fish | NA | NA | NA | $mMN.m^{-2}.y^{-1}$ | Insufficient data |
| Annual consumption of demersal fish by fish | NA | NA | NA | $mMN.m^{-2}.y^{-1}$ | Insufficient data |
| Annual consumption of omnivorous zooplankton by fish and fish larvae | NA | NA | NA | $mMN.m^{-2}.y^{-1}$ | Insufficient data |
| Annual consumption of omnivorous zooplankton by carnivorous zooplankton | NA | NA | NA | $mMN.m^{-2}.y^{-1}$ | Insufficient data |
| Annual consumption of benthos by fish | NA | NA | NA | $mMN.m^{-2}.y^{-1}$ | Insufficient data |
| Annual food consumption by birds | NA | NA | NA | $mMN.m^{-2}.y^{-1}$ | Insufficient data |
| Proportion planktivorous fish in diet of birds | NA | NA | NA | dimensionless | Insufficient data |
| Proportion demersal fish in diet of birds | NA | NA | NA | dimensionless | Insufficient data |
| Proportion migratory fish in diet of birds | NA | NA | NA | dimensionless | Insufficient data |
| Proportion discards in diet of birds | NA | NA | NA | dimensionless | Insufficient data |
| Annual food consumption by pinnipeds | NA | NA | NA | $mMN.m^{-2}.y^{-1}$ | Insufficient data |
| Proportion pelagic fish in diet of pinnipeds | NA | NA | NA | dimensionless | Insufficient data |
| Proportion demersal fish in diet of pinnipeds | NA | NA | NA | dimensionless | Insufficient data |

| | | | | | |
|--|---|----------|-------|-----------------------------------|-------------------|
| Proportion migratory fish in diet of pinnipeds | NA | NA | NA | dimensionless | Insufficient data |
| Annual food consumption by cetaceans | NA | NA | NA | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | Insufficient data |
| Proportion pelagic fish in diet of cetaceans | NA | NA | NA | dimensionless | Insufficient data |
| Proportion demersal fish in diet of cetaceans | NA | NA | NA | dimensionless | Insufficient data |
| Proportion migratory fish in diet of cetaceans | NA | NA | NA | dimensionless | Insufficient data |
| Annual planktivorous fish landings (live weight) | Analysis of EuroSTAT ICES landings data (Lassen <i>et al.</i> , 2012) | 2.928 | 0.6 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | Period-specific |
| Annual demersal fish landings (live weight) | Analysis of EuroSTAT ICES landings data (Lassen <i>et al.</i> , 2012) | 0.6408 | 0.3 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | Period-specific |
| Annual migratory fish landings (live weight) | Analysis of EuroSTAT ICES landings data (Lassen <i>et al.</i> , 2012) | 0.9925 | 0.25 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | Period-specific |
| Annual suspension/deposit feeding benthos landings (live weight) | Analysis of EuroSTAT ICES landings data (Lassen <i>et al.</i> , 2012) | 0.07444 | 0.028 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | Period-specific |
| Annual carnivore/scavenge feeding benthos landings (live weight) | Analysis of EuroSTAT ICES landings data (Lassen <i>et al.</i> , 2012) | 0.12801 | 0.05 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | Period-specific |
| Annual carnivorous zooplankton landings (live weight) | Analysis of EuroSTAT ICES landings data (Lassen <i>et al.</i> , 2012) | 0.006976 | 0.004 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | Period-specific |
| Annual kelp landings (live weight) | Analysis of EuroSTAT ICES landings data (Lassen <i>et al.</i> , 2012) | 0 | 0 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | Period-specific |
| Annual carbon gross PB ratio of kelp | Brady-Campbell <i>et al.</i> (1984) | 2 | 0.5 | y^{-1} | General value |
| Annual gross PB ratio larvae of suspension/deposit feeding benthos | Mackinson & Daskalov (2007) | 10 | 5 | y^{-1} | General value |

| | | | | | |
|--|--|------|-------|---------------------|-----------------|
| Annual gross PB ratio larvae of carnivore/scavenge feeding benthos | Mackinson & Daskalov (2007) | 10 | 5 | y^{-1} | General value |
| Annual gross PB ratio suspension /deposit feeding benthos | Mackinson & Daskalov (2007) | 10 | 3 | y^{-1} | General value |
| Annual gross PB ratio carnivore/scavenge feeding benthos | Mackinson & Daskalov (2007) | 1.2 | 1 | y^{-1} | General value |
| Annual gross PB ratio omnivorous zooplankton | Mackinson & Daskalov (2007) | 20 | 10 | y^{-1} | General value |
| Annual gross PB ratio carnivorous zooplankton | Mackinson & Daskalov (2007) | 5 | 1.315 | y^{-1} | General value |
| Annual gross PB ratio larvae of planktivorous fish | Mackinson & Daskalov (2007) | 4 | 2 | y^{-1} | General value |
| Annual gross PB ratio larvae of demersal fish | Mackinson & Daskalov (2007) | 4 | 2 | y^{-1} | General value |
| Annual gross PB ratio planktivorous fish | Mackinson & Daskalov (2007) | 1.72 | 0.86 | y^{-1} | General value |
| Annual gross PB ratio demersal fish | Mackinson & Daskalov (2007) | 0.88 | 0.44 | y^{-1} | General value |
| Annual gross PB ratio migratory fish | Mackinson & Daskalov (2007) | 1.3 | 0.6 | y^{-1} | General value |
| Annual net PB ratio birds | Mackinson & Daskalov (2007) | 0.28 | 0.14 | y^{-1} | General value |
| Annual net PB ratio pinnipeds | Mackinson & Daskalov (2007) | 0.09 | 0.045 | y^{-1} | General value |
| Annual net PB ratio cetaceans | Mackinson & Daskalov (2007) | 0.02 | 0.01 | y^{-1} | General value |
| Annual average proportion of kelp C uptake which is exuded | Abdullah & Fredriksen (2004) | 0.3 | 0.1 | dimensionless | General value |
| Annual average molar NC ratio of kelp | Broch & Slagstad (2012); Sjotun <i>et al.</i> (1996) | 0.12 | 0.2 | dimensionless | General value |
| Annual denitrification | Brion <i>et al.</i> (2004) | 129 | 42 | $mMN.m^{-2}.y^{-1}$ | Period-specific |
| Proportion of demersal fish catch discarded | Heath & Cook (2015) | 0.51 | 0.1 | dimensionless | Period-specific |

| | | | | | |
|---|--|----------|---------|------------------------------------|-------------------|
| Annual average ammonia concentration in porewater of sand grain size 0.25mm | Serpetti (2012); Serpetti <i>et al.</i> (2016) | 19.24 | 9 | mMN.m ⁻³ | General value |
| Annual average ammonia concentration in porewater of mud grain size 0.12mm | Serpetti (2012); Serpetti <i>et al.</i> (2016) | 63.45 | 22 | mMN.m ⁻³ | General value |
| Annual average nitrate concentration in porewater of sand grain size 0.25mm | Serpetti (2012); Serpetti <i>et al.</i> (2016) | 4.15 | 2 | mMN.m ⁻³ | General value |
| Annual average nitrate concentration in porewater of mud grain size 0.12mm | Serpetti (2012); Serpetti <i>et al.</i> (2016) | 2.34 | 1 | mMN.m ⁻³ | General value |
| Annual average organic N content of sand grain size 0.25mm (0.19-0.43mm) | Serpetti (2012); Serpetti <i>et al.</i> (2016) | 0.051515 | 0.02441 | %N (gN.(g dry sed) ⁻¹) | General value |
| Annual average organic N content of mud grain size 0.12mm (0.03-0.07mm) | Serpetti (2012); Serpetti <i>et al.</i> (2016) | 0.07357 | 0.0343 | %N (gN.(g dry sed) ⁻¹) | General value |
| Average winter (Nov-Feb) nitrate concentration shallow layer | NA | NA | NA | mMN.m ⁻³ | Insufficient data |
| Average summer (May-Aug) nitrate concentration shallow layer | NA | NA | NA | mMN.m ⁻³ | Insufficient data |
| Average winter (Nov-Feb) nitrate concentration deep layer | NA | NA | NA | mMN.m ⁻³ | Insufficient data |
| Average summer (May-Aug) nitrate concentration deep layer | NA | NA | NA | mMN.m ⁻³ | Insufficient data |
| Average winter (Nov-Feb) ammonia concentration shallow layer | NA | NA | NA | mMN.m ⁻³ | Insufficient data |
| Average summer (May-Aug) ammonia concentration shallow layer | NA | NA | NA | mMN.m ⁻³ | Insufficient data |
| Average winter (Nov-Feb) ammonia concentration deep | NA | NA | NA | mMN.m ⁻³ | Insufficient data |

| | | | | | |
|--|--|----------|----------|--------------------------------------|-------------------|
| layer | | | | | |
| Average summer (May-Aug) ammonia concentration deep layer | NA | NA | NA | mMN.m ⁻³ | Insufficient data |
| Inshore offshore ratio of annual mean carnivorous zooplankton depth averaged concentration | Analysis of Continuous Plankton Recorder data | 0.904055 | 0.2 | dimensionless | Period-specific |
| Inshore offshore ratio of annual mean omnivorous zooplankton depth averaged concentration | Analysis of Continuous Plankton Recorder data | 1.675727 | 0.271539 | dimensionless | Period-specific |
| Inshore offshore ratio of annual mean phytoplankton surface layer concentration | Analysis of ICES hydro-chemical data | 3.744403 | 1 | dimensionless | Period-specific |
| Inshore offshore ratio of annual mean nitrate surface layer concentration | Analysis of ICES hydro-chemical data | 3.000139 | 1 | dimensionless | Period-specific |
| Inshore offshore ratio of annual mean ammonia surface layer concentration | Analysis of ICES hydro-chemical data | 2.429353 | 0.9 | dimensionless | Period-specific |
| Inshore offshore ratio of annual mean planktivorous fish density (m ⁻²) | Analysis of ICES International Bottom Trawl Survey data and Stock Assessment Working Group reports | 0.67 | 0.49 | dimensionless | Period-specific |
| Inshore offshore ratio of annual mean demersal fish density (m ⁻²) | Analysis of ICES International Bottom Trawl Survey data and Stock Assessment Working Group reports | 0.39 | 0.31 | dimensionless | Period-specific |
| Annual bycatch of birds | Insufficient data available | 6.25E-07 | 3.00E-07 | mMN.m ⁻² .y ⁻¹ | NA |
| Annual bycatch of pinnipeds | Insufficient data available | 2.74E-05 | 5.00E-05 | mMN.m ⁻² .y ⁻¹ | NA |
| Annual bycatch of cetaceans | Insufficient data available | 0.000275 | 1.00E-04 | mMN.m ⁻² .y ⁻¹ | NA |
| Proportion of kelp annual nitrogen uptake exported as beach-cast | Zemke-White <i>et al.</i> (2005) | 0.15 | 0.05 | dimensionless | General value |

| | | | | | |
|------------------------------|---|----------|----------|--------------------------------------|-----------------|
| Cetacean (Minke whale) catch | Analysis of data from Norwegian Directorate of Fisheries and International Whaling Commission | 8.38E-05 | 4.00E-05 | mMN.m ⁻² .y ⁻¹ | Period-specific |
|------------------------------|---|----------|----------|--------------------------------------|-----------------|

Model fitting

The major unknown parameters of the model system are within ecology model, especially those related to the temperature-dependent physiological and behavioural functioning of the guilds and geochemical cycling processes. We expect these parameters to remain constant over time unless there have been extensive changes in species composition of the guilds. Since there is no realistic prospect of independently estimating most of these parameters we require to estimate them by fitting the stationary model to target data on the observed state of the system in a given time period. Constraining the ecology model parameters is clearly dependent on the diversity and quality of the target data, but also conditional on the quality of the information on environmental drivers and fishing fleet parameters

Among the fishing fleet parameters that are required, we can assume that the power parameters and seabed impact properties of the fishing gears remain constant over time. These are defining feature of each of the gear types which determine selectivity with respect to the harvestable guilds and seabed disturbance. Of course, it is likely that in reality there have been some changes in these gear properties, but we can assume that at the coarse taxonomic resolution of the model guilds these are likely to be small. Realistic changes in selectivity and seabed impact are probably at the scale of the species within each guild and hence not resolved by this model. Changes in selectivity and seabed impact at the scale of the model guilds would more likely justify the definition of different gear types.

Similarly, the scaling parameters linking effort to harvest ratios of each guild must be assumed to remain constant over time. To assume otherwise would risk confounding the translation of changes in fishing gear activity and distribution into changes in harvest ratio.

The remaining drivers and properties of the model which define the conditions that make one time period different from another, are the environmental conditions, fishing gear activity, distribution and discard rates, and the seasonal immigration flux of migratory fish from outside the system.

Unfortunately the data required to independently determine the defining properties of the fishing fleet model are not uniformly available over time in the North Sea (Table 29). Full data requirements were not satisfied for either of the time periods 1970-1999 and 2003-2013. As a result, we had to develop an iterative process to arrive at a parameter set which provided a credible fit to the observed target data for both time periods.

TABLE 29 Quality of key datasets needed for fitting of StrathE2E2 in the North Sea.

| Data type | 1970-1999 | 2003-2013 |
|---|--|---|
| Target data to which the model can be fitted | Comprehensive coverage of data fields as a consequence of large research programmes during the period | Many missing fields in the target data |
| Gear activity rates and spatial distributions | Unknown | Well known |
| Discard rates for each gear | Unknown | Well known |
| Power parameters for each gear | Unknown | Well known |
| Harvest ratios for ecology model guilds | Reasonably well known for fish based on long-standing stock assessments (ICES 2016). Less certain for invertebrates and top predators. | Reasonably well known based on long-standing stock assessments (ICES 2016), by-catch and strandings data. |
| Environmental driving data | Well known from NEMO-ERSEM outputs and World Ocean Atlas (Butenschön <i>et al.</i> 2016, Garcia <i>et al.</i> 2014) | Well known from NEMO-ERSEM outputs and World Ocean Atlas (Butenschön <i>et al.</i> 2016, Garcia <i>et al.</i> 2014) |

The starting point for the iterative scheme (Figure 10) relied on four ‘solid’ sets of information:

- 1) the environmental driving data for the two periods derived from the NEMO-ERSEM model outputs,
- 2) the 1970-1999 and 2003-2013 guild-level harvest ratios for fish and invertebrates,
- 3) the 2003-2013 gear activity and discard rate data derived from the STECF database,
- 4) the comprehensive assemblage of 1970-1999 target data on the ecological state of the system.

Gear activity rates for 1970-1999 were not available, so we estimated approximate values by up-scaling the well-established values for 2003-2013. The basis for the up-scaling was the synthesis of 60-year changes in functional guild aggregated fishing mortality rates in the North Sea compiled for the ICES Greater North Sea Eco-region review (ICES, 2016). The digitised data show that 1970-1999 averaged mortality rates for pelagic fish stocks were 1.513-times higher than during 2003-2013, demersal fish rates 1.616-times higher. On the other hand, mortality rates of benthic invertebrates were lower during 1970-1999 than in the more recent period, by a factor of 0.401. We therefore scaled the 2003-2013 activity densities of the pelagic-fish-targeting gears (pelagic trawls and seine, and mackerel longlines) by a factor of 1.513; the demersal-fish-targeting gears (demersal beam trawl, seine, otter trawl, longlines and gillnets) by 1.616; and the benthos-targeting gears (shrimp trawl, Nephrops trawl, creels and pots, and mollusc dredges) by 0.401. We adopted the proportional spatial distribution of activity by each gear from 2003-2013. This provided enough information to calculate initial values for the scaling parameters linking effort to harvest ratios using the function `e2e_calculate_hrscale()`.

The patterns of inshore and offshore harvest ratios for each of the ecology model guilds, given these parameters, represents an element of known driving data for the model. Based on these data we then obtained a first set of ecology model parameters by fitting to the target data, using the function **e2e_optimize_eco()**.

The next step was to focus on the 2003-2013 period. This time, we adopted the ecology model parameters from the previous step, the known activity, spatial distribution and discard rate data from STECF, and instead estimate the effort – harvest ratio scaling parameters required to produce the best fit to the 2003-2013 target data using the function **e2e_optimize_hr()**.

Stage 3 returned to the 1970-1999 period. This time we adopted the newly estimated effort – harvest ratio scaling parameters from stage 2, and focussed on finding the gear activity rates required to reproduce the harvest ratios for each ecology model guild which were independently derived prior to stage 1. This is not a trivial task since the gears had overlapping selectivity patterns for the model resource guilds. The process involved using the function **e2e_optimize_act()** to optimize the gear activity rates (but not their spatial distributions) so as to maximise the likelihood of the expected harvest ratios given the other parameters of the fishing fleet model. Some groups of gears were constrained to vary in concert to a degree, rather than completely independently. For example the subset of gears targeting demersal fish (demersal seine, demersal otter trawl, demersal gill nets and long-lines) were linked so that changes in their activities at each iteration of the annealing process were proportional to each other plus or minus some random variation.

Stage 4 was a repeat of stage 1 but using the gear activity rates from stage 3 and the effort – harvest ratio scaling parameters from stage 2. A further cycle through the loop of fitting stages produced minimal changes in the ecology model parameters, the effort – harvest ratio scaling parameters, or the estimates of 1970-1999 activity rates.

In the case of the whaling fleet in the North Sea, there were no data on activity rates, so a notional value was assigned for 2003-2013 when the whaling catch is known, and the effort – harvest ratio scaling parameter manually adjusted to achieve the target catch of cetaceans. This was entirely justified since the whaler catch is confined to cetaceans with no by-catch of other guilds.

For the 1970-1999 model the annealing process converged to a robust set of ecology model parameters within 11,000 iterations, with each iteration entailing a 50 year run for each proposed parameter set (stages 1, 4, 7 in the fitting scheme). Convergence was deemed to be attained when the system completed 200 iterations without finding any improvement in the likelihood. At convergence, the overall likelihood of the observed data given the parameters, driving data and the model structure was 0.421. For the 2003-2013 period the overall likelihood with the same ecology model parameters and effort – harvest ratio scaling values was 0.496 (but note that there were fewer target data available for the 2003-2013 period). All of the observed data values, the maximum likelihood model value, and the corresponding partial likelihoods, are given in Tables 32 and 33. Convergence of the fitting process for effort – harvest ratio scaling values (stages 2, 5) and 1970-1999 gear activity rates (stages 3, 6) was within 1000 iteration since far fewer parameters were involved in the process.

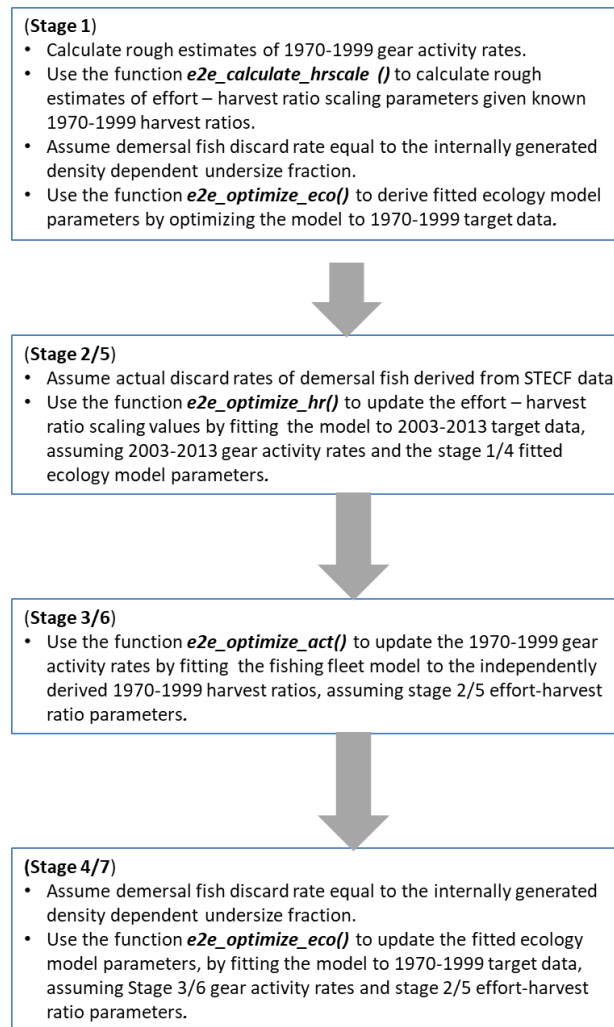


FIGURE 10 Workflow diagram for the scheme devised to fit the combined fleet and ecology model to the target data available for the two periods 1970-1999 and 2003-2013.

Results of the model fitting procedure

TABLE 30. North Sea domain-wide activity rates ($\text{s.m}^{-2}.\text{d}^{-1}$) for each gear during 2003-2013 derived from the STECF data, and estimates for 1970-1999 derived by fitting the combined fishing fleet and ecology model to the observed ecosystem data.

| Gear category | STECF activity 2003-2013 | Estimated activity 1970-1999 |
|--|--------------------------|------------------------------|
| Pelagic trawls & seines | 2.170E-06 | 6.828E-06 |
| Sandeel & sprat trawl (Otter30-70mm+TR3) | 4.230E-06 | 5.101E-06 |
| Longline mackerel | 1.680E-06 | 5.965E-06 |
| Beam trawls for demersal fish (BT1+BT2) | 1.150E-05 | 9.075E-05 |
| Demersal seine | 1.720E-08 | 1.395E-07 |
| Demersal otter trawl (TR1) | 2.160E-05 | 1.807E-04 |
| Gill nets & longlines for demersal fish | 7.920E-06 | 8.178E-05 |
| Beam trawl for shrimp | 1.270E-05 | 1.447E-05 |
| Nephrops trawl (TR2) | 1.720E-05 | 9.659E-06 |
| Creels | 2.400E-05 | 5.415E-06 |
| Mollusc dredge | 3.110E-06 | 5.822E-06 |
| Whaling vessels | 1.980E-08 | 1.243E-08 |

TABLE 31 Harvest ratios in the two periods (1970-1999 and 2003-2013) and scaling coefficients for each ecology model resource guild on conclusion of the fitting procedure.

| Ecology model resource guild | 1970-1999 inshore harvest ratio | 1970-1999 offshore harvest ratio | 2003-2013 inshore harvest ratio | 2003-2013 offshore harvest ratio | Effort to harvest ratio scaling coefficient |
|------------------------------------|---------------------------------|----------------------------------|---------------------------------|----------------------------------|---|
| Planktivorous fish | 9.640E-04 | 8.053E-04 | 7.116E-04 | 5.534E-04 | 0.069711 |
| Demersal fish | 1.502E-03 | 9.684E-04 | 3.075E-04 | 2.025E-04 | 0.075483 |
| Migratory fish | 2.272E-03 | 2.245E-03 | 1.063E-03 | 9.875E-04 | 0.369567 |
| Suspension/deposit feeding benthos | 4.094E-04 | 2.411E-04 | 3.320E-04 | 1.956E-04 | 12.66203 |
| Carnivore/scavenge feeding benthos | 7.448E-04 | 1.033E-04 | 9.568E-04 | 1.155E-04 | 0.630055 |
| Carnivorous zooplankton | 1.967E-03 | 3.077E-03 | 3.756E-04 | 6.034E-04 | 15.7164 |
| Birds | 7.301E-06 | 3.668E-06 | 2.013E-06 | 1.531E-06 | 2.717908 |

| | | | | | |
|-------------|-----------|-----------|-----------|-----------|----------|
| Pinnipeds | 3.206E-04 | 4.195E-05 | 4.751E-05 | 6.216E-06 | 2.781545 |
| Cetaceans | 1.672E-03 | 2.686E-04 | 7.861E-04 | 1.614E-04 | 18.10814 |
| Macrophytes | 0 | 0 | 0 | 0 | 1 |

TABLE 32 Observational indices of the 1970-1999 state of the North Sea ecosystem to which the model was fitted using 1970-1999 environmental drivers and fishing fleet inputs, Indices include fluxes between guilds, and annual ratios, and their standard deviations. For each index (*i*) the partial likelihood between observed and modelled values with parameter set θ , obtained from the simulated annealing scheme, was calculated as $\exp(-\chi_{\theta i}^2)$, where $\chi_{\theta i}^2 = \frac{(\text{observed}_i - \text{model}_{\theta,i})^2}{2\sigma_i^2}$ (σ_i is the standard deviation of observed index *i*), and the overall likelihood is given by $\exp(-(\text{mean}(\chi_{\theta}^2)))$. The results are shown graphically in Figure 11.

| Description | Observational value | s.d. of observational value | Units | Maximum likelihood model value | Partial likelihood |
|---|---------------------|-----------------------------|--------------------------------------|--------------------------------|--------------------|
| Annual total primary production | 1522 | 150.94 | mMN.m ⁻² .y ⁻¹ | 1216.76 | 0.129411 |
| Annual new production from depth integrated nitrate | 624.4 | 66.4 | mMN.m ⁻² .y ⁻¹ | 494.024 | 0.145445 |
| Annual within forest net production of kelp | 600 | 100 | gC.m ⁻² .y ⁻¹ | 597.316 | 0.99964 |
| Annual omnivorous zooplankton gross production | 339.6 | 25.16 | mMN.m ⁻² .y ⁻¹ | 320.8222 | 0.756862 |
| Annual carnivorous zooplankton gross production | 44.35 | 2.516 | mMN.m ⁻² .y ⁻¹ | 38.80518 | 0.088125 |
| Annual planktivorous fish gross production | 29.97 | 3.509 | mMN.m ⁻² .y ⁻¹ | 23.38265 | 0.171688 |
| Annual demersal fish gross production | 11.5 | 2.277 | mMN.m ⁻² .y ⁻¹ | 7.369128 | 0.192894 |
| Annual suspension /deposit feeding benthos gross production | 1248 | 449 | mMN.m ⁻² .y ⁻¹ | 598.3658 | 0.351101 |
| Annual carnivore/scavenge feeding benthos gross production | 21.1 | 7.6 | mMN.m ⁻² .y ⁻¹ | 20.68533 | 0.998513 |
| Annual net production of birds | 8.452E-04 | 2.00E-04 | mMN.m ⁻² .y ⁻¹ | 0.000878 | 0.986591 |
| Annual net production of pinnipeds | 7.245E-04 | 3.50e-04 | mMN.m ⁻² .y ⁻¹ | 0.000809 | 0.971189 |
| Annual net production of cetaceans | 1.691E-03 | 8.00E-04 | mMN.m ⁻² .y ⁻¹ | 0.001162 | 0.803953 |
| Annual monthly max concentration of benthos suspension /deposit feeder larvae | 1.185 | 0.4421 | mMN.m ⁻³ | 0.636286 | 0.462905 |
| Annual monthly max concentration of benthos carnivore/scavenge feeder larvae | 0.334 | 0.1013 | mMN.m ⁻³ | 0.288244 | 0.903021 |

| | | | | | |
|---|---------|----------|-----------------------------------|----------|----------|
| Annual consumption of planktivorous fish by fish | 23.48 | 9.057 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | 2.879647 | 0.075266 |
| Annual consumption of demersal fish by fish | 2.138 | 0.503 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | 1.102063 | 0.119935 |
| Annual consumption of omnivorous zooplankton by fish and fish larvae | 92.28 | 13.019 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | 79.08908 | 0.598522 |
| Annual consumption of omnivorous zooplankton by carnivorous zooplankton | 60.38 | 25.157 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | 106.8211 | 0.181924 |
| Annual consumption of benthos by fish | 12.58 | 6.289 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | 7.312845 | 0.704276 |
| Annual food consumption by birds | 0.6538 | 0.325 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | 0.099423 | 0.233439 |
| Proportion planktivorous fish in diet of birds | 0.6 | 0.2 | Dimensionless | 0.612792 | 0.997957 |
| Proportion demersal fish in diet of birds | 0.1 | 0.05 | Dimensionless | 0.146752 | 0.645879 |
| Proportion migratory fish in diet of birds | 0.05 | 0.015 | Dimensionless | 0.050157 | 0.999945 |
| Proportion discards in diet of birds | 0.05 | 0.02 | Dimensionless | 0.00681 | 0.097132 |
| Annual food consumption by pinnipeds | 0.2161 | 0.105 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | 0.139659 | 0.767279 |
| Proportion pelagic fish in diet of pinnipeds | 0.2910 | 0.0728 | Dimensionless | 0.320438 | 0.921604 |
| Proportion demersal fish in diet of pinnipeds | 0.6969 | 0.1742 | Dimensionless | 0.633588 | 0.936145 |
| Proportion migratory fish in diet of pinnipeds | 0.01208 | 0.0030 | Dimensionless | 0.010932 | 0.929773 |
| Annual food consumption by cetaceans | 0.9691 | 0.48 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | 0.48652 | 0.603313 |
| Proportion pelagic fish in diet of cetaceans | 0.6632 | 0.1658 | Dimensionless | 0.727068 | 0.92836 |
| Proportion demersal fish in diet of cetaceans | 0.0995 | 0.04 | Dimensionless | 0.123305 | 0.837755 |
| Proportion migratory fish in diet of cetaceans | 0.08014 | 0.035 | Dimensionless | 0.099164 | 0.862626 |
| Annual planktivorous fish landings (live weight) | 5.555 | 0.2 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | 5.648772 | 0.895909 |
| Annual demersal fish landings (live weight) | 1.735 | 0.08 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | 1.764646 | 0.933643 |
| Annual migratory fish landings (live weight) | 0.775 | 0.308 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | 0.778053 | 0.999951 |
| Annual suspension /deposit feeding benthos landings (live weight) | 0.0953 | 0.0382 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | 0.094009 | 0.99943 |
| Annual carnivore/scavenge feeding benthos landings (live weight) | 0.0829 | 0.0169 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | 0.05639 | 0.292193 |
| Annual carnivorous zooplankton landings (live weight) | 0.00147 | 9.32E-04 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | 0.002913 | 0.301576 |
| Annual kelp landings (live weight) | 0 | 0 | $\text{mMN.m}^{-2}.\text{y}^{-1}$ | 0 | NA |
| Annual carbon gross PB ratio of kelp | 2 | 0.5 | y^{-1} | 1.318367 | 0.394851 |
| Annual gross PB ratio larvae of suspension/deposit feeding benthos | 10 | 5 | y^{-1} | 15.49276 | 0.546944 |
| Annual gross PB ratio larvae of | 10 | 5 | y^{-1} | 15.86236 | 0.502909 |

| | | | | | |
|---|---------|---------|--------------------------------------|----------|----------|
| carnivore/scavenge feeding benthos | | | | | |
| Annual gross PB ratio suspension/deposit feeding benthos | 10 | 3 | y ⁻¹ | 11.94798 | 0.809925 |
| Annual gross PB ratio carnivore/scavenge feeding benthos | 1.2 | 1 | y ⁻¹ | 1.939656 | 0.760678 |
| Annual gross PB ratio omnivorous zooplankton | 20 | 10 | y ⁻¹ | 10.49988 | 0.636824 |
| Annual gross PB ratio carnivorous zooplankton | 5 | 1.315 | y ⁻¹ | 3.732141 | 0.628264 |
| Annual gross PB ratio larvae of planktivorous fish | 4 | 2 | y ⁻¹ | 0.651836 | 0.246283 |
| Annual gross PB ratio larvae of demersal fish | 4 | 2 | y ⁻¹ | 6.365609 | 0.496827 |
| Annual gross PB ratio planktivorous fish | 1.72 | 0.86 | y ⁻¹ | 1.022148 | 0.719477 |
| Annual gross PB ratio demersal fish | 0.88 | 0.44 | y ⁻¹ | 0.299076 | 0.418294 |
| Annual gross PB ratio migratory fish | 1.3 | 0.6 | y ⁻¹ | 0.185255 | 0.178011 |
| Annual net PB ratio birds | 0.28 | 0.14 | y ⁻¹ | 0.126261 | 0.547194 |
| Annual net PB ratio pinnipeds | 0.09 | 0.045 | y ⁻¹ | 0.041177 | 0.555126 |
| Annual net PB ratio cetaceans | 0.02 | 0.01 | y ⁻¹ | 0.029723 | 0.623302 |
| Annual average proportion of kelp C uptake which is exuded | 0.3 | 0.1 | Dimensionless | 0.296805 | 0.99949 |
| Annual average molar NC ratio of kelp | 0.12 | 0.2 | Dimensionless | 0.123307 | 0.999863 |
| Annual denitrification | 129 | 42 | mMN.m ⁻² .y ⁻¹ | 208.0052 | 0.170465 |
| Proportion of demersal fish catch discarded | 0.37 | 0.075 | Dimensionless | 0.360811 | 0.992522 |
| Annual average ammonia concentration in porewater of sand grain size 0.25mm | 19.24 | 9 | mMN.m ⁻³ | 4.715929 | 0.271946 |
| Annual average ammonia concentration in porewater of mud grain size 0.12mm | 63.45 | 22 | mMN.m ⁻³ | 51.95817 | 0.872469 |
| Annual average nitrate concentration in porewater of sand grain size 0.25mm | 4.15 | 2 | mMN.m ⁻³ | 5.160632 | 0.880142 |
| Annual average nitrate concentration in porewater of mud grain size 0.12mm | 2.34 | 1 | mMN.m ⁻³ | 3.121486 | 0.736858 |
| Annual average organic N content of sand grain size 0.25mm (0.19-0.43mm) | 0.05152 | 0.02441 | %N (gN.(g dry sed) ⁻¹) | 0.049978 | 0.998019 |
| Annual average organic N content of mud grain size 0.12mm (0.03-0.07mm) | 0.07357 | 0.0343 | %N (gN.(g dry sed) ⁻¹) | 0.066265 | 0.977575 |
| Average winter (Nov-Feb) nitrate concentration shallow layer | 9.998 | 2.135 | mMN.m ⁻³ | 9.420257 | 0.964048 |
| Average summer (May-Aug) nitrate concentration | 2.161 | 1.089 | mMN.m ⁻³ | 3.215862 | 0.625538 |

| | | | | | |
|--|----------|----------|--------------------------------------|----------|----------|
| shallow layer | | | | | |
| Average winter (Nov-Feb) nitrate concentration deep layer | 6.995 | 0.836 | mMN.m ⁻³ | 9.516758 | 0.010572 |
| Average summer (May-Aug) nitrate concentration deep layer | 2.837 | 0.917 | mMN.m ⁻³ | 5.697196 | 0.007717 |
| Average winter (Nov-Feb) ammonia concentration shallow layer | 2.367 | 0.774 | mMN.m ⁻³ | 1.781435 | 0.751128 |
| Average summer (May-Aug) ammonia concentration shallow layer | 1.737 | 0.669 | mMN.m ⁻³ | 3.181125 | 0.097311 |
| Average winter (Nov-Feb) ammonia concentration deep layer | 0.853 | 0.32 | mMN.m ⁻³ | 1.571777 | 0.080246 |
| Average summer (May-Aug) ammonia concentration deep layer | 1.338 | 0.708 | mMN.m ⁻³ | 2.84671 | 0.103264 |
| Inshore:offshore ratio of annual mean carnivorous zooplankton depth averaged concentration | 0.904055 | 0.385 | Dimensionless | 1.434845 | 0.386596 |
| Inshore:offshore ratio of annual mean omnivorous zooplankton depth averaged concentration | 1.676 | 0.2715 | Dimensionless | 1.404356 | 0.606908 |
| Inshore:offshore ratio of annual mean phytoplankton surface layer concentration | 3.744 | 1 | Dimensionless | 0.992409 | 0.022669 |
| Inshore:offshore ratio of annual mean nitrate surface layer concentration | 3.000 | 1 | Dimensionless | 0.876221 | 0.10482 |
| Inshore:offshore ratio of annual mean ammonia surface layer concentration | 2.4294 | 0.9 | Dimensionless | 1.589962 | 0.647314 |
| Inshore:offshore ratio of annual mean planktivorous fish density (m ⁻²) | 0.67 | 0.49 | Dimensionless | 0.419324 | 0.877341 |
| Inshore:offshore ratio of annual mean demersal fish density (m ⁻²) | 0.39 | 0.31 | Dimensionless | 0.875153 | 0.293866 |
| Annual bycatch of birds | NA | NA | mMN.m ⁻² .y ⁻¹ | 8.99E-07 | NA |
| Annual bycatch of pinnipeds | NA | NA | mMN.m ⁻² .y ⁻¹ | 5.19E-05 | NA |
| Annual bycatch of cetaceans | NA | NA | mMN.m ⁻² .y ⁻¹ | 0.000394 | NA |
| Proportion of kelp annual nitrogen uptake exported as beach-cast | 0.15 | 0.05 | Dimensionless | 0.085593 | 0.436199 |
| Annual cetacean catch | 4.05E-05 | 2.00E-05 | mMN.m ⁻² .y ⁻¹ | 4.04E-05 | 0.999992 |
| Overall model | | | | | 0.421073 |

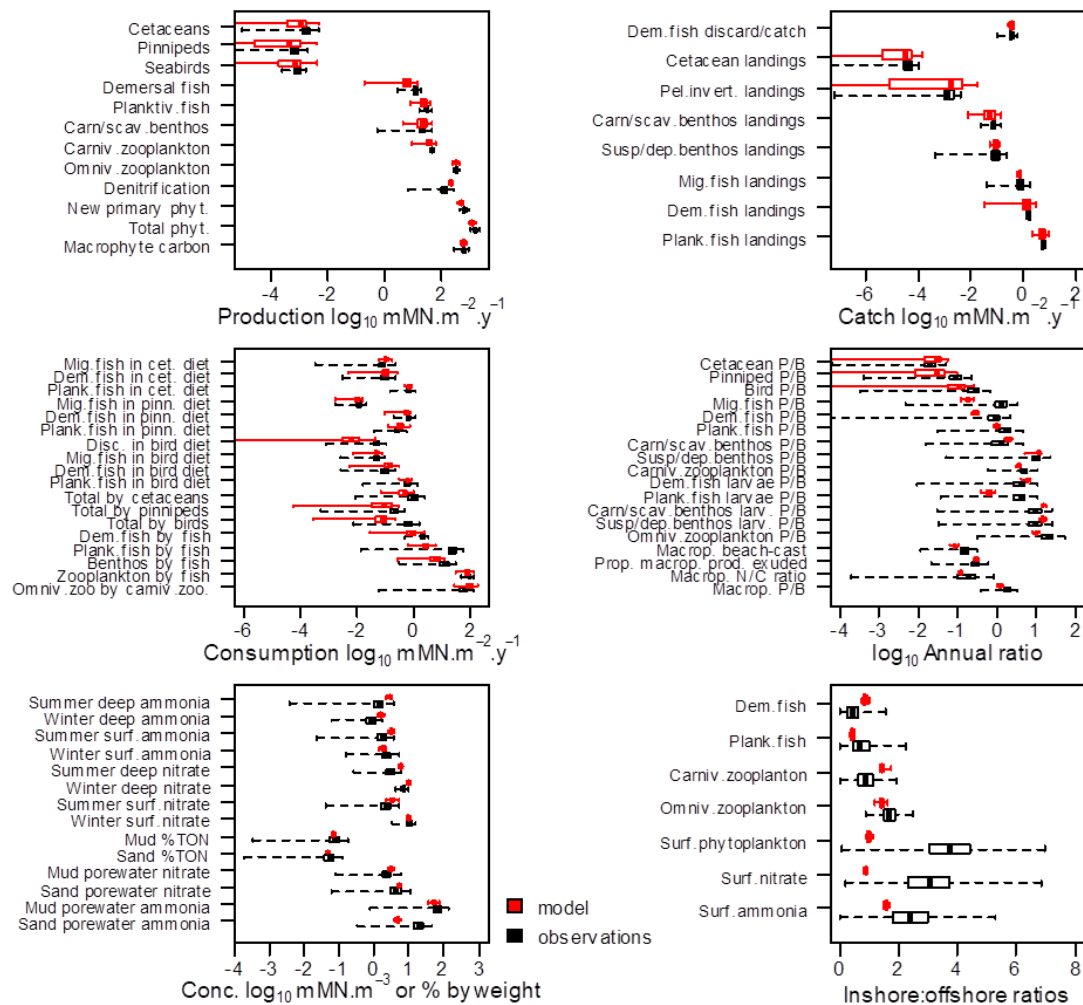


FIGURE 11 Annual integrated or averaged results from the best-fit 1970-1999 stationary model and the corresponding observed data from the North Sea. Red boxes and whiskers show the 0.5, 25, 50, 75 and 99.5 centiles of the likelihood distribution of model results given the uncertainty in fitted parameter values. Black boxes and whiskers show the equivalent variability in measurements from the North Sea aggregated over the period 1970-1999. Drawn with the function `e2e_compare_obs(,selection="ANNUAL")`.

TABLE 33 Observational indices of the 2003-2013 state of the North Sea ecosystem to which the model was fitted using 2003-2013 environmental drivers and fishing fleet inputs, Indices include fluxes between guilds, and annual ratios, and their standard deviations. For each index (*i*) the partial likelihood between observed and modelled values with parameter set θ , obtained from the simulated annealing scheme, was calculated as $\exp(-\chi_{\theta i}^2)$, where $\chi_{\theta i}^2 = \frac{(observed_i - model_{\theta,i})^2}{2\sigma_i^2}$ (σ_i is the standard deviation of observed index *i*), and the overall likelihood is given by $\exp(-(\text{mean}(\chi_{\theta}^2))$. The results are shown graphically in Figure 12.

| Description | Observational value | s.d. of observational value | Units | Maximum likelihood model value | Partial likelihood |
|---|---------------------|-----------------------------|--------------------------------------|--------------------------------|--------------------|
| Annual total primary production | 1522 | 150.94 | mMN.m ⁻² .y ⁻¹ | 1234.26 | 0.162507 |
| Annual new production from depth integrated nitrate | 672.8 | 73.00 | mMN.m ⁻² .y ⁻¹ | 482.3586 | 0.033325 |
| Annual within forest net production of kelp | 600 | 100 | gC.m ⁻² .y ⁻¹ | 600.0001 | 1 |
| Annual omnivorous zooplankton gross production | 339.6 | 25.157 | mMN.m ⁻² .y ⁻¹ | 319.3574 | 0.723445 |
| Annual carnivorous zooplankton gross production | NA | NA | mMN.m ⁻² .y ⁻¹ | 45.57989 | NA |
| Annual planktivorous fish gross production | NA | NA | mMN.m ⁻² .y ⁻¹ | 13.67917 | NA |
| Annual demersal fish gross production | NA | NA | mMN.m ⁻² .y ⁻¹ | 19.80397 | NA |
| Annual suspension /deposit feeding benthos gross production | NA | NA | mMN.m ⁻² .y ⁻¹ | 603.6545 | NA |
| Annual carnivore/scavenge feeding benthos gross production | NA | NA | mMN.m ⁻² .y ⁻¹ | 29.87239 | NA |
| Annual net production of birds | NA | NA | mMN.m ⁻² .y ⁻¹ | 0.00137 | NA |
| Annual net production of pinnipeds | NA | NA | mMN.m ⁻² .y ⁻¹ | 0.009809 | NA |
| Annual net production of cetaceans | NA | NA | mMN.m ⁻² .y ⁻¹ | 0.00133 | NA |
| Annual monthly max concentration of benthos suspension /deposit feeder larvae | NA | NA | mMN.m ⁻³ | 0.608316 | NA |
| Annual monthly max concentration of benthos carnivore/scavenge feeder larvae | NA | NA | mMN.m ⁻³ | 0.443911 | NA |
| Annual consumption of planktivorous fish by fish | NA | NA | mMN.m ⁻² .y ⁻¹ | 3.428481 | NA |
| Annual consumption of demersal fish by fish | NA | NA | mMN.m ⁻² .y ⁻¹ | 2.384715 | NA |
| Annual consumption of omnivorous zooplankton by fish and fish larvae | NA | NA | mMN.m ⁻² .y ⁻¹ | 71.53949 | NA |

| | | | | | |
|---|----------|-------|--------------------------------------|----------|----------|
| Annual consumption of omnivorous zooplankton by carnivorous zooplankton | NA | NA | mMN.m ⁻² .y ⁻¹ | 123.7521 | NA |
| Annual consumption of benthos by fish | NA | NA | mMN.m ⁻² .y ⁻¹ | 22.24755 | NA |
| Annual food consumption by birds | NA | NA | mMN.m ⁻² .y ⁻¹ | 0.125026 | NA |
| Proportion planktivorous fish in diet of birds | NA | NA | Dimensionless | 0.349124 | NA |
| Proportion demersal fish in diet of birds | NA | NA | Dimensionless | 0.347781 | NA |
| Proportion migratory fish in diet of birds | NA | NA | Dimensionless | 0.112013 | NA |
| Proportion discards in diet of birds | NA | NA | Dimensionless | 0.004897 | NA |
| Annual food consumption by pinnipeds | NA | NA | mMN.m ⁻² .y ⁻¹ | 0.537479 | NA |
| Proportion pelagic fish in diet of pinnipeds | NA | NA | Dimensionless | 0.101002 | NA |
| Proportion demersal fish in diet of pinnipeds | NA | NA | Dimensionless | 0.86852 | NA |
| Proportion migratory fish in diet of pinnipeds | NA | NA | Dimensionless | 0.014271 | NA |
| Annual food consumption by cetaceans | NA | NA | mMN.m ⁻² .y ⁻¹ | 0.555505 | NA |
| Proportion pelagic fish in diet of cetaceans | NA | NA | Dimensionless | 0.435075 | NA |
| Proportion demersal fish in diet of cetaceans | NA | NA | Dimensionless | 0.29101 | NA |
| Proportion migratory fish in diet of cetaceans | NA | NA | Dimensionless | 0.224505 | NA |
| Annual planktivorous fish landings (live weight) | 2.928 | 0.6 | mMN.m ⁻² .y ⁻¹ | 2.266516 | 0.544589 |
| Annual demersal fish landings (live weight) | 0.6408 | 0.3 | mMN.m ⁻² .y ⁻¹ | 0.829583 | 0.820374 |
| Annual migratory fish landings (live weight) | 0.9925 | 0.25 | mMN.m ⁻² .y ⁻¹ | 1.00489 | 0.998773 |
| Annual suspension /deposit feeding benthos landings (live weight) | 0.07444 | 0.028 | mMN.m ⁻² .y ⁻¹ | 0.075691 | 0.999003 |
| Annual carnivore/scavenge feeding benthos landings (live weight) | 0.12801 | 0.05 | mMN.m ⁻² .y ⁻¹ | 0.129282 | 0.999677 |
| Annual carnivorous zooplankton landings (live weight) | 0.006976 | 0.004 | mMN.m ⁻² .y ⁻¹ | 0.00657 | 0.994866 |
| Annual kelp landings (live weight) | 0 | 0 | mMN.m ⁻² .y ⁻¹ | 0 | NA |
| Annual carbon gross PB ratio of kelp | 2 | 0.5 | y ⁻¹ | 1.332273 | 0.40995 |
| Annual gross PB ratio larvae of suspension /deposit feeding benthos | 10 | 5 | y ⁻¹ | 15.56301 | 0.538515 |
| Annual gross PB ratio larvae of carnivore/scavenge feeding benthos | 10 | 5 | y ⁻¹ | 15.70319 | 0.52177 |
| Annual gross PB ratio suspension /deposit feeding benthos | 10 | 3 | y ⁻¹ | 12.20704 | 0.762913 |
| Annual gross PB ratio carnivore/scavenge feeding | 1.2 | 1 | y ⁻¹ | 1.971445 | 0.742625 |

| | | | | | |
|---|----------|---------|--------------------------------------|----------|----------|
| benthos | | | | | |
| Annual gross PB ratio omnivorous zooplankton | 20 | 10 | y ⁻¹ | 11.53846 | 0.699081 |
| Annual gross PB ratio carnivorous zooplankton | 5 | 1.315 | y ⁻¹ | 3.597897 | 0.566413 |
| Annual gross PB ratio larvae of planktivorous fish | 4 | 2 | y ⁻¹ | 0.633989 | 0.242621 |
| Annual gross PB ratio larvae of demersal fish | 4 | 2 | y ⁻¹ | 5.564162 | 0.736515 |
| Annual gross PB ratio planktivorous fish | 1.72 | 0.86 | y ⁻¹ | 1.029416 | 0.724402 |
| Annual gross PB ratio demersal fish | 0.88 | 0.44 | y ⁻¹ | 0.301554 | 0.421408 |
| Annual gross PB ratio migratory fish | 1.3 | 0.6 | y ⁻¹ | 0.179503 | 0.17486 |
| Annual net PB ratio birds | 0.28 | 0.14 | y ⁻¹ | 0.157159 | 0.680486 |
| Annual net PB ratio pinnipeds | 0.09 | 0.045 | y ⁻¹ | 0.140334 | 0.534959 |
| Annual net PB ratio cetaceans | 0.02 | 0.01 | y ⁻¹ | 0.029476 | 0.638279 |
| Annual average proportion of kelp C uptake which is exuded | 0.3 | 0.1 | Dimensionless | 0.297185 | 0.999604 |
| Annual average molar NC ratio of kelp | 0.12 | 0.2 | Dimensionless | 0.12224 | 0.999937 |
| Annual denitrification | 129 | 42 | mMN.m ⁻² .y ⁻¹ | 205.4724 | 0.190595 |
| Proportion of demersal fish catch discarded | 0.51 | 0.1 | Dimensionless | 0.514979 | 0.998761 |
| Annual average ammonia concentration in porewater of sand grain size 0.25mm | 19.24 | 9 | mMN.m ⁻³ | 4.794285 | 0.275783 |
| Annual average ammonia concentration in porewater of mud grain size 0.12mm | 63.45 | 22 | mMN.m ⁻³ | 53.57287 | 0.904129 |
| Annual average nitrate concentration in porewater of sand grain size 0.25mm | 4.15 | 2 | mMN.m ⁻³ | 5.054439 | 0.902803 |
| Annual average nitrate concentration in porewater of mud grain size 0.12mm | 2.34 | 1 | mMN.m ⁻³ | 3.045961 | 0.779431 |
| Annual average organic N content of sand grain size 0.25mm (0.19-0.43mm) | 0.051515 | 0.02441 | %N (gN.(g dry sed) ⁻¹) | 0.049978 | 0.99802 |
| Annual average organic N content of mud grain size 0.12mm (0.03-0.07mm) | 0.07357 | 0.0343 | %N (gN.(g dry sed) ⁻¹) | 0.066284 | 0.977694 |
| Average winter (Nov-Feb) nitrate concentration shallow layer | NA | NA | mMN.m ⁻³ | 9.339934 | NA |
| Average summer (May-Aug) nitrate concentration shallow layer | NA | NA | mMN.m ⁻³ | 2.930569 | NA |
| Average winter (Nov-Feb) nitrate concentration deep layer | NA | NA | mMN.m ⁻³ | 9.431013 | NA |
| Average summer (May-Aug) nitrate concentration | NA | NA | mMN.m ⁻³ | 5.615285 | NA |

| | | | | | |
|--|----------|----------|--------------------------------------|----------|----------|
| deep layer | | | | | |
| Average winter (Nov-Feb) ammonia concentration shallow layer | NA | NA | mMN.m ⁻³ | 1.813984 | NA |
| Average summer (May-Aug) ammonia concentration shallow layer | NA | NA | mMN.m ⁻³ | 3.234808 | NA |
| Average winter (Nov-Feb) ammonia concentration deep layer | NA | NA | mMN.m ⁻³ | 1.596596 | NA |
| Average summer (May-Aug) ammonia concentration deep layer | NA | NA | mMN.m ⁻³ | 2.904538 | NA |
| Inshore:offshore ratio of annual mean carnivorous zooplankton depth averaged concentration | 0.904055 | 0.2 | Dimensionless | 1.397925 | 0.047413 |
| Inshore:offshore ratio of annual mean omnivorous zooplankton depth averaged concentration | 1.675727 | 0.271539 | Dimensionless | 1.278671 | 0.343326 |
| Inshore:offshore ratio of annual mean phytoplankton surface layer concentration | 3.744403 | 1 | Dimensionless | 1.002478 | 0.023305 |
| Inshore:offshore ratio of annual mean nitrate surface layer concentration | 3.000139 | 1 | Dimensionless | 0.879365 | 0.105521 |
| Inshore:offshore ratio of annual mean ammonia surface layer concentration | 2.429353 | 0.9 | Dimensionless | 1.606846 | 0.658623 |
| Inshore:offshore ratio of annual mean planktivorous fish density (m ⁻²) | 0.67 | 0.49 | dimensionless | 0.368966 | 0.828022 |
| Inshore:offshore ratio of annual mean demersal fish density (m ⁻²) | 0.39 | 0.31 | Dimensionless | 0.944781 | 0.201622 |
| Annual bycatch of birds | 6.25E-07 | 3.00E-07 | mMN.m ⁻² .y ⁻¹ | 4.59E-07 | 0.857692 |
| Annual bycatch of pinnipeds | 2.74E-05 | 5.00E-05 | mMN.m ⁻² .y ⁻¹ | 2.71E-05 | 0.99998 |
| Annual bycatch of cetaceans | 0.000275 | 1.00E-04 | mMN.m ⁻² .y ⁻¹ | 0.000276 | 0.999996 |
| Proportion of kelp annual nitrogen uptake exported as beach-cast | 0.15 | 0.05 | Dimensionless | 0.083798 | 0.41622 |
| Cetacean catch | 8.38E-05 | 4.00E-05 | mMN.m ⁻² .y ⁻¹ | 8.84E-05 | 0.993414 |
| Overall model | | | | | 0.496398 |

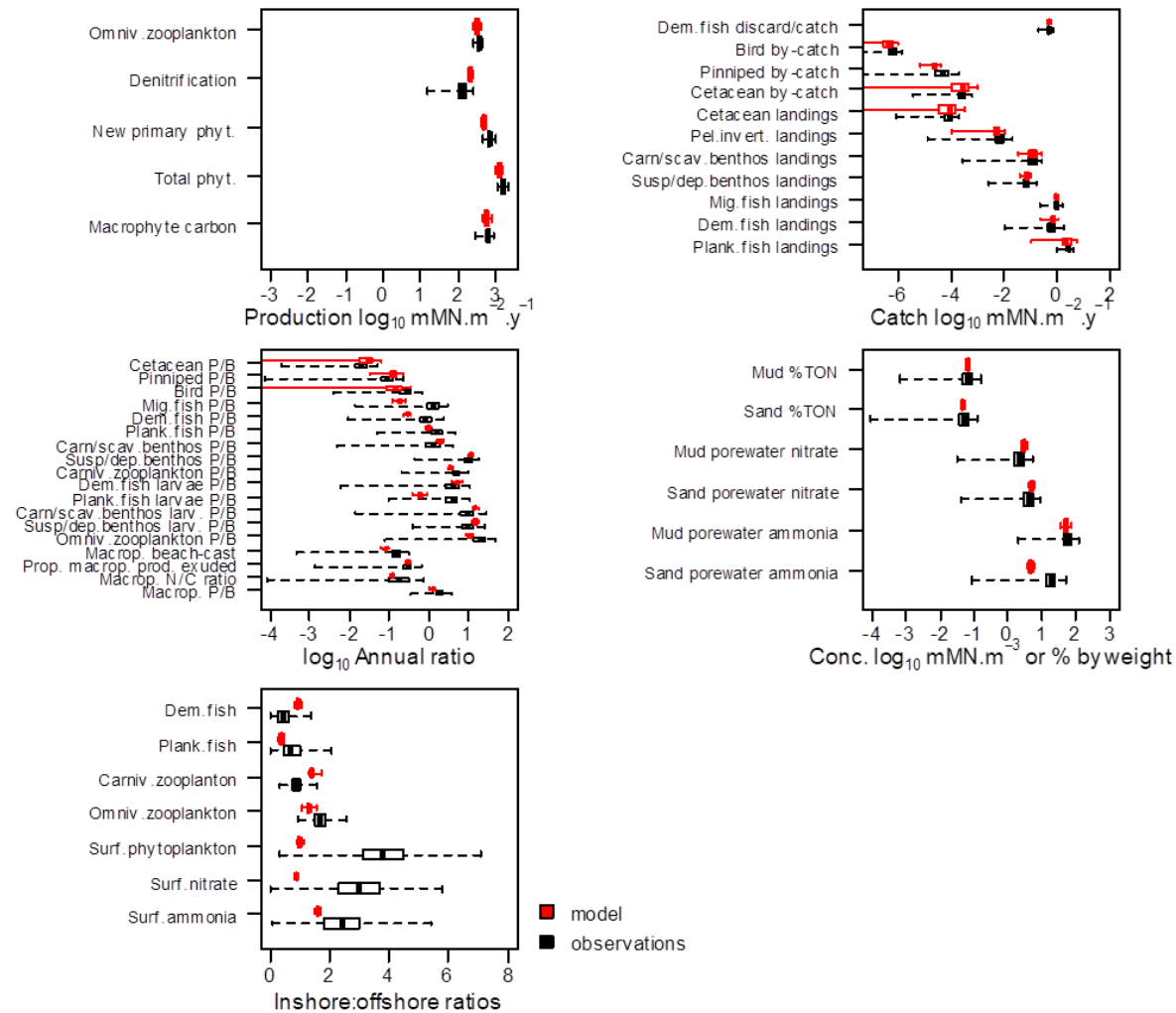


FIGURE 12 Annual integrated or averaged results from the best-fit 2003-2013 stationary model and the corresponding observed data from the North Sea. Red boxes and whiskers show the 0.5, 25, 50, 75 and 99.5 centiles of the likelihood distribution of model results given the uncertainty in fitted parameter values. Black boxes and whiskers show the equivalent variability in measurements from the North Sea aggregated over the period 2003-2013. Drawn with the function `e2e_compare_obs(,selection="ANNUAL")`.

TABLE 34 Maximum likelihood preference parameters $pref_{\text{resource-consumer}}$ for all resource-consumer links in the model, estimated using the simulated annealing scheme to fit the 1970-1999 model to the observed data on ecosystem state. Preferences for each consumer guild (columns) sum to 1.0. Values shown are constrained to 3 decimal places. Note that the preference parameters are inputs to the model, and do not represent proportions in the diet, Diet composition is an emergent output of the model.

| Resources | ID | Consumers | | | | | | | | | | | | | | | |
|--|----|-----------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| | | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 |
| Water column ammonia | 1 | 0.271 | 0.271 | | | | | | | | | | | | | | |
| Water column nitrate | 2 | 0.729 | 0.729 | | | | | | | | | | | | | | |
| Suspended detritus | 3 | | | 0.001 | | | | | | | 0.976 | 0.312 | 0.295 | | | | |
| Sediment detritus | 4 | | | | | | | | | | | | 0.478 | | | | |
| Macrophyte debris | 5 | | | | | | | | | | | | | 0.007 | | | |
| Corpses | 6 | | | | | | | | | 0.010 | | | | 0.420 | 0.088 | 0.053 | |
| Fishery discards | 7 | | | | | | | | | 0.097 | | | | | 0.259 | 0.059 | 0.132 |
| Macrophytes | 8 | | | | | | | | | | | | | 0.013 | | | |
| Phytoplankton | 9 | | | 0.906 | | | | | | | 0.024 | 0.688 | 0.227 | | | | |
| Omnivorous zooplankton | 10 | | | | 0.653 | 0.147 | 0.947 | 0.361 | 0.366 | | | | | | | | 0.010 |
| Carnivorous zooplankton | 11 | | | | | | | 0.152 | 0.014 | 0.075 | | | | | 0.091 | 0.000 | 0.018 |
| Larvae of planktivorous fish | 12 | | | | 0.017 | | | 0.128 | 0.050 | 0.116 | | | | | | | |
| Larvae of demersal fish | 13 | | | | 0.041 | | | 0.179 | 0.058 | 0.111 | | | | | | | |
| Planktivorous fish | 14 | | | | | | | | | 0.106 | | | | | 0.233 | 0.135 | 0.310 |
| Migratory fish | 15 | | | | | | | | | 0.018 | | | | | 0.184 | 0.048 | 0.390 |
| Demersal fish | 16 | | | | | | | | | 0.006 | | | | | 0.145 | 0.705 | 0.136 |
| Larvae of suspension/deposit feeding benthos | 17 | | | 0.002 | 0.192 | 0.588 | 0.043 | 0.088 | 0.285 | | | | | | | | |
| Larvae of carnivorous/scavenge feeding benthos | 18 | | | 0.092 | 0.096 | 0.265 | 0.010 | 0.092 | 0.227 | | | | | | | | |
| Suspension/deposit feeding benthos | 19 | | | | | | | | | 0.390 | | | | 0.561 | 0.000 | 0.000 | 0.000 |
| Carnivorous/scavenge feeding benthos | 20 | | | | | | | | | 0.072 | | | | | 0.000 | 0.000 | 0.000 |
| Birds | 21 | | | | | | | | | | | | | | | 0.000 | 0.000 |
| Pinnipeds | 22 | | | | | | | | | | | | | | | | 0.005 |
| Cetaceans | 23 | | | | | | | | | | | | | | | | |

TABLE 35 Fitted uptake, mortality, migration and exploitable fraction parameters for the maximum likelihood model, estimated using the simulated annealing scheme. Maximum uptake rates are given at the Q_{10} reference temperature.

| Consumer guild | Maximum carbon uptake rate | Carbon exudation rate | Maximum nitrogen uptake rate | Nitrogen uptake half-saturation coefficient | Beddington-DeAngelis parameter | Density dependent mortality coefficient | Active migration coefficient | Maximum exploitable fraction of the stock |
|--|----------------------------|-----------------------|------------------------------|---|--------------------------------|---|------------------------------|---|
| Macrophytes | 0.026 | 1.811E-08 | 0.009 | 18.272 | | 3.099E-08 | | 0.041 |
| Phytoplankton – shallow | | | 2.876 | 4.302 | | 5.142E-02 | | |
| Phytoplankton – deep | | | | | | 6.328E-02 | | |
| Omnivorous zooplankton | | | 2.156 | 3.390 | | 3.346E-05 | | |
| Carnivorous zooplankton | | | 0.154 | 1.086 | | 2.090E-04 | | 0.004 |
| Larvae of planktivorous fish | | | 0.315 | 6.408 | | 4.101E-06 | | |
| Larvae of demersal fish | | | 0.233 | 2.036 | | 5.787E-07 | | |
| Planktivorous fish | | | 0.134 | 2.127 | | 1.906E-05 | 1.579E-03 | 0.874 |
| Migratory fish | | | 0.033 | 2.264 | | 1.691E-06 | 3.545E-05 | 0.684 |
| Demersal fish | | | 0.011 | 0.526 | | 3.760E-05 | 2.205E-03 | 0.900 |
| Larvae of suspension/deposit feeding benthos | | | 0.334 | 1.033 | | 9.633E-06 | | |
| Larvae of carnivorous/scavenge feeding benthos | | | 0.993 | 2.371 | | 2.540E-07 | | |
| Suspension/deposit feeding benthos | | | 0.910 | 1.722 | | 3.674E-04 | | 0.030 |
| Carnivorous/scavenge feeding benthos | | | 0.027 | 9.016 | | 3.979E-04 | | 0.155 |
| Birds | | | 0.648 | 0.995 | 4553.179 | 4.367E-02 | 9.759E-03 | 0.100 |
| Pinnipeds | | | 0.430 | 1.669 | 2113.567 | 5.091E-03 | 9.837E-03 | 0.099 |
| Cetaceans | | | 0.398 | 0.519 | 895.342 | 1.050E-03 | 9.985E-03 | 0.313 |

TABLE 36 Fitted values for parameters of microbiological rates at the Q_{10} reference temperature, and other related parameters, for the maximum likelihood model, estimated using the simulated annealing scheme.

| Description | Value |
|--|--------------|
| Water column detritus mineralisation rate | 1.676E-05 |
| Upper layer water column nitrification rate | 1.187E-07 |
| Upper layer water column denitrification rate | 2.380E-09 |
| Lower layer water column nitrification rate | 3.052E-02 |
| Lower layer water column denitrification rate | 4.873E-07 |
| Formation rate parameter for refractory detritus | 4.962E-01 |
| Mineralisation rate scaling parameter for refractory detritus | 1.627E-04 |
| Proportion of refractory detritus digestible by benthos | 2.614E-02 |
| Sediment detritus mineralisation rate | 1.485E-02 |
| Grain size sensitivity for sediment detritus mineralisation rate | -1.598E-08 |
| Sediment nitrification rate | 6.276E-05 |
| Grain size sensitivity for sediment nitrification rate | -2.645E-06 |
| Sediment denitrification rate | 5.281E-01 |
| Grain size sensitivity for sediment denitrification rate | 1.537E-07 |
| Conversion rate of discards to corpses | 1.837E-02 |
| Conversion rate of corpses to sediment detritus | 2.233E-01 |
| Detritus sinking rate in the upper layers | 2.669E-02 |
| Detritus sinking rate in the lower layer | 9.813E-02 |
| Density dependent self-shading parameter for macrophytes | 3.580E-06 |
| Wave-dependent beach-cast rate for macrophyte debris | 4.422E-05 |
| Fitting parameter for undersize demersal fish function | 1.481E-03 |

TABLE 37 Fixed assimilation, metabolism, fecundity, and exploitation threshold parameters which were not subject to fitting. The only parameter which differed between the 1970-1999 and 2003-2013 simulation periods was the threshold for zero remaining exploitable biomass of carnivorous zooplankton. The value was set to 14 mMN.m⁻² as shown here for 1970-1999, and 4 mMN.m⁻² for 2003-2013 to reflect an observed trend in squid abundance and landings in the North Sea (Pierce *et al.*, 1998; van der Kooij *et al.*, 2016)

| Consumer guild | Assimilation efficiency | Background metabolic rate at Q ₁₀ reference temperature (d ⁻¹) | Annual weight-specific fecundity | Threshold for zero exploitable biomass remaining (mMN.m ⁻²) | Minimum inedible biomass (mMN.m ⁻²) |
|--|-------------------------|---|----------------------------------|---|---|
| Macrophytes | | | | 1 | |
| Phytoplankton | | | | | |
| Omnivorous zooplankton | 0.34 | 1.000E-02 | | | |
| Carnivorous zooplankton | 0.34 | 5.000E-03 | | 14 | 1 |
| Larvae of planktivorous fish | 0.34 | 5.000E-05 | | | |
| Larvae of demersal fish | 0.34 | 5.000E-05 | | | |
| Planktivorous fish | 0.275 | 1.400E-03 | 0.25 | 0.1 | |
| Migratory fish | 0.25 | 4.000E-04 | | 0.1 | |
| Demersal fish | 0.25 | 8.200E-04 | 0.25 | 0.1 | |
| Larvae of suspension/deposit feeding benthos | 0.34 | 1.000E-02 | | | |
| Larvae of carnivorous/scavenge feeding benthos | 0.34 | 1.000E-02 | | | |
| Suspension/deposit feeding benthos | 0.34 | 1.000E-02 | 0.10 | 20 | |
| Carnivorous/scavenge feeding benthos | 0.34 | 1.000E-03 | 0.40 | 1 | |
| Birds | 0.15 | 5.500E-03 | | 0.0006 | |
| Pinnipeds | 0.15 | 2.800E-03 | | 0.002 | |
| Cetaceans | 0.15 | 5.000E-03 | | 0.03 | |

TABLE 38. Other fixed parameters which were not subject to fitting.

| Description | Value |
|--|--------------|
| Irradiance at maximum carbon uptake by macrophytes ($E.m^{-2}.d^{-1}$) | 3 |
| Minimum nitrogen:carbon molar ratio for macrophytes | 0.02 |
| Maximum nitrogen:carbon molar ratio for macrophytes | 0.15 |
| Irradiance at maximum nitrogen uptake by phytoplankton ($E.m^{-2}.d^{-1}$) | 4 |
| Autotroph Q_{10} value | 1.20 |
| Heterotroph uptake Q_{10} value | 1.32 |
| Metabolic and bacterial Q_{10} value | 1.44 |
| Q_{10} reference temperature | 10 |

Sensitivity analysis

The sensitivity analysis (e2e_run_sens()) tested the influence of individual parameters and the fishing and environmental drivers on the overall likelihood of the observational target data given the 1970-1999 model. The analysis was conducted on the basis of the factorial sampling scheme (Morris, 1991) as implemented in the StrathE2E2 package with 16 trajectories of the model (randomised parameter sets). For each trajectory, multiple model runs were executed, with a different single internal parameter being varied in each successive run, drawn from a symmetrical random uniform distribution. With a total of 450 parameters and drivers being included in the analysis therefore required a total of 7216 model runs each of 40 years.

The analysis included the fixed and fitted parameters of the ecology model, fishing fleet model parameters, harvest ratios on each model guild, environmental and biological event drivers, and the physical configuration parameters (layer thicknesses, inshore/offshore areas, sediment properties).

For the ecology model component, the number of parameters in the sensitivity analysis was greater than the number of input parameters for the model. This was because within the model each prey-predator pair is represented by a discrete uptake function defined by a maximum uptake rate and half-saturation coefficient. These are derived by the combination of the preference matrix and, for each predator, the single values of maximum uptake rate and half-saturation coefficient which are the input parameters to the model.

Of the 450 parameters in the analysis, 120 emerged as having significant sensitivity with respect to overall model fit to the observed data (95% probability that the distributions of elementary effects were non-zero). Of these 64% were ecology model parameters (42% fitted, 23% fixed). Physical configuration parameters made up 14% of the significantly sensitive set; harvest ratios only 7%. This does not mean that model was insensitive to harvest ratios; just that compared to many of the other parameters in the model the harvest ratios were not among the most sensitive for the overall fit given the range of variations imposed in the analysis (Table 39, Figure 13).

The parameter class with the highest proportion of significantly sensitive terms was the fixed ecology model group (27 out of 52; 52%). The standard deviations of the significantly sensitive parameters indicated that they all had strong interactions with other parameters. 51% of the physical configuration parameters had significant sensitivity.

The main conclusions from the sensitivity analysis were:

- 1) assimilation efficiencies and maximum uptake rates of mid-trophic level guilds were the most sensitive parameters for the overall model. This implies that the food web as a whole was potentially sensitive to processes that might change the community-level feeding responses of these guilds, such as species invasions and replacements, and physiological effects not represented in the model such as ocean acidification (temperature effects are included in the model already and the Q_{10} for uptake and metabolic rates emerge as having significant sensitivity).

- 2) A high proportion of significantly sensitive ecology model parameters (35%) were independently fixed and not included in the simulated annealing fitting process. This indicates that the fixed parameters were well chosen. Given that the model as a whole was somewhat under-constrained by the observed fitting data, the selection of high-sensitivity parameters for independent constraint was a sound choice.

Parameter sensitivity analysis of the 1970-1999 maximum likelihood model

TABLE 39 One-at-a-time factorial sampling scheme sensitivity analysis results based on 16 trajectories, measuring the sensitivity of the overall likelihood of the 1970-1999 model to each of the parameters and drivers. The mean elemental effect (EE_mean) is a measure of the sensitivity of the model to each individual parameter. The corresponding standard deviation of the elemental effect (EE_sd) is a measure of the susceptibility of each parameter to interactions with other parameters. The results are ranked by decreasing values of the absolute value of EE_mean, so the model is most sensitive to the first parameters in the list. The column labelled ‘Signif’ indicates whether the value of EE_mean was significantly different from zero ($p < 0.05$). Significant parameters are highlighted by grey-shaded cells. Parameters with EE_mean and EE_sd = NA were set to zero in the baseline model or not included in the sensitivity analysis.

| Parameter class | Parameter description | Model guild or feature | EE_mean | EE_sd | Signif. |
|----------------------|---------------------------------------|---|---------|--------|---------|
| Ecology model fixed | Assimilation efficiency | Planktivorous fish | -3.6625 | 0.6965 | sig |
| Ecology model fitted | Maximum uptake rate | Omnivorous zooplankton by carnivorous zooplankton | -3.5883 | 0.7679 | sig |
| Ecology model fixed | Assimilation efficiency | Carnivorous zooplankton | -3.2346 | 0.8094 | sig |
| Ecology model fitted | Maximum uptake rate | Omnivorous zooplankton by planktivorous fish | -2.8473 | 0.8103 | sig |
| Ecology model fixed | Assimilation efficiency | Demersal fish larvae | -2.8221 | 2.4440 | sig |
| Ecology model fitted | Maximum uptake rate | Omnivorous zooplankton by demersal fish larvae | -2.5204 | 2.3415 | sig |
| Ecology model fixed | Assimilation efficiency | Omnivorous zooplankton | -2.4244 | 2.1161 | sig |
| Ecology model fitted | Uptake half saturation coefficient | Phytoplankton by omnivorous zooplankton | -2.2406 | 1.0706 | sig |
| Ecology model fitted | Maximum uptake rate | Phytoplankton by omnivorous zooplankton | -2.1823 | 2.2996 | sig |
| Ecology model fitted | Uptake half saturation coefficient | Omnivorous zooplankton by carnivorous zooplankton | -2.1378 | 2.0290 | sig |
| Ecology model fitted | Uptake half saturation coefficient | Omnivorous zooplankton by demersal fish larvae | -2.0745 | 1.0685 | sig |
| Ecology model fitted | Uptake half saturation coefficient | Omnivorous zooplankton by planktivorous fish | -1.6102 | 1.2772 | sig |
| Ecology model fixed | Background metabolic rate coefficient | Planktivorous fish | -1.2816 | 0.9279 | sig |
| Ecology model fixed | Background metabolic rate coefficient | Carnivorous zooplankton | -1.1676 | 1.6671 | sig |
| Ecology model fitted | Maximum uptake rate | Nitrate by phytoplankton | -0.7665 | 0.7604 | sig |

| | | | | | |
|-------------------------|---|---|---------|--------|-----|
| Physical configuration | Vertical thickness | Offshore zone upper layer | -0.6712 | 0.7987 | sig |
| Ecology model fitted | Density dependent mortality coefficient | Phytoplankton upper layer | -0.6194 | 0.7932 | sig |
| Environmental driver | Boundary concentration | Lower layer nitrate | -0.5881 | 0.8173 | sig |
| Ecology model fitted | Maximum uptake rate | Carnivorous zooplankton by planktivorous fish | -0.5332 | 0.8725 | sig |
| Ecology model fitted | Uptake half saturation coefficient | Carnivorous zooplankton by planktivorous fish | -0.5243 | 0.6670 | sig |
| Ecology model fixed | Background metabolic rate coefficient | Demersal fish | -0.5155 | 0.7296 | sig |
| Physical configuration | Vertical thickness | Offshore zone lower layer | -0.4606 | 0.6248 | sig |
| Ecology model fixed | Maximum exploitable fraction of stock | Demersal fish | -0.4599 | 0.6137 | sig |
| Ecology model fitted | Maximum uptake rate | Demersal fish larvae by planktivorous fish | -0.4035 | 0.6252 | sig |
| Biological event driver | Spawning rate | Demersal fish | -0.3835 | 1.1114 | ns |
| Ecology model fixed | Annual fecundity | Demersal fish | -0.3835 | 1.1114 | ns |
| Ecology model fixed | Q10 | Heterotrophic uptake | -0.3792 | 0.6648 | sig |
| Ecology model fixed | Q10 reference temperature | All temperature dependent processes | -0.3473 | 0.5915 | sig |
| Ecology model fixed | Saturation light intensity for uptake | Nutrient by phytoplankton | -0.3391 | 0.3910 | sig |
| Ecology model fixed | Background metabolic rate coefficient | Omnivorous zooplankton | -0.3349 | 0.5699 | sig |
| Ecology model fitted | Uptake half saturation coefficient | Nitrate by phytoplankton | -0.3075 | 0.3207 | sig |
| Ecology model fixed | Q10 | Autotrophic uptake | -0.3069 | 0.5698 | sig |
| Ecology model fixed | Assimilation efficiency | Demersal fish | -0.2921 | 0.9224 | ns |
| Harvest ratio | Harvest ratio offshore | Demersal fish | -0.2481 | 0.4380 | sig |
| Biological event driver | Recruitment rate | Demersal fish | -0.2289 | 0.5044 | ns |
| Ecology model fixed | Maximum exploitable fraction of stock | Planktivorous fish | -0.2246 | 0.3973 | sig |
| Ecology model fitted | Density dependent mortality coefficient | Carnivorous zooplankton | -0.2039 | 0.8739 | ns |
| Environmental driver | Sea surface irradiance | Inshore and offshore zones | -0.1995 | 0.4546 | ns |
| Harvest ratio | Harvest ratio offshore | Planktivorous fish | -0.1553 | 0.3158 | ns |
| Ecology model fitted | Density dependent mortality coefficient | Demersal fish | -0.1539 | 0.3946 | ns |
| Physical configuration | Coefficient | Light attenuation coefficient vs SPM | -0.1526 | 0.2887 | sig |
| Ecology model fitted | Density dependent mortality coefficient | Phytoplankton lower layer | -0.1441 | 0.3600 | ns |
| Ecology model fixed | Assimilation efficiency | Carnivore/scavenge feeding benthos larvae | -0.1370 | 0.2756 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Demersal fish larvae by planktivorous fish | -0.1333 | 0.7183 | ns |
| Ecology model fitted | Maximum uptake rate | Suspension/deposit feeding benthos by demersal fish | -0.1186 | 0.6585 | ns |

| | | | | | |
|------------------------|---|--|---------|--------|-----|
| Ecology model fixed | Assimilation efficiency | Suspension/deposit feeding benthos larvae | -0.1154 | 0.2465 | ns |
| Ecology model fixed | Assimilation efficiency | Birds | -0.1149 | 0.0904 | sig |
| Ecology model fitted | Maximum uptake rate | Suspended detritus by suspension/deposit feeding benthos larvae | -0.1008 | 0.1877 | sig |
| Ecology model fitted | Density dependent mortality coefficient | Planktivorous fish | -0.0961 | 0.3005 | ns |
| Environmental driver | Suspended particulate matter | Offshore zone | -0.0908 | 0.2183 | ns |
| Physical configuration | Vertical thickness | Inshore zone | -0.0802 | 0.3032 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Suspension/deposit feeding benthos by demersal fish | -0.0783 | 0.2371 | ns |
| Harvest ratio | Harvest ratio inshore | Demersal fish | -0.0768 | 0.2227 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Ammonia by phytoplankton | -0.0757 | 0.1484 | sig |
| Physical configuration | Intercept | Light attenuation coefficient vs SPM | -0.0665 | 0.1355 | ns |
| Fishing fleet model | Discard rate offshore | Cetaceans | -0.0662 | 0.0446 | sig |
| Environmental driver | Temperature | Offshore zone upper layer | -0.0613 | 0.5035 | ns |
| Ecology model fixed | Background metabolic rate coefficient | Birds | -0.0545 | 0.0714 | sig |
| Ecology model fitted | Uptake half saturation coefficient | Planktivorous fish by demersal fish | -0.0518 | 0.1447 | ns |
| Ecology model fitted | Maximum uptake rate | Suspended detritus by carnivore/scavenge feeding benthos larvae | -0.0483 | 0.1526 | ns |
| Ecology model fitted | Maximum uptake rate | Planktivorous fish by birds | -0.0444 | 0.0408 | sig |
| Ecology model fitted | Density dependent mortality coefficient | Suspension/deposit feeding benthos | -0.0426 | 0.1783 | ns |
| Ecology model fitted | Maximum uptake rate | Ammonia by phytoplankton | -0.0424 | 0.2467 | ns |
| Ecology model fitted | Maximum uptake rate | Suspension/deposit feeding benthos larvae by carnivorous zooplankton | -0.0420 | 0.1940 | ns |
| Ecology model fitted | Maximum uptake rate | Phytoplankton by suspension/deposit feeding benthos | -0.0414 | 0.1048 | ns |
| Physical configuration | Vertical thickness | Benthic boundary feeding layer | -0.0402 | 0.1262 | ns |
| Ecology model fitted | Maximum uptake rate | Phytoplankton by carnivore/scavenge feeding benthos larvae | -0.0402 | 0.1987 | ns |
| Ecology model fixed | Background metabolic rate coefficient | Pinnipeds | -0.0381 | 0.0466 | sig |
| Ecology model fitted | Maximum uptake rate | Suspension/deposit feeding benthos larvae by planktivorous fish | -0.0374 | 0.0934 | ns |
| Ecology model fitted | Maximum uptake rate | Demersal fish larvae by carnivorous zooplankton | -0.0362 | 0.1413 | ns |
| Ecology model fixed | Assimilation efficiency | Cetaceans | -0.0358 | 0.0188 | sig |
| Ecology model fitted | Uptake half saturation coefficient | Carnivore/scavenge feeding benthos larvae by omnivorous zooplankton | -0.0349 | 0.2127 | ns |

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|------------------------|--|---|---------|--------|-----|
| Ecology model fitted | Maximum uptake rate | Suspended detritus by suspension/deposit feeding benthos | -0.0344 | 0.3513 | ns |
| Ecology model fixed | Assimilation efficiency | Suspension/deposit feeding benthos | -0.0337 | 0.3413 | ns |
| Ecology model fixed | Assimilation efficiency | Pinnipeds | -0.0330 | 0.0496 | sig |
| Fishing fleet model | Coefficient | Demersal fish quota-limited undersize vs nitrogen mass | -0.0317 | 0.1777 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Suspended detritus by suspension/deposit feeding benthos | -0.0313 | 0.1528 | ns |
| Ecology model fixed | Background metabolic rate coefficient | Suspension/deposit feeding benthos | -0.0298 | 0.1011 | ns |
| Ecology model fitted | Denitrification rate coefficient | Sediment porewater nitrate | -0.0290 | 0.0576 | sig |
| Ecology model fitted | Uptake half saturation coefficient | Planktivorous fish larvae by planktivorous fish | -0.0285 | 0.0747 | ns |
| Physical configuration | Proportion of depth range occupied | Phytoplankton inshore | -0.0265 | 0.0886 | ns |
| Environmental driver | Suspended particulate matter | Inshore zone | -0.0261 | 0.0824 | ns |
| Environmental driver | Boundary volume inflow rate | Lower layer offshore | -0.0248 | 0.1538 | ns |
| Ecology model fixed | Background metabolic rate coefficient | Cetaceans | -0.0230 | 0.0212 | sig |
| Ecology model fixed | Assimilation efficiency | Planktivorous fish larvae | -0.0228 | 0.0750 | ns |
| Fishing fleet model | Discard rate offshore | Demersal fish | -0.0217 | 0.1422 | ns |
| Ecology model fitted | Conversion rate coefficient | Labile to refractory sediment detritus | -0.0216 | 0.1194 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Suspended detritus by carnivore/scavenge feeding benthos larvae | -0.0196 | 0.0905 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Carnivorous zooplankton by demersal fish | -0.0193 | 0.0829 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Demersal fish larvae by carnivorous zooplankton | 0.0183 | 0.1414 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Suspension/deposit feeding benthos larvae by demersal fish larvae | -0.0182 | 0.0774 | ns |
| Ecology model fitted | Density dependent mortality coefficient | Carnivore/scavenge feeding benthos | -0.0180 | 0.0615 | ns |
| Environmental driver | Vertical diffusion rate | Offshore zone | -0.0176 | 0.1588 | ns |
| Ecology model fitted | Coefficient | Macrophyte self shading | -0.0158 | 0.0079 | sig |
| Ecology model fitted | Threshold biomass for zero exploitable stock remaining | Carnivorous zooplankton offshore | 0.0158 | 0.1118 | ns |
| Ecology model fitted | Maximum uptake rate | Demersal fish by pinnipeds | -0.0152 | 0.0225 | sig |
| Ecology model fitted | Nitrification rate coefficient | Lower layer ammonia | 0.0148 | 0.0425 | ns |
| Ecology model fitted | Maximum uptake rate | Carnivore/scavenge feeding benthos larvae by planktivorous fish | -0.0146 | 0.0490 | ns |
| Ecology model fitted | Maximum uptake rate | Carnivorous zooplankton by demersal fish | 0.0143 | 0.1090 | ns |

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|-------------------------|--|--|---------|--------|-----|
| Environmental driver | River volume inflow rate | Inshore zone | 0.0140 | 0.1085 | ns |
| Harvest ratio | Harvest ratio offshore | Carnivorous zooplankton | -0.0139 | 0.0304 | ns |
| Ecology model fixed | Maximum exploitable fraction of stock | Carnivorous zooplankton | -0.0139 | 0.0304 | ns |
| Ecology model fitted | Maximum uptake rate | Omnivorous zooplankton by planktivorous fish larvae | -0.0136 | 0.0541 | ns |
| Environmental driver | Boundary concentration | River nitrate | 0.0135 | 0.0960 | ns |
| Ecology model fixed | Maximum exploitable fraction of stock | Carnivore/scavenge feeding benthos | 0.0134 | 0.0208 | sig |
| Ecology model fitted | Maximum uptake rate | Carnivore/scavenge feeding benthos larvae by omnivorous zooplankton | -0.0132 | 0.2488 | ns |
| Ecology model fitted | Maximum uptake rate | Suspension/deposit feeding benthos larvae by demersal fish larvae | 0.0131 | 0.0944 | ns |
| Biological event driver | Recruitment rate | Suspension/deposit feeding benthos | 0.0131 | 0.0967 | ns |
| Physical configuration | Sediment porosity | Offshore sandy sediments | -0.0130 | 0.0198 | sig |
| Environmental driver | Boundary concentration | Upper layer offshore nitrate | -0.0125 | 0.2323 | ns |
| Environmental driver | Boundary concentration | Lower layer ammonia | -0.0124 | 0.0547 | ns |
| Ecology model fitted | Maximum uptake rate | Demersal fish by birds | -0.0120 | 0.0158 | sig |
| Physical configuration | Sediment porosity | Inshore sandy sediments | -0.0115 | 0.0260 | ns |
| Biological event driver | Spawning rate | Suspension/deposit feeding benthos | -0.0112 | 0.0527 | ns |
| Ecology model fixed | Annual fecundity | Suspension/deposit feeding benthos | -0.0112 | 0.0527 | ns |
| Ecology model fitted | Maximum uptake rate | demersal fish larvae by demersal fish | -0.0111 | 0.0497 | ns |
| Environmental driver | Boundary concentration | Inshore nitrate | 0.0106 | 0.1158 | ns |
| Environmental driver | Boundary volume inflow rate | Inshore zone | 0.0103 | 0.0681 | ns |
| Ecology model fitted | Maximum uptake rate | Planktivorous fish by cetaceans | -0.0102 | 0.0328 | ns |
| Fishing fleet model | Damage mortality rate by fishing gears | Carnivore/scavenge feeding benthos inshore | -0.0097 | 0.0104 | sig |
| Harvest ratio | Harvest ratio offshore | Carnivore/scavenge feeding benthos | 0.0097 | 0.0143 | sig |
| Ecology model fitted | Uptake half saturation coefficient | demersal fish larvae by demersal fish | 0.0089 | 0.0463 | ns |
| Ecology model fitted | Maximum uptake rate | Planktivorous fish larvae by planktivorous fish | 0.0088 | 0.0779 | ns |
| Ecology model fixed | Q10 | Metabolism and microbial rates | 0.0087 | 0.3557 | ns |
| Ecology model fitted | Maximum uptake rate | Planktivorous fish by pinnipeds | -0.0087 | 0.0177 | ns |
| Ecology model fitted | Maximum uptake rate | Carnivore/scavenge feeding benthos larvae by carnivorous zooplankton | -0.0082 | 0.0542 | ns |
| Ecology model fixed | Background metabolic rate coefficient | Carnivore/scavenge feeding benthos larvae | -0.0082 | 0.0525 | ns |

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|-------------------------|------------------------------------|---|---------|--------|-----|
| Environmental driver | Temperature | Inshore zone | 0.0080 | 0.2322 | ns |
| Ecology model fitted | Maximum uptake rate | Demersal fish by cetaceans | -0.0080 | 0.0155 | sig |
| Ecology model fitted | Scaling parameter | Linking demersal fish survey and model abundance | -0.0079 | 0.0500 | ns |
| Physical configuration | Hydraulic conductivity | Offshore muddy sediments | -0.0077 | 0.0096 | sig |
| Environmental driver | Significant wave height | Inshore zone | 0.0074 | 0.0136 | sig |
| Ecology model fitted | Uptake half saturation coefficient | Suspended detritus by suspension/deposit feeding benthos larvae | 0.0074 | 0.0762 | ns |
| Ecology model fixed | Assimilation efficiency | Carnivore/scavenge feeding benthos | 0.0072 | 0.0537 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Demersal fish by pinnipeds | -0.0072 | 0.0100 | sig |
| Ecology model fitted | Uptake half saturation coefficient | Carnivore/scavenge feeding benthos by demersal fish | -0.0071 | 0.0392 | ns |
| Ecology model fitted | Bedding DeAngelis parameter | Cetaceans | -0.0069 | 0.0147 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Planktivorous fish larvae by demersal fish | -0.0066 | 0.0263 | ns |
| Fishing fleet model | Discard rate inshore | Demersal fish | -0.0065 | 0.0643 | ns |
| Ecology model fitted | Maximum uptake rate | Carnivore/scavenge feeding benthos by demersal fish | 0.0062 | 0.0511 | ns |
| Ecology model fitted | Maximum uptake rate | Planktivorous fish larvae by demersal fish | 0.0057 | 0.0283 | ns |
| Environmental driver | Boundary concentration | Upper layer offshore detritus | -0.0057 | 0.0201 | ns |
| Biological event driver | Immigration rate | Migratory fish | -0.0056 | 0.0045 | sig |
| Fishing fleet model | Discard rate inshore | Carnivore/scavenge feeding benthos | -0.0056 | 0.0075 | sig |
| Environmental driver | Temperature | Lower layer offshore | 0.0055 | 0.0489 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Carnivore/scavenge feeding benthos larvae by planktivorous fish | 0.0054 | 0.0454 | ns |
| Ecology model fitted | Disintegration rate | Macrophyte debris to detritus | -0.0053 | 0.0079 | sig |
| Ecology model fitted | Bedding DeAngelis parameter | Birds | -0.0053 | 0.0223 | ns |
| Ecology model fitted | Mineralisation rate coefficient | Labile sediment detritus | -0.0052 | 0.0185 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Planktivorous fish by birds | -0.0052 | 0.0150 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Suspension/deposit feeding benthos larvae by planktivorous fish | 0.0052 | 0.0759 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Planktivorous fish by cetaceans | -0.0051 | 0.0104 | ns |
| Ecology model fitted | Conversion rate | Macrophyte debris to beach-cast | 0.0051 | 0.0079 | sig |
| Physical configuration | Hydraulic conductivity | Offshore sandy sediments | -0.0050 | 0.0066 | sig |
| Ecology model fitted | Uptake half saturation coefficient | Omnivorous zooplankton by planktivorous fish larvae | 0.0050 | 0.0484 | ns |

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| Ecology model fitted | Uptake half saturation coefficient | Carnivore/scavenge feeding benthos larvae by carnivorous zooplankton | 0.0050 | 0.0478 | ns |
| Ecology model fitted | Bedding DeAngelis parameter | Pinnipeds | -0.0050 | 0.0070 | sig |
| Ecology model fixed | Background metabolic rate coefficient | Carnivore/scavenge feeding benthos | -0.0050 | 0.0114 | ns |
| Ecology model fitted | Maximum uptake rate | Migratory fish by cetaceans | -0.0048 | 0.0045 | sig |
| Environmental driver | Boundary concentration | Inshore phytoplankton | 0.0046 | 0.0313 | ns |
| Harvest ratio | Harvest ratio inshore | Carnivore/scavenge feeding benthos | 0.0044 | 0.0066 | sig |
| Ecology model fitted | Uptake half saturation coefficient | Suspension/deposit feeding benthos larvae by carnivorous zooplankton | 0.0043 | 0.1502 | ns |
| Ecology model fitted | Density dependent mortality coefficient | Cetaceans | -0.0043 | 0.0057 | sig |
| Ecology model fixed | Maximum exploitable fraction of stock | Cetaceans | -0.0042 | 0.0066 | sig |
| Harvest ratio | Harvest ratio offshore | Cetaceans | -0.0042 | 0.0066 | sig |
| Ecology model fitted | Maximum uptake rate | Planktivorous fish by demersal fish | 0.0041 | 0.2413 | ns |
| Physical configuration | Sediment porosity | Inshore coarse sediments | -0.0041 | 0.0094 | ns |
| Fishing fleet model | Exponent | Demersal fish quota-limited undersize vs nitrogen mass | 0.0040 | 0.0364 | ns |
| Ecology model fixed | Saturation light intensity for uptake | Nutrient by macrophytes | -0.0040 | 0.0032 | sig |
| Fishing fleet model | Discard rate all areas | Demersal fish all areas | -0.0040 | 0.2204 | ns |
| Ecology model fitted | Maximum uptake rate | Corpses by carnivore/scavenge feeding benthos | 0.0039 | 0.0118 | ns |
| Biological event driver | Spawning rate | Carnivore/scavenge feeding benthos | -0.0039 | 0.0442 | ns |
| Ecology model fixed | Annual fecundity | Carnivore/scavenge feeding benthos | -0.0039 | 0.0443 | ns |
| Ecology model fitted | Density dependent mortality coefficient | Birds | -0.0038 | 0.0093 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Demersal fish by birds | 0.0038 | 0.0082 | ns |
| Ecology model fitted | Conversion rate coefficient | Corpses to labile sediment detritus | -0.0036 | 0.0065 | sig |
| Ecology model fitted | Uptake half saturation coefficient | Planktivorous fish larvae by carnivorous zooplankton | -0.0036 | 0.0168 | ns |
| Biological event driver | Spawning rate | Planktivorous fish | 0.0036 | 0.0537 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Phytoplankton by suspension/deposit feeding benthos | 0.0036 | 0.0935 | ns |
| Ecology model fixed | Annual fecundity | Planktivorous fish | 0.0036 | 0.0537 | ns |
| Environmental driver | Volume outflow rate | Inshore zone | 0.0035 | 0.0277 | ns |
| Biological event driver | Recruitment rate | Carnivore/scavenge feeding benthos | 0.0035 | 0.0091 | ns |
| Fishing fleet model | Discard rate offshore | Carnivore/scavenge feeding benthos | -0.0034 | 0.0045 | sig |
| Ecology model fixed | Background metabolic rate coefficient | Suspension/deposit feeding benthos larvae | 0.0033 | 0.0504 | ns |

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| Ecology model fitted | Density dependent mortality coefficient | Omnivorous zooplankton | 0.0032 | 0.0238 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Migratory fish by pinnipeds | -0.0032 | 0.0039 | sig |
| Ecology model fitted | Maximum uptake rate | Suspension/deposit feeding benthos by carnivore/scavenge feeding benthos | -0.0032 | 0.0872 | ns |
| Physical configuration | Proportion of depth range occupied | Macrophytes inshore | -0.0031 | 0.0027 | sig |
| Ecology model fitted | Maximum uptake rate | Suspension/deposit feeding benthos larvae by planktivorous fish larvae | -0.0031 | 0.0155 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Corpses by carnivore/scavenge feeding benthos | -0.0030 | 0.0073 | ns |
| Environmental driver | Boundary concentration | Upper layer offshore ammonia | -0.0029 | 0.0180 | ns |
| Fishing fleet model | Damage mortality rate by fishing gears | Suspension/deposit feeding benthos offshore | -0.0028 | 0.0182 | ns |
| Ecology model fitted | Maximum uptake rate | Planktivorous fish larvae by carnivorous zooplankton | 0.0027 | 0.0184 | ns |
| Environmental driver | Boundary concentration | Inshore detritus | 0.0026 | 0.0186 | ns |
| Ecology model fitted | Maximum uptake rate | Carnivorous zooplankton by birds | -0.0026 | 0.0069 | ns |
| Ecology model fitted | Density dependent mortality coefficient | Pinnipeds | 0.0026 | 0.0144 | ns |
| Environmental driver | Boundary concentration | River ammonia | 0.0025 | 0.0162 | ns |
| Ecology model fitted | Threshold biomass for zero exploitable stock remaining | Cetaceans offshore | 0.0025 | 0.0084 | ns |
| Ecology model fitted | Sinking rate coefficient | Upper layer suspended detritus | 0.0025 | 0.0466 | ns |
| Fishing fleet model | Damage mortality rate by fishing gears | Carnivore/scavenge feeding benthos offshore | -0.0024 | 0.0163 | ns |
| Ecology model fitted | Maximum uptake rate | Migratory fish by birds | -0.0024 | 0.0025 | sig |
| Ecology model fitted | Conversion rate coefficient | Discards to corpses | -0.0023 | 0.0026 | sig |
| Environmental driver | Volume exchange rate | Offshore to inshore zone | 0.0023 | 0.0104 | ns |
| Environmental driver | Boundary volume inflow rate | Inshore zone | 0.0023 | 0.0190 | ns |
| Ecology model fitted | Wave height dependent conversion rate | Macrophytes to macrophyte debris | 0.0022 | 0.0059 | ns |
| Physical configuration | Hydraulic conductivity | Inshore sandy sediments | -0.0022 | 0.0039 | sig |
| Ecology model fitted | Uptake half saturation coefficient | Suspension/deposit feeding benthos larvae by planktivorous fish larvae | 0.0021 | 0.0142 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Demersal fish by cetaceans | 0.0020 | 0.0058 | ns |
| Ecology model fitted | Maximum uptake rate | Demersal fish by demersal fish | -0.0020 | 0.0122 | ns |
| Ecology model fitted | Maximum uptake rate | Discards by birds | 0.0020 | 0.0028 | sig |
| Ecology model fixed | Background metabolic rate coefficient | Demersal fish larvae | -0.0019 | 0.0100 | ns |
| Environmental driver | Boundary concentration | Lower layer phytoplankton | 0.0018 | 0.0185 | ns |

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|-------------------------|--|--|---------|--------|-----|
| Ecology model fixed | Inedible biomass inshore | Carnivorous zooplankton | -0.0018 | 0.0244 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Carnivore/scavenge feeding benthos larvae by demersal fish larvae | -0.0017 | 0.0096 | ns |
| Environmental driver | Atmospheric deposition rate | Inshore nitrate | 0.0017 | 0.0073 | ns |
| Ecology model fitted | Density dependent mortality coefficient | Suspension/deposit feeding benthos larvae | 0.0017 | 0.0186 | ns |
| Ecology model fitted | Maximum uptake rate | Carnivore/scavenge feeding benthos larvae by demersal fish larvae | 0.0016 | 0.0105 | ns |
| Ecology model fixed | Assimilation efficiency | Migratory fish | 0.0016 | 0.0025 | sig |
| Ecology model fitted | Maximum uptake rate | Migratory fish by pinnipeds | 0.0016 | 0.0061 | ns |
| Ecology model fitted | Active migration coefficient | Planktivorous fish | 0.0015 | 0.0140 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Phytoplankton by carnivore/scavenge feeding benthos larvae | 0.0015 | 0.1828 | ns |
| Environmental driver | Boundary concentration | Inshore ammonia | 0.0014 | 0.0102 | ns |
| Ecology model fitted | Threshold biomass for zero exploitable stock remaining | Carnivore/scavenge feeding benthos inshore | -0.0014 | 0.0019 | sig |
| Ecology model fitted | Uptake half saturation coefficient | Demersal fish by demersal fish | 0.0014 | 0.0088 | ns |
| Biological event driver | Recruitment rate | Planktivorous fish | -0.0013 | 0.0035 | ns |
| Ecology model fitted | Remobilisation parameter | Refractory to labile sediment detritus | 0.0013 | 0.0797 | ns |
| Fishing fleet model | Coefficient | Demersal fish non-quota undersize vs nitrogen mass | -0.0013 | 0.0274 | ns |
| Environmental driver | Boundary concentration | Upper layer offshore phytoplankton | 0.0013 | 0.0328 | ns |
| Ecology model fitted | Maximum uptake rate | Phytoplankton by suspension/deposit feeding benthos larvae | -0.0012 | 0.0064 | ns |
| Physical configuration | Sediment porosity | Offshore muddy sediments | -0.0012 | 0.0010 | sig |
| Ecology model fitted | Carbon exudation rate | Macrophytes | 0.0012 | 0.0033 | ns |
| Ecology model fixed | Maximum exploitable fraction of stock | Suspension/deposit feeding benthos | -0.0011 | 0.0007 | sig |
| Harvest ratio | Harvest ratio inshore | Planktivorous fish | 0.0011 | 0.0786 | ns |
| Ecology model fitted | Maximum uptake rate | Corpses by birds | -0.0011 | 0.0031 | ns |
| Physical configuration | Sediment porosity | Offshore coarse sediments | -0.0011 | 0.0018 | sig |
| Ecology model fitted | Uptake half saturation coefficient | Suspension/deposit feeding benthos by carnivore/scavenge feeding benthos | -0.0010 | 0.0206 | ns |
| Ecology model fitted | Maximum uptake rate | Carbon by macrophytes | 0.0010 | 0.0035 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Phytoplankton by suspension/deposit feeding benthos larvae | 0.0010 | 0.0053 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Migratory fish by cetaceans | 0.0010 | 0.0018 | sig |

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|-------------------------|--|---|---------|--------|-----|
| Ecology model fitted | Maximum uptake rate | Omnivorous zooplankton by migratory fish | 0.0010 | 0.0075 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Discards by birds | -0.0010 | 0.0016 | sig |
| Ecology model fitted | Maximum uptake rate | Sediment detritus by suspension/deposit feeding benthos | 0.0010 | 0.0910 | ns |
| Ecology model fitted | Threshold biomass for zero exploitable stock remaining | Carnivore/scavenge feeding benthos offshore | -0.0009 | 0.0013 | sig |
| Ecology model fitted | Uptake half saturation coefficient | Suspension/deposit feeding benthos larvae by omnivorous zooplankton | -0.0009 | 0.0056 | ns |
| Ecology model fitted | Maximum uptake rate | Suspension/deposit feeding benthos larvae by omnivorous zooplankton | 0.0008 | 0.0063 | ns |
| Environmental driver | Atmospheric deposition rate | Offshore nitrate | 0.0008 | 0.0065 | ns |
| Ecology model fitted | Threshold biomass for zero exploitable stock remaining | Demersal fish offshore | 0.0008 | 0.0054 | ns |
| Biological event driver | Emigration rate | Migratory fish | -0.0008 | 0.0063 | ns |
| Environmental driver | Atmospheric deposition rate | Inshore ammonia | 0.0008 | 0.0053 | ns |
| Ecology model fixed | Maximum exploitable fraction of stock | Migratory fish | -0.0007 | 0.0015 | sig |
| Harvest ratio | Harvest ratio offshore | Migratory fish | -0.0007 | 0.0014 | sig |
| Environmental driver | Boundary concentration | Lower layer detritus | 0.0007 | 0.0523 | ns |
| Ecology model fitted | Maximum uptake rate | Corpses by demersal fish | 0.0007 | 0.0040 | ns |
| Fishing fleet model | Exponent | Demersal fish non-quota proportion in catch vs nitrogen mass | 0.0007 | 0.0075 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Corpses by demersal fish | -0.0006 | 0.0032 | ns |
| Ecology model fixed | Inedible biomass offshore | Carnivorous zooplankton | -0.0006 | 0.0096 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Suspended detritus by omnivorous zooplankton | -0.0006 | 0.0045 | ns |
| Ecology model fitted | Mineralisation rate scaling parameter | Refractory sediment detritus | 0.0005 | 0.0104 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Discards by demersal fish | -0.0005 | 0.0023 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Migratory fish by demersal fish | -0.0005 | 0.0023 | ns |
| Ecology model fitted | Maximum uptake rate | Migratory fish by demersal fish | 0.0005 | 0.0025 | ns |
| Environmental driver | Natural disturbance rate | Offshore sandy sediments | 0.0005 | 0.0029 | ns |
| Physical configuration | Physical disturbance depth | Offshore sandy sediments | 0.0005 | 0.0029 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Omnivorous zooplankton by migratory fish | -0.0005 | 0.0074 | ns |
| Harvest ratio | Harvest ratio offshore | Suspension/deposit feeding benthos | -0.0004 | 0.0004 | sig |
| Fishing fleet model | Coefficient | Demersal fish non-quota proportion in catch vs nitrogen mass | -0.0004 | 0.0112 | ns |

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|------------------------|--|--|---------|--------|-----|
| Fishing fleet model | Damage mortality rate by fishing gears | Suspension/deposit feeding benthos inshore | 0.0004 | 0.0029 | ns |
| Ecology model fixed | Background metabolic rate coefficient | Planktivorous fish larvae | 0.0004 | 0.0021 | ns |
| Ecology model fitted | Maximum uptake rate | Nitrate by macrophytes | 0.0004 | 0.0007 | sig |
| Fishing fleet model | Exponent | Demersal fish non-quota undersize vs nitrogen mass | 0.0004 | 0.0053 | ns |
| Ecology model fitted | Maximum uptake rate | Discards by demersal fish | 0.0004 | 0.0023 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Nitrate by macrophytes | -0.0004 | 0.0006 | sig |
| Ecology model fitted | Threshold biomass for zero exploitable stock remaining | Demersal fish inshore | 0.0004 | 0.0030 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Migratory fish by birds | -0.0004 | 0.0012 | ns |
| Ecology model fitted | Maximum uptake rate | Planktivorous fish larvae by migratory fish | 0.0004 | 0.0016 | ns |
| Ecology model fitted | Maximum uptake rate | Carnivore/scavenge feeding benthos larvae by planktivorous fish larvae | -0.0004 | 0.0026 | ns |
| Environmental driver | Atmospheric deposition rate | Offshore ammonia | -0.0004 | 0.0038 | ns |
| Fishing fleet model | Discard rate offshore | Migratory fish | 0.0004 | 0.0005 | sig |
| Ecology model fitted | Uptake half saturation coefficient | Planktivorous fish larvae by migratory fish | -0.0004 | 0.0014 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Carnivorous zooplankton by migratory fish | -0.0003 | 0.0019 | ns |
| Ecology model fitted | Maximum uptake rate | Carnivorous zooplankton by cetaceans | 0.0003 | 0.0009 | ns |
| Fishing fleet model | Discard rate offshore | Planktivorous fish | 0.0003 | 0.0005 | sig |
| Ecology model fitted | Uptake half saturation coefficient | Carnivore/scavenge feeding benthos larvae by planktivorous fish larvae | 0.0003 | 0.0025 | ns |
| Physical configuration | Hydraulic conductivity | Inshore coarse sediments | -0.0003 | 0.0006 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Planktivorous fish by pinnipeds | -0.0003 | 0.0092 | ns |
| Ecology model fitted | Maximum uptake rate | Carnivorous zooplankton by migratory fish | 0.0003 | 0.0019 | ns |
| Physical configuration | Hydraulic conductivity | Inshore muddy sediments | -0.0003 | 0.0003 | sig |
| Ecology model fitted | Uptake half saturation coefficient | Corpses by birds | 0.0002 | 0.0017 | ns |
| Ecology model fitted | Density dependent mortality coefficient | Planktivorous fish larvae | 0.0002 | 0.0011 | ns |
| Ecology model fitted | Maximum uptake rate | Suspended detritus by omnivorous zooplankton | 0.0002 | 0.0028 | ns |
| Ecology model fitted | Density dependent mortality coefficient | Demersal fish larvae | -0.0002 | 0.0007 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Demersal fish larvae by migratory fish | 0.0002 | 0.0009 | ns |
| Ecology model fitted | Maximum uptake rate | Demersal fish larvae by migratory fish | -0.0002 | 0.0009 | ns |
| Ecology model fitted | Maximum uptake rate | Omnivorous zooplankton by cetaceans | 0.0002 | 0.0007 | ns |

| | | | | | |
|------------------------|--|---|-----------|----------|-----|
| Ecology model fitted | Maximum uptake rate | Ammonia by macrophytes | 0.0002 | 0.0003 | sig |
| Ecology model fitted | Uptake half saturation coefficient | Carnivorous zooplankton by birds&mammala | 0.0002 | 0.0036 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Ammonia by macrophytes | -0.0002 | 0.0002 | sig |
| Ecology model fitted | Uptake half saturation coefficient | Corpses by pinnipeds | -0.0002 | 0.0007 | ns |
| Physical configuration | Sediment porosity | Inshore muddy sediments | -0.0002 | 0.0002 | sig |
| Fishing fleet model | Offal as proportion of live weight | All guilds | 0.0001 | 0.0002 | sig |
| Ecology model fitted | Maximum uptake rate | Corpses by pinnipeds | 0.0001 | 0.0012 | ns |
| Ecology model fitted | Threshold biomass for zero exploitable stock remaining | Suspension/deposit feeding benthos offshore | -0.0001 | 0.0002 | sig |
| Harvest ratio | Harvest ratio inshore | Suspension/deposit feeding benthos | -0.0001 | 0.0002 | sig |
| Ecology model fitted | Uptake half saturation coefficient | Carnivorous zooplankton by cetaceans | -0.0001 | 0.0003 | ns |
| Ecology model fixed | N:C molar ratio maximum | Macrophytes | -0.0001 | 0.0034 | ns |
| Environmental driver | Natural disturbance rate | Inshore sandy sediments | 0.0001 | 0.0007 | ns |
| Ecology model fitted | Maximum uptake rate | Suspension/deposit feeding benthos larvae by migratory fish | 0.0001 | 0.0003 | ns |
| Fishing fleet model | Penetration depth by fishing gears | Offshore muddy sediments | -0.0001 | 0.0012 | ns |
| Ecology model fixed | Maximum exploitable fraction of stock | Pinnipeds | 0.0001 | 0.0007 | ns |
| Physical configuration | Physical disturbance depth | Inshore sandy sediments | 9.53E-05 | 6.98E-04 | ns |
| Fishing fleet model | Processing at sea rate offshore | Demersal fish | 9.17E-05 | 1.59E-04 | sig |
| Fishing fleet model | Penetration depth by fishing gears | Offshore sandy sediments | 8.92E-05 | 6.41E-04 | ns |
| Physical configuration | Hydraulic conductivity | Offshore coarse sediments | -8.39E-05 | 1.27E-04 | sig |
| Ecology model fitted | Uptake half saturation coefficient | Suspension/deposit feeding benthos larvae by migratory fish | -8.20E-05 | 1.58E-04 | sig |
| Harvest ratio | Harvest ratio offshore | Pinnipeds | 8.06E-05 | 3.95E-04 | ns |
| Ecology model fixed | N:C molar ratio minimum | Macrophytes | -7.65E-05 | 9.42E-05 | sig |
| Ecology model fitted | Uptake half saturation coefficient | Omnivorous zooplankton by cetaceans | -7.35E-05 | 1.71E-04 | ns |
| Ecology model fitted | Threshold biomass for zero exploitable stock remaining | Planktivorous fish inshore | -7.08E-05 | 7.75E-04 | ns |
| Fishing fleet model | Abrasion rate by fishing gears | Offshore muddy sediments | -7.02E-05 | 1.31E-03 | ns |
| Harvest ratio | Harvest ratio inshore | Pinnipeds | 6.63E-05 | 4.27E-04 | ns |
| Ecology model fitted | Density dependent mortality coefficient | Carnivore/scavenge feeding benthos larvae | -6.56E-05 | 2.25E-04 | ns |
| Ecology model fitted | Threshold biomass for zero exploitable stock remaining | Planktivorous fish offshore | -6.44E-05 | 1.26E-03 | ns |

| | | | | | |
|------------------------|--|---|-----------|----------|-----|
| Ecology model fitted | Threshold biomass for zero exploitable stock remaining | Suspension/deposit feeding benthos inshore | -5.55E-05 | 8.85E-05 | sig |
| Fishing fleet model | Discard rate inshore | Planktivorous fish | 5.50E-05 | 9.29E-05 | sig |
| Physical configuration | Physical disturbance depth | Inshore muddy sediments | -5.40E-05 | 4.64E-05 | sig |
| Ecology model fitted | Maximum uptake rate | Discards by pinnipeds | 5.28E-05 | 1.99E-04 | ns |
| Harvest ratio | Harvest ratio offshore | Birds | -4.82E-05 | 7.36E-05 | sig |
| Environmental driver | Natural disturbance rate | Offshore muddy sediments | -4.74E-05 | 2.67E-03 | ns |
| Fishing fleet model | Processing at sea rate inshore | Carnivore/scavenge feeding benthos | -4.68E-05 | 6.27E-05 | sig |
| Ecology model fitted | Uptake half saturation coefficient | Sediment detritus by suspension/deposit feeding benthos | -4.34E-05 | 9.26E-05 | ns |
| Ecology model fitted | Threshold biomass for zero exploitable stock remaining | Pinnipeds inshore | -4.33E-05 | 1.12E-04 | ns |
| Physical configuration | Bioturbation depth | Offshore muddy sediments | -4.32E-05 | 6.34E-05 | sig |
| Ecology model fitted | Uptake half saturation coefficient | Pinnipeds by cetaceans | -4.20E-05 | 9.34E-05 | ns |
| Ecology model fitted | Active migration coefficient | Birds | -3.84E-05 | 4.19E-05 | sig |
| Ecology model fitted | Uptake half saturation coefficient | Carnivore/scavenge feeding benthos larvae by migratory fish | -3.73E-05 | 9.90E-05 | ns |
| Ecology model fitted | Maximum uptake rate | Macrophytes by carnivorous/scavenge feeding benthos | 3.70E-05 | 1.46E-04 | ns |
| Physical configuration | Physical disturbance depth | Offshore muddy sediments | -3.64E-05 | 2.65E-03 | ns |
| Ecology model fitted | Threshold biomass for zero exploitable stock remaining | Pinnipeds offshore | -3.49E-05 | 6.66E-05 | sig |
| Fishing fleet model | Abrasion rate by fishing gears | Offshore coarse sediments | -3.43E-05 | 1.04E-04 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Discards by pinnipeds | -3.27E-05 | 6.75E-05 | ns |
| Ecology model fitted | Nitrification rate coefficient | Sediment porewater ammonia | -3.26E-05 | 4.17E-05 | sig |
| Ecology model fitted | Denitrification rate coefficient | Lower layer nitrate | -3.24E-05 | 6.06E-05 | sig |
| Physical configuration | Bioturbation depth | Inshore coarse sediments | -3.23E-05 | 6.01E-05 | sig |
| Ecology model fitted | Active migration coefficient | Demersal fish | -3.15E-05 | 4.05E-03 | ns |
| Fishing fleet model | Abrasion rate by fishing gears | Offshore rock | -3.13E-05 | 3.94E-05 | sig |
| Ecology model fitted | Uptake half saturation coefficient | Macrophyte debris by carnivorous/scavenge feeding benthos | 3.09E-05 | 1.14E-04 | ns |
| Physical configuration | Physical disturbance depth | Offshore coarse sediments | -2.92E-05 | 4.56E-05 | sig |
| Ecology model fitted | Maximum uptake rate | Pinnipeds by cetaceans | 2.91E-05 | 1.41E-04 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Macrophytes by carnivorous/scavenge feeding benthos | -2.83E-05 | 4.13E-05 | sig |

| | | | | | |
|------------------------|--|--|-----------|----------|-----|
| Ecology model fixed | Background metabolic rate coefficient | Migratory fish | -2.83E-05 | 3.39E-04 | ns |
| Fishing fleet model | Penetration depth by fishing gears | Offshore coarse sediments | -2.65E-05 | 3.43E-05 | sig |
| Ecology model fitted | Density dependent mortality coefficient | Migratory fish | 2.56E-05 | 1.39E-04 | ns |
| Fishing fleet model | Penetration depth by fishing gears | Inshore muddy sediments | -2.56E-05 | 4.34E-05 | sig |
| Fishing fleet model | Penetration depth by fishing gears | Inshore coarse sediments | -2.55E-05 | 5.55E-05 | ns |
| Fishing fleet model | Abrasion rate by fishing gears | Offshore sandy sediments | 2.50E-05 | 4.74E-04 | ns |
| Environmental driver | Natural disturbance rate | Offshore coarse sediments | -2.39E-05 | 3.98E-05 | sig |
| Fishing fleet model | Abrasion rate by fishing gears | Inshore sandy sediments | 2.34E-05 | 1.54E-04 | ns |
| Ecology model fitted | Uptake half saturation coefficient | Discards by cetaceans | -2.34E-05 | 1.20E-05 | sig |
| Harvest ratio | Harvest ratio inshore | Migratory fish | 2.34E-05 | 1.30E-04 | ns |
| Ecology model fitted | Sinking rate coefficient | Lower layer suspended detritus | -2.29E-05 | 6.12E-05 | ns |
| Fishing fleet model | Processing at sea rate inshore | Demersal fish | 2.25E-05 | 8.11E-05 | ns |
| Fishing fleet model | Discard rate inshore | Pinnipeds | 2.25E-05 | 1.26E-04 | ns |
| Physical configuration | Bioturbation depth | Offshore coarse sediments | 2.12E-05 | 8.27E-05 | ns |
| Ecology model fitted | Mineralisation rate sensitivity to grain size | Labile sediment detritus | 2.12E-05 | 9.89E-05 | ns |
| Ecology model fitted | Maximum uptake rate | Discards by cetaceans | 2.10E-05 | 8.24E-05 | ns |
| Ecology model fitted | Denitrification rate sensitivity to grain size | Sediment porewater nitrate | -2.09E-05 | 4.82E-05 | ns |
| Physical configuration | Bioturbation depth | Inshore sandy sediments | -2.06E-05 | 4.74E-05 | ns |
| Environmental driver | Natural disturbance rate | Inshore coarse sediments | -1.91E-05 | 2.74E-05 | sig |
| Physical configuration | Bioturbation depth | Offshore sandy sediments | -1.89E-05 | 4.62E-05 | ns |
| Environmental driver | Natural disturbance rate | Inshore muddy sediments | -1.74E-05 | 5.59E-05 | ns |
| Ecology model fitted | Active migration coefficient | Pinnipeds | -1.61E-05 | 4.91E-05 | ns |
| Fishing fleet model | Discard rate offshore | Pinnipeds | -1.59E-05 | 3.87E-05 | ns |
| Fishing fleet model | Discard rate offshore | Suspension/deposit feeding benthos | -1.47E-05 | 4.46E-05 | ns |
| Ecology model fitted | Nitrification rate sensitivity to grain size | Sediment porewater ammonia | -1.41E-05 | 4.12E-05 | ns |
| Fishing fleet model | Discard rate inshore | Migratory fish | 1.33E-05 | 4.45E-05 | ns |
| Fishing fleet model | Discard rate inshore | Suspension/deposit feeding benthos | -1.29E-05 | 3.39E-05 | ns |
| Fishing fleet model | Abrasion rate by fishing gears | Inshore rock | -1.29E-05 | 5.84E-05 | ns |
| Physical configuration | Hydraulic conductivity | Reference value for sediment-dependent processes | -1.28E-05 | 4.43E-05 | ns |

| | | | | | |
|------------------------|--|---|-----------|----------|----|
| Ecology model fitted | Threshold biomass for zero exploitable stock remaining | Birds offshore | -1.20E-05 | 5.31E-05 | ns |
| Ecology model fitted | Mineralisation rate coefficient | Suspended detritus | 1.20E-05 | 1.35E-04 | ns |
| Ecology model fitted | Maximum uptake rate | Carnivore/scavenge feeding benthos larvae by migratory fish | -1.14E-05 | 6.94E-05 | ns |
| Ecology model fitted | Nitrification rate coefficient | Upper layer ammonia | -1.12E-05 | 2.48E-05 | ns |
| Ecology model fitted | Active migration coefficient | Migratory fish | -1.05E-05 | 1.51E-04 | ns |
| Ecology model fitted | Denitrification rate coefficient | Upper layer nitrate | -9.96E-06 | 3.57E-05 | ns |
| Fishing fleet model | Discard rate inshore | Birds | -8.70E-06 | 4.93E-05 | ns |
| Ecology model fitted | Threshold biomass for zero exploitable stock remaining | Migratory fish inshore | -8.39E-06 | 4.98E-05 | ns |
| Fishing fleet model | Discard rate offshore | Birds | -7.30E-06 | 5.25E-05 | ns |
| Fishing fleet model | Penetration depth by fishing gears | Inshore sandy sediments | 7.26E-06 | 1.24E-04 | ns |
| Ecology model fitted | Threshold biomass for zero exploitable stock remaining | Migratory fish offshore | 6.46E-06 | 5.60E-05 | ns |
| Ecology model fitted | Threshold biomass for zero exploitable stock remaining | Birds inshore | -3.96E-06 | 3.57E-05 | ns |
| Fishing fleet model | Abrasion rate by fishing gears | Inshore muddy sediments | 3.19E-06 | 5.12E-05 | ns |
| Ecology model fixed | Maximum exploitable fraction of stock | Birds | -3.10E-06 | 2.42E-05 | ns |
| Fishing fleet model | Abrasion rate by fishing gears | Inshore coarse sediments | 3.00E-06 | 6.66E-05 | ns |
| Fishing fleet model | Processing at sea rate offshore | Carnivore/scavenge feeding benthos | -2.81E-06 | 5.54E-05 | ns |
| Ecology model fitted | Active migration coefficient | Cetaceans | 2.70E-06 | 5.34E-05 | ns |
| Ecology model fitted | Maximum uptake rate | Macrophyte debris by carnivorous/scavenge feeding benthos | 1.61E-06 | 7.35E-05 | ns |
| Physical configuration | Bioturbation depth | Inshore muddy sediments | 1.40E-06 | 7.28E-05 | ns |
| Physical configuration | Physical disturbance depth | Inshore coarse sediments | -8.58E-07 | 4.76E-05 | ns |
| Harvest ratio | Harvest ratio inshore | Birds | 3.85E-07 | 4.78E-05 | ns |
| Environmental driver | Upwelling rate | Offshore zone | NA | NA | NA |
| Environmental driver | Boundary concentration | River detritus | NA | NA | NA |
| Harvest ratio | Harvest ratio inshore | Carnivorous zooplankton | NA | NA | NA |
| Harvest ratio | Harvest ratio inshore | Cetaceans | NA | NA | NA |
| Harvest ratio | Harvest ratio inshore | Macrophytes | NA | NA | NA |
| Harvest ratio | Harvest ratio offshore | Macrophytes | NA | NA | NA |

| | | | | | |
|----------------------|------------------------------------|---|----|----|----|
| Fishing fleet model | Discard rate inshore | Carnivorous zooplankton | NA | NA | NA |
| Fishing fleet model | Discard rate offshore | Carnivorous zooplankton | NA | NA | NA |
| Fishing fleet model | Discard rate inshore | Cetaceans | NA | NA | NA |
| Fishing fleet model | Discard rate inshore | Macrophytes | NA | NA | NA |
| Fishing fleet model | Discard rate offshore | Macrophytes | NA | NA | NA |
| Fishing fleet model | Processing at sea rate inshore | Planktivorous fish | NA | NA | NA |
| Fishing fleet model | Processing at sea rate offshore | Planktivorous fish | NA | NA | NA |
| Fishing fleet model | Processing at sea rate inshore | Migratory fish | NA | NA | NA |
| Fishing fleet model | Processing at sea rate offshore | Migratory fish | NA | NA | NA |
| Fishing fleet model | Processing at sea rate inshore | Suspension/deposit feeding benthos | NA | NA | NA |
| Fishing fleet model | Processing at sea rate offshore | Suspension/deposit feeding benthos | NA | NA | NA |
| Fishing fleet model | Processing at sea rate inshore | Carnivorous zooplankton | NA | NA | NA |
| Fishing fleet model | Processing at sea rate offshore | Carnivorous zooplankton | NA | NA | NA |
| Fishing fleet model | Processing at sea rate inshore | Birds | NA | NA | NA |
| Fishing fleet model | Processing at sea rate offshore | Birds | NA | NA | NA |
| Fishing fleet model | Processing at sea rate inshore | Pinnipeds | NA | NA | NA |
| Fishing fleet model | Processing at sea rate offshore | Pinnipeds | NA | NA | NA |
| Fishing fleet model | Processing at sea rate inshore | Cetaceans | NA | NA | NA |
| Fishing fleet model | Processing at sea rate offshore | Cetaceans | NA | NA | NA |
| Fishing fleet model | Processing at sea rate inshore | Macrophytes | NA | NA | NA |
| Fishing fleet model | Processing at sea rate offshore | Macrophytes | NA | NA | NA |
| Fishing fleet model | Penetration depth by fishing gears | Inshore rock | NA | NA | NA |
| Fishing fleet model | Penetration depth by fishing gears | Offshore rock | NA | NA | NA |
| Ecology model fitted | Maximum uptake rate | Suspension/deposit feeding benthos by birds | NA | NA | NA |
| Ecology model fitted | Uptake half saturation coefficient | Suspension/deposit feeding benthos by birds | NA | NA | NA |
| Ecology model fitted | Maximum uptake rate | Carnivore/scavenge feeding benthos by birds | NA | NA | NA |
| Ecology model fitted | Uptake half saturation coefficient | Carnivore/scavenge feeding benthos by birds | NA | NA | NA |
| Ecology model fitted | Maximum uptake rate | Carnivorous zooplankton by pinnipeds | NA | NA | NA |
| Ecology model fitted | Uptake half saturation coefficient | Carnivorous zooplankton by pinnipeds | NA | NA | NA |
| Ecology model fitted | Maximum uptake rate | Suspension/deposit feeding benthos by pinnipeds | NA | NA | NA |

| | | | | | |
|----------------------|--|---|----|----|----|
| Ecology model fitted | Uptake half saturation coefficient | Suspension/deposit feeding benthos by pinnipeds | NA | NA | NA |
| Ecology model fitted | Maximum uptake rate | Carnivore/scavenge feeding benthos by pinnipeds | NA | NA | NA |
| Ecology model fitted | Uptake half saturation coefficient | Carnivore/scavenge feeding benthos by pinnipeds | NA | NA | NA |
| Ecology model fitted | Maximum uptake rate | Birds by pinnipeds | NA | NA | NA |
| Ecology model fitted | Uptake half saturation coefficient | Birds by pinnipeds | NA | NA | NA |
| Ecology model fitted | Maximum uptake rate | Suspension/deposit feeding benthos by cetaceans | NA | NA | NA |
| Ecology model fitted | Uptake half saturation coefficient | Suspension/deposit feeding benthos by cetaceans | NA | NA | NA |
| Ecology model fitted | Maximum uptake rate | Carnivore/scavenge feeding benthos by cetaceans | NA | NA | NA |
| Ecology model fitted | Uptake half saturation coefficient | Carnivore/scavenge feeding benthos by cetaceans | NA | NA | NA |
| Ecology model fitted | Maximum uptake rate | Birds by cetaceans | NA | NA | NA |
| Ecology model fitted | Uptake half saturation coefficient | Birds by cetaceans | NA | NA | NA |
| Ecology model fitted | Threshold biomass for zero exploitable stock remaining | Carnivorous zooplankton inshore | NA | NA | NA |
| Ecology model fitted | Threshold biomass for zero exploitable stock remaining | Cetaceans inshore | NA | NA | NA |
| Ecology model fitted | Threshold biomass for zero exploitable stock remaining | Macrophytes inshore | NA | NA | NA |
| Ecology model fixed | Maximum exploitable fraction of stock | Macrophytes | NA | NA | NA |

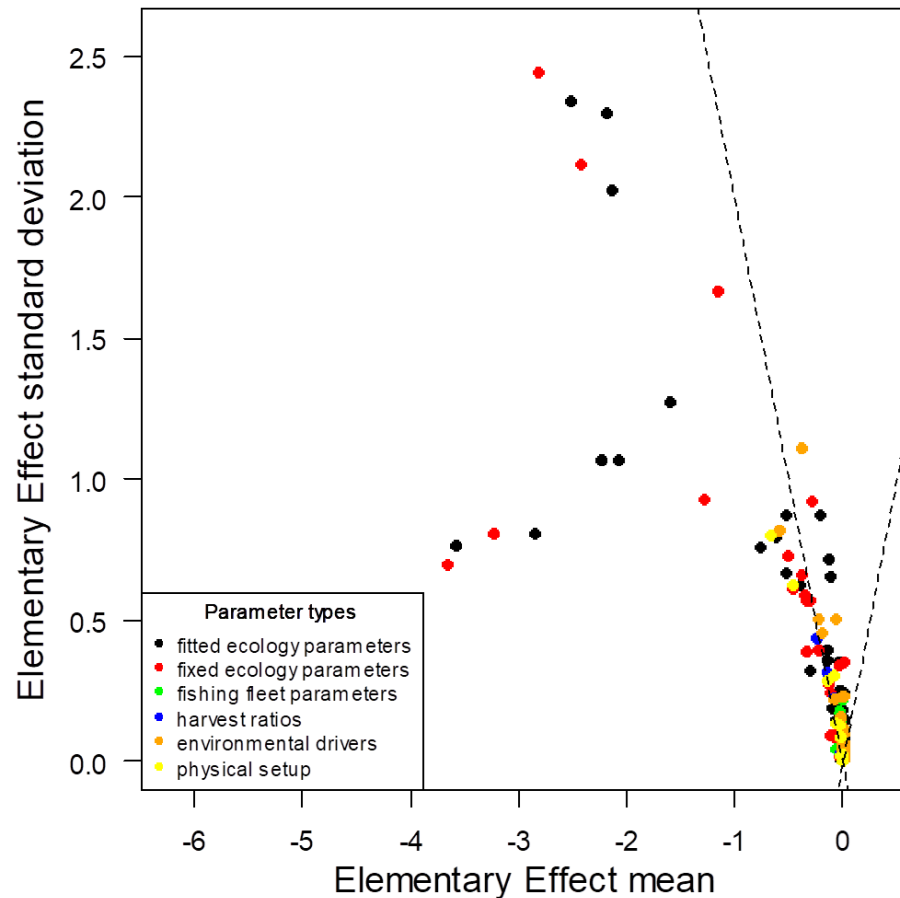


FIGURE 13 Means (EE_mean) and standard deviations (EE_sd) of the distributions of elementary effects of parameters in the sensitivity analysis of the 1970-1999 North Sea model with respect to the likelihood of the observed target data. Black symbols indicate ecology model parameters which were optimized by simulated annealing; red symbols indicate the fixed parameters which were not optimized; green symbols indicate fishing fleet model parameters; blue symbols indicate harvest ratios; orange symbols indicate environmental and biological event drivers; yellow symbols indicate the physical setup parameters. The wedge formed by the two dashed lines corresponds to ± 2 standard errors of the mean, so for points falling outside of the wedge there is a significant expectation that the distribution of elementary effects is non-zero. Drawn with the function `e2e_plot_sens_mc()`.

Performance of the maximum likelihood fitted model

The data required to derive credible intervals around the maximum likelihood fitted model were generated by 1000 runs of the 1970-1999 model, each of 40 years. The parameters values for each run were drawn from symmetrical random uniform distributions around the maximum likelihood values with a bandwidth of $\pm 15\%$.

Annual average biomass density in the model (mMN.m^{-2}) varied by around 5 orders of magnitude across the guilds. Biomass density was higher in the offshore zone than the inshore for all guilds. In general, the uncertainty in annual average biomass increased with the trophic level of guilds (Figure 14).

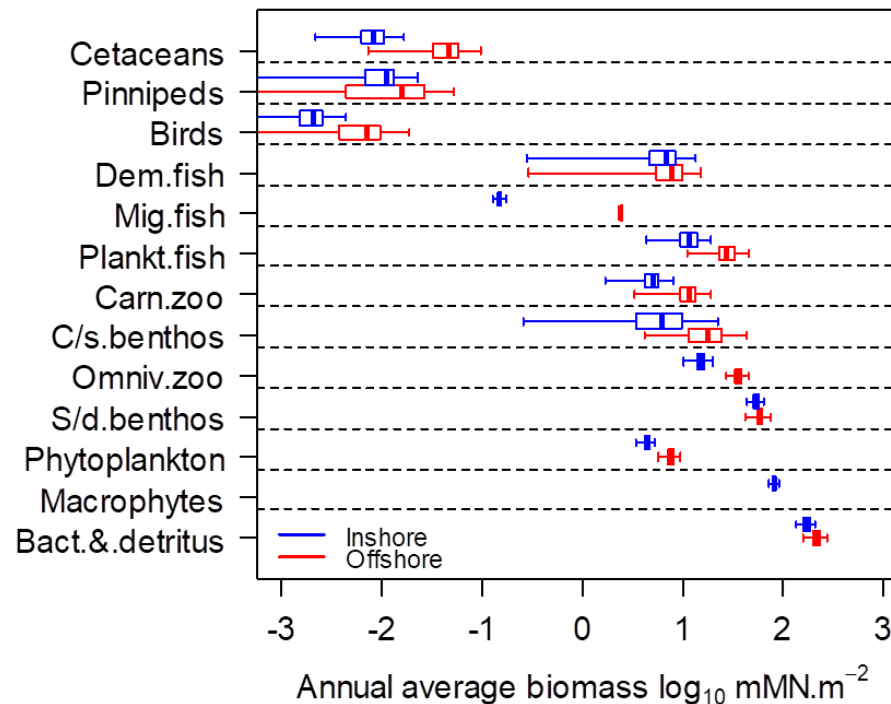


FIGURE 14 Credible intervals of the annual averaged values of living state variables in the model. X-axis shows \log_{10} values of the average biomass density (mMN m^{-2}) over a stationary annual cycle for the 1970-1999 fitted model. Blue box-and-whisker plots refer to the inshore/shallow zone, red to the offshore/deep zone. Whiskers span the 99% likelihood interval, boxes span 50%, and the central tick-mark indicates the median value. Drawn with the function `e2e_plot_biomass()`.

Stationary annual cycles of the each of the model state variables and their credible intervals, with the maximum likelihood fitted parameter set, fixed parameters, and driving data corresponding to the 1970-1999 period, are shown in Figures 15-22.

The annual cycles of the fish and the birds & mammals guilds in the model show the effects of the dynamic, food-motivated active migrations (Figure 23). The primary driver for the feeding migrations is the timing of seasonal peaks of omnivorous and carnivorous zooplankton concentrations, which higher in the inshore zone than offshore during summer, and conversely higher offshore in winter. These gradients drive an inshore movement of fish in the spring and an offshore movement in winter. This in turn drives an inshore spring immigration of especially birds, and offshore winter movement.

For some of the nutrient and plankton variables in the model we have corresponding monthly averaged observational data from various sources, aggregated up to the scale of the whole model domain (i.e. combining both the inshore and offshore sub-domains). Comparison of the observed and modelled monthly averaged data is shown in Figure 24. This represents an independent qualitative test of the model performance. The results show some under-prediction of the biomass of omnivorous zooplankton and meroplankton (larvae of benthic taxa). However the overall agreement is good, particularly with respect to the timing of peaks in abundance.

Data generated by the NetIndices package showed that the certainty of mean trophic level of the living state variables in the whole domain was relatively high (i.e. narrow credible intervals), except for the birds guild. However, the omnivory index was more uncertain for all guilds especially the birds (Figure 25).

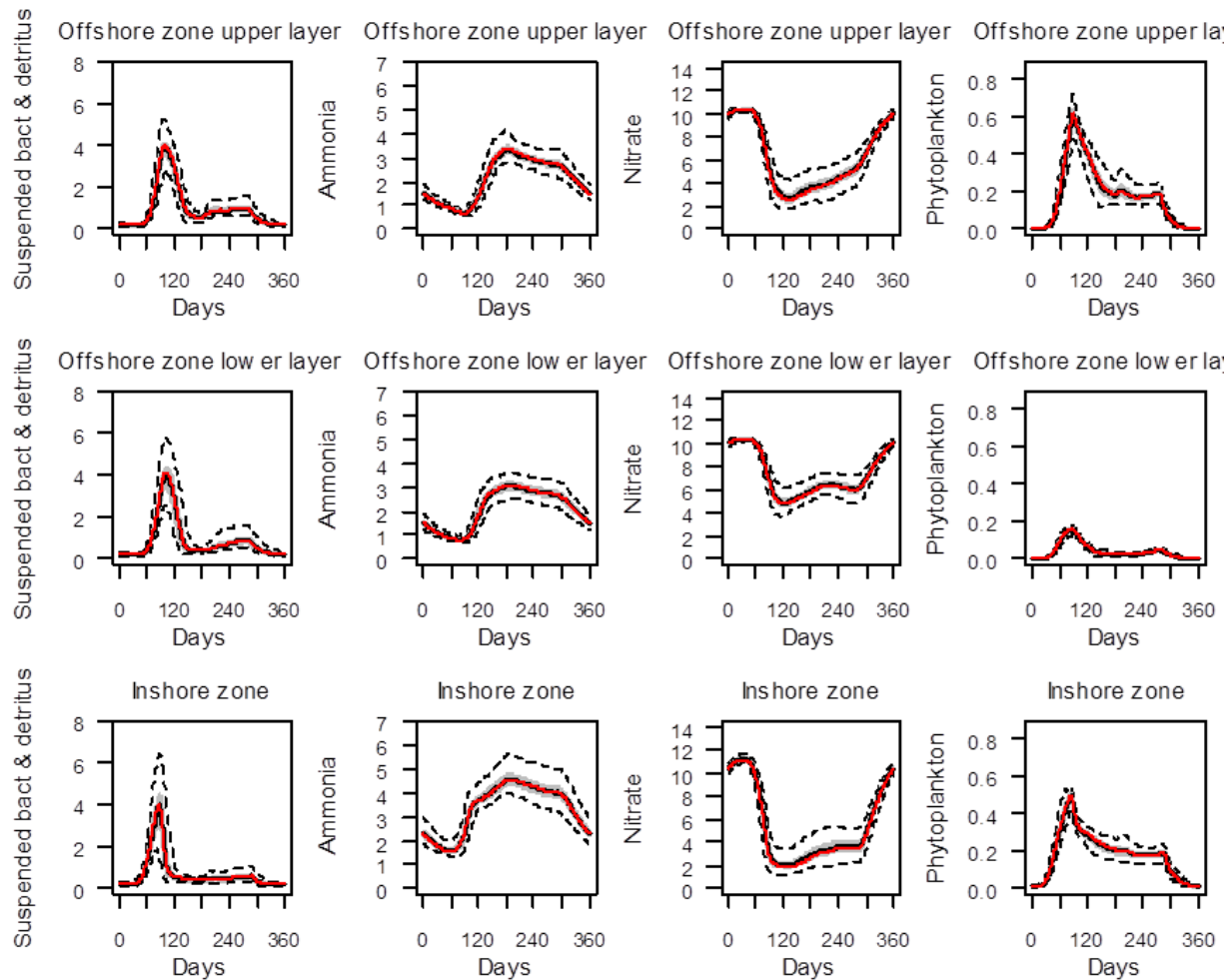


FIGURE 15 Stationary annual cycles of water column suspended detritus (implicitly including bacteria), dissolved nutrients, and phytoplankton for the 1970-1999 fitted model. Dashed lines span the 99% likelihood interval of model outputs given the observed indices of the state of the North Sea during this period. Grey shading spans the 50% likelihood interval. Solid black line is the median of the likelihood distribution and the red line is the maximum likelihood model. In this case the grey shading and black and red lines are almost coincident for all variables. Columns of panels are different variables output from the model, rows are different spatial compartments. Units for all variables: mMN m^{-3} .

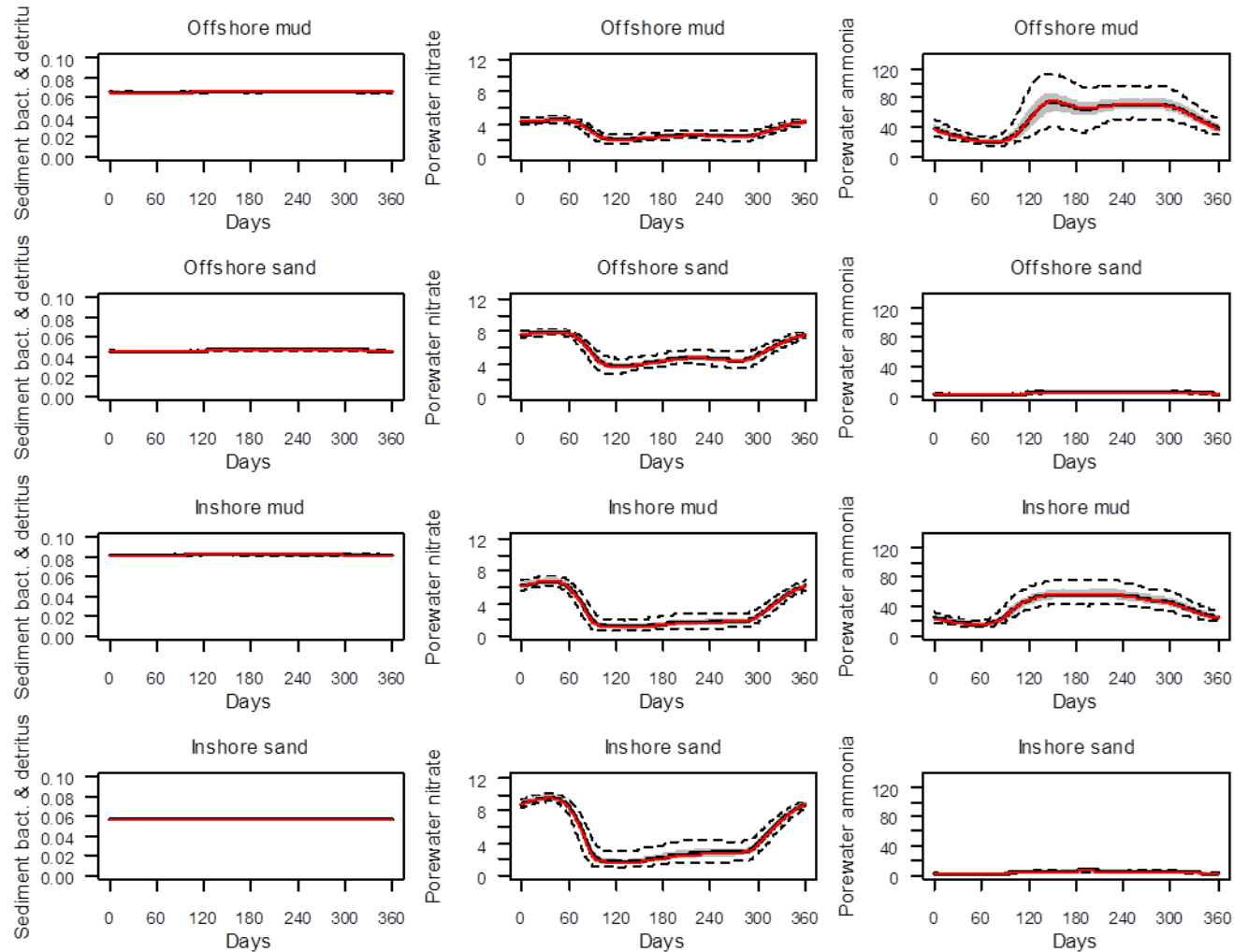


FIGURE 16 Stationary annual cycles of sediment detritus (implicitly including bacteria) and porewater nutrients for the 1970-1999 fitted model. Dashed lines span the 99% likelihood interval of model outputs given the observed indices of the state of the North Sea during this period. Grey shading spans the 50% likelihood interval. Solid black line is the median of the likelihood distribution and the red line is the maximum likelihood model. In this case the grey shading and black and red lines are almost coincident for all variables. Columns of panels are different variables output from the model, rows are different spatial compartments. Units: detritus %N by weight, dissolved nutrients mMN m^{-3} .

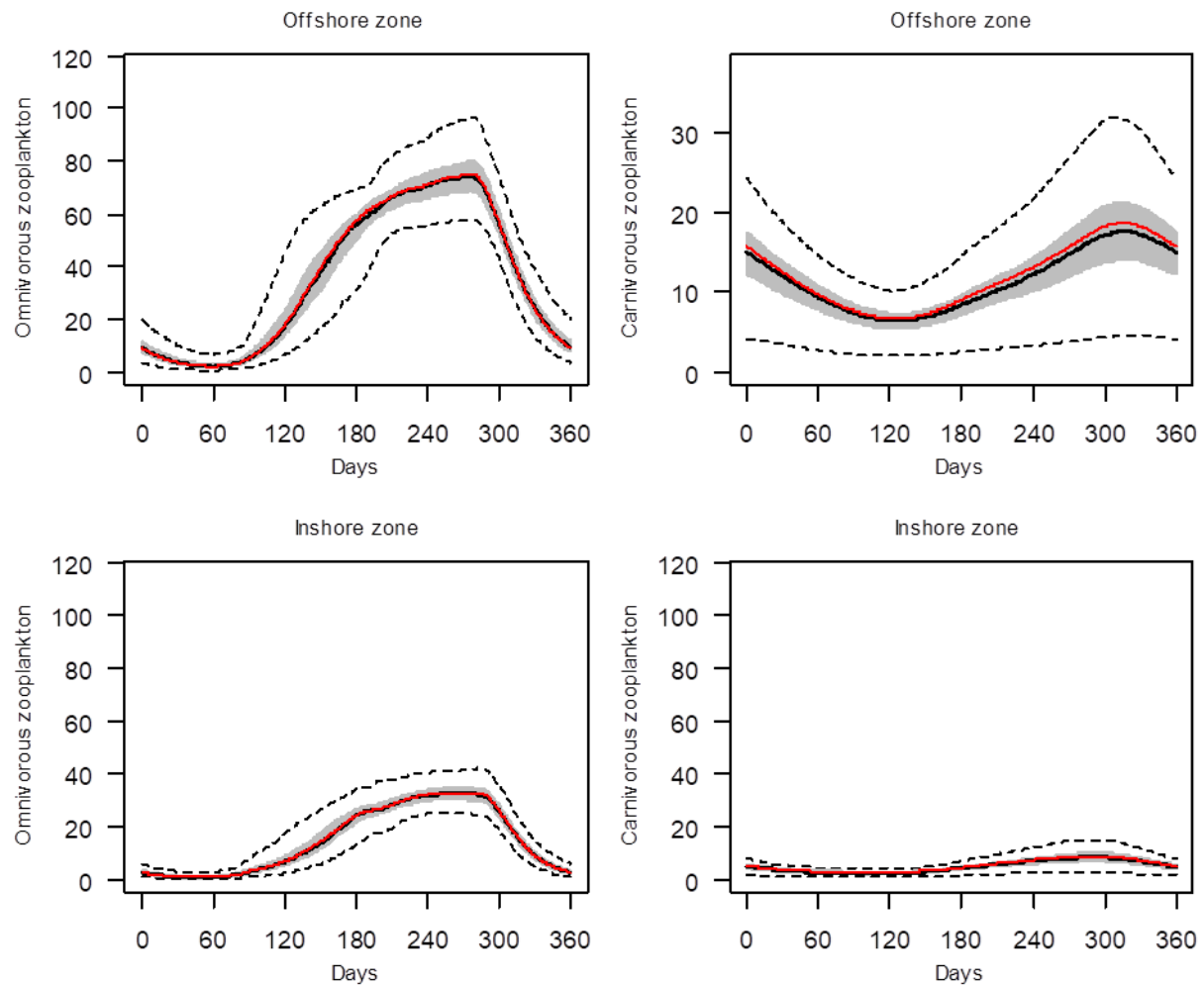


FIGURE 17 Stationary annual cycles of zooplankton guilds for the 1970-1999 fitted model. Dashed lines span the 99% likelihood interval of model outputs given the observed indices of the state of the North Sea during this period. Grey shading spans the 50% likelihood interval. Solid black line is the median of the likelihood distribution and the red line is the maximum likelihood model. Columns of panels are different variables output from the model, rows are different spatial compartments. Units for all variables: mMN m^{-2} .

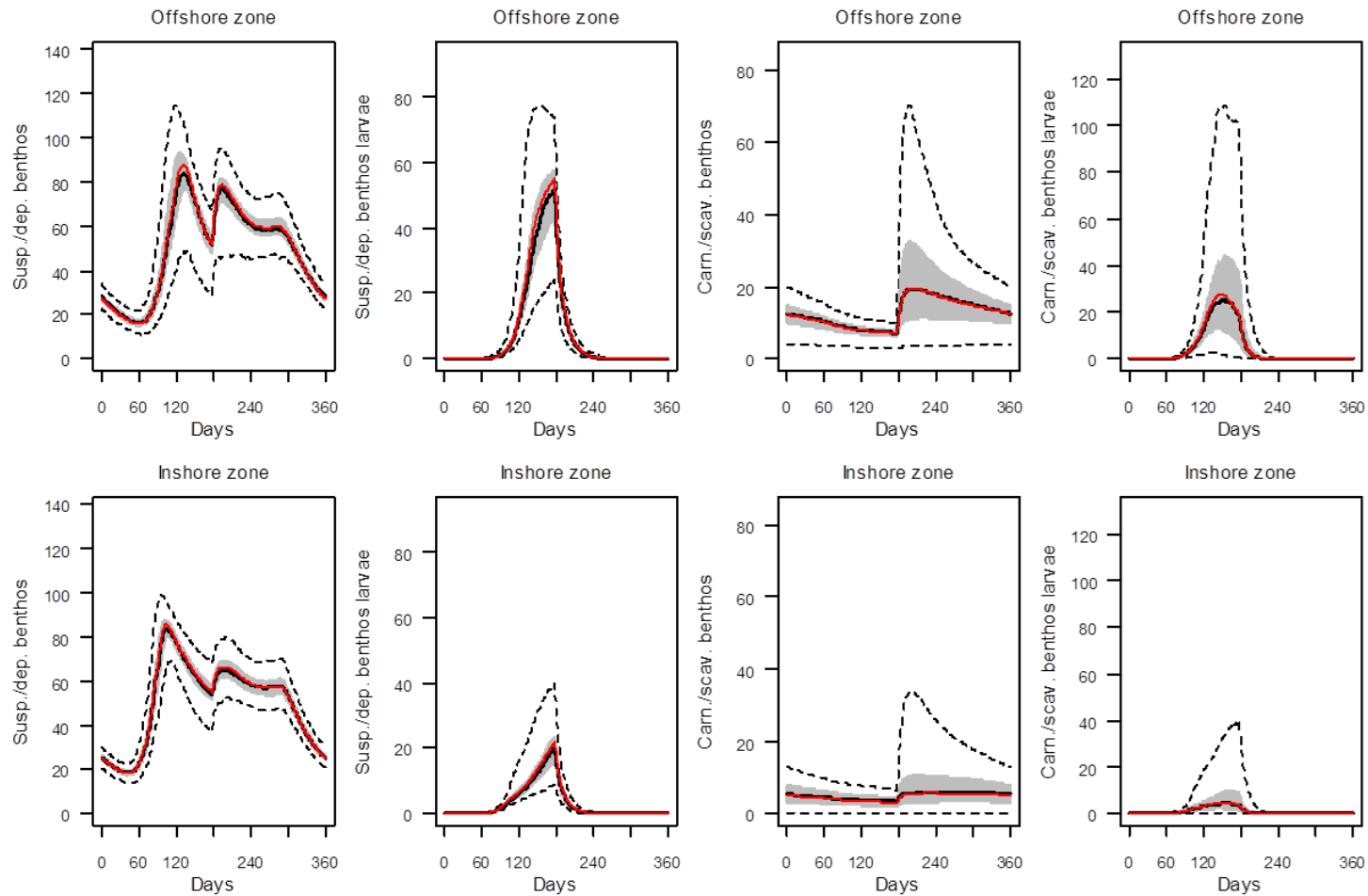


FIGURE 18 Stationary annual cycles of benthos guilds and their larval stages for the 1970-1999 fitted model. Dashed lines span the 99% likelihood interval of model outputs given the observed indices of the state of the North Sea during this period. Grey shading spans the 50% likelihood interval. Solid black line is the median of the likelihood distribution and the red line is the maximum likelihood model. Columns of panels are different variables output from the model, rows are different spatial compartments. Units for all variables: mMN m^{-2} .

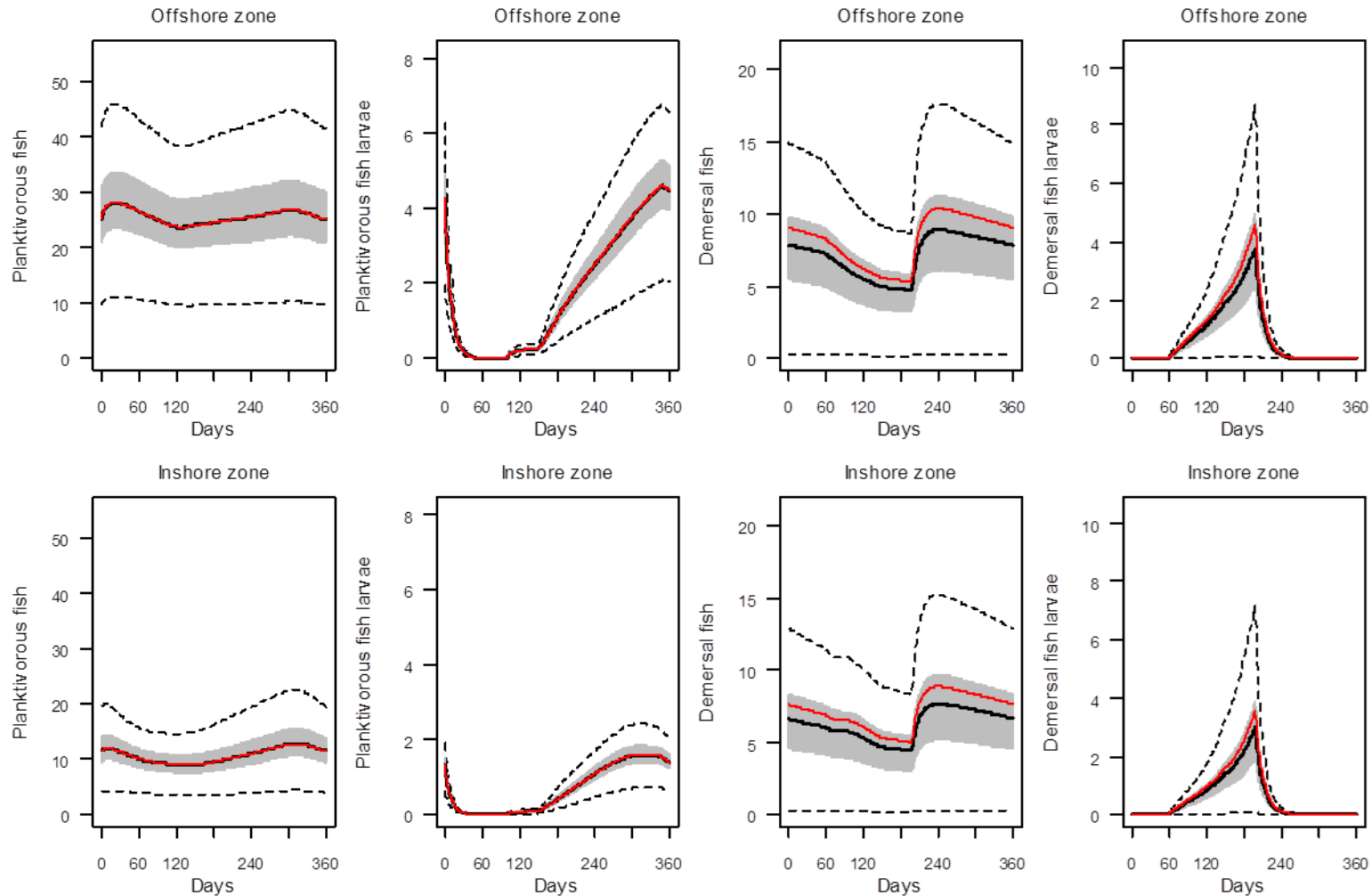


FIGURE 19 Stationary annual cycles of planktivorous and demersal fish guilds and their larval stages for the 1970-1999 fitted model. Dashed lines span the 99% likelihood interval of model outputs given the observed indices of the state of the North Sea during this period. Grey shading spans the 50% likelihood interval. Solid black line is the median of the likelihood distribution and the red line is the maximum likelihood model. Columns of panels are different variables output from the model, rows are different spatial compartments. Units for all variables: mMN m^{-2} .

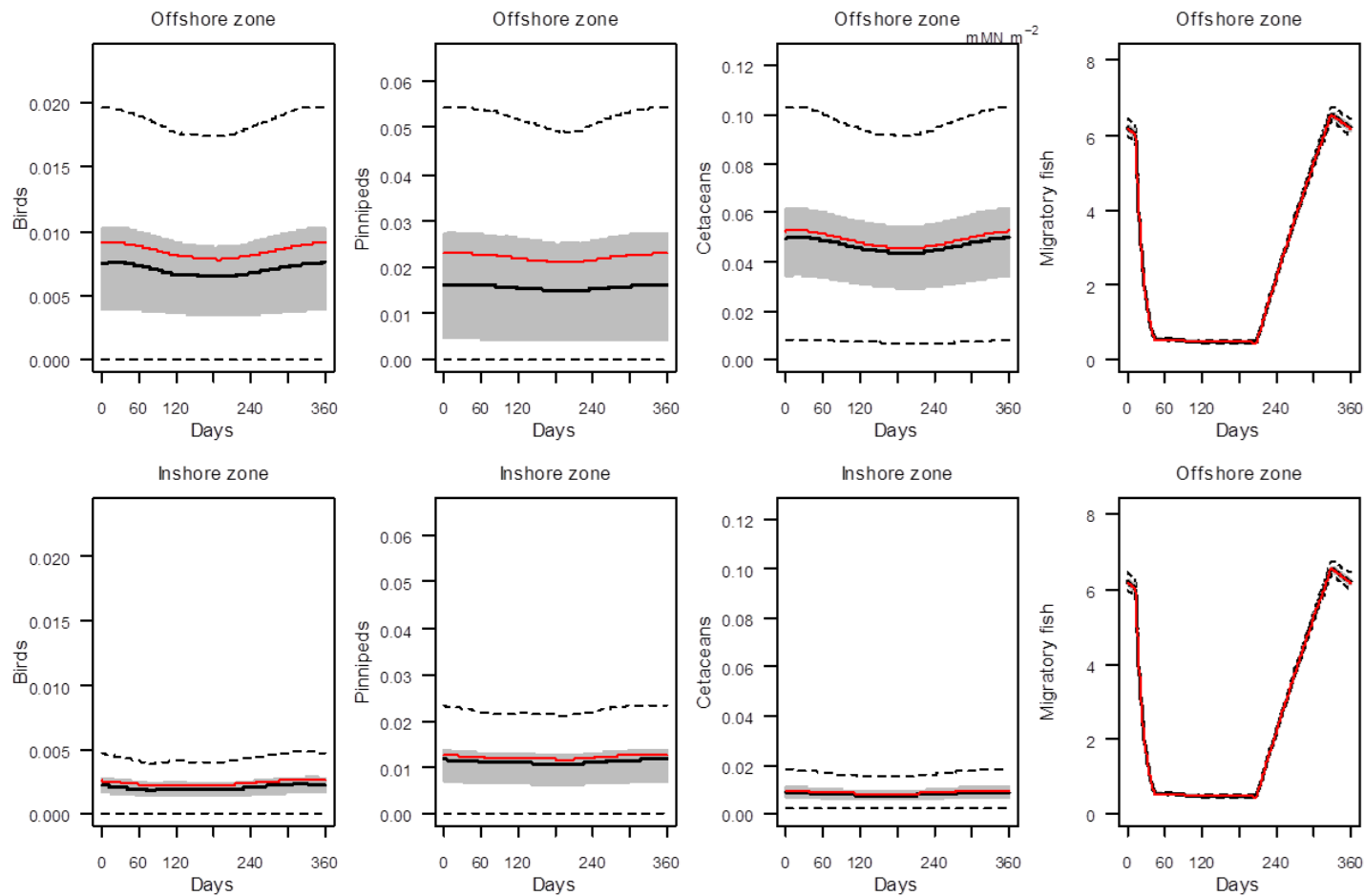


FIGURE 20 Stationary annual cycles of the birds & mammals guilds, migratory fish, and dead corpses for the 1970-1999 fitted model. Dashed lines span the 99% likelihood interval of model outputs given the observed indices of the state of the North Sea during this period. Grey shading spans the 50% likelihood interval. Solid black line is the median of the likelihood distribution and the red line is the maximum likelihood model. Columns of panels are different variables output from the model, rows are different spatial compartments. Units for all variables: mMN m^{-2} .

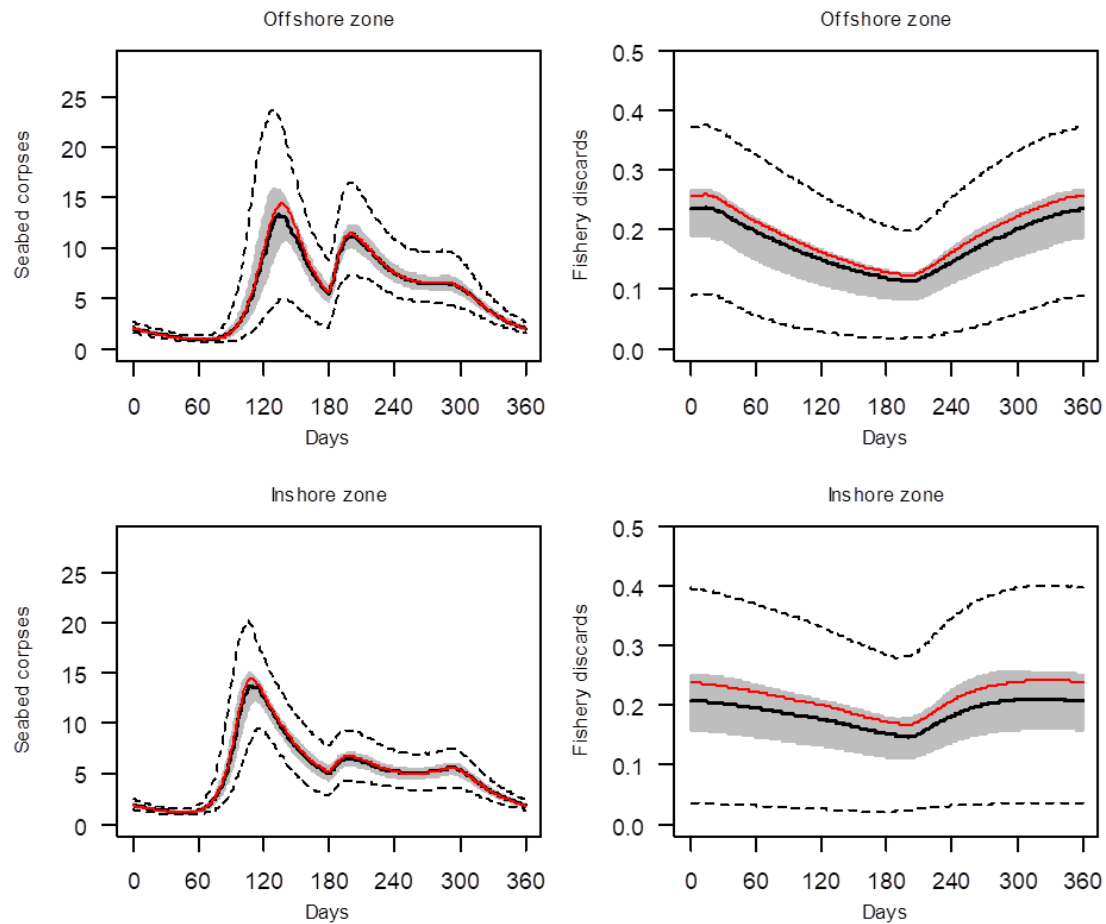


FIGURE 21 Stationary annual cycles of seabed corpses and fishery discards for the 1970-1999 fitted model. Dashed lines span the 99% likelihood interval of model outputs given the observed indices of the state of the North Sea during this period. Grey shading spans the 50% likelihood interval. Solid black line is the median of the likelihood distribution and the red line is the maximum likelihood model. Columns of panels are different variables output from the model, rows are different spatial compartments. Units for all variables: mMN m^{-2} .

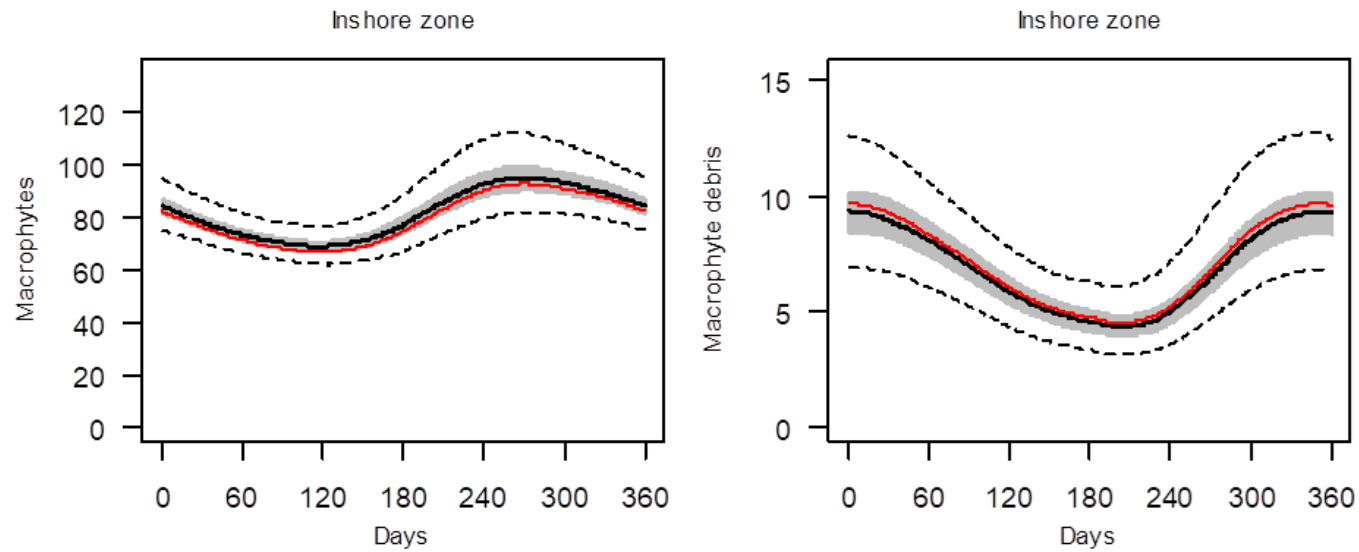


FIGURE 22 Stationary annual cycles of macrophytes and macrophyte debris in the inshore zone for the 1970-1999 fitted model. Dashed lines span the 99% likelihood interval of model outputs given the observed indices of the state of the North Sea during this period. Grey shading spans the 50% likelihood interval. Solid black line is the median of the likelihood distribution and the red line is the maximum likelihood model. Macrophytes are absent from the offshore zone. Units for all variables: mMN m^{-2} .

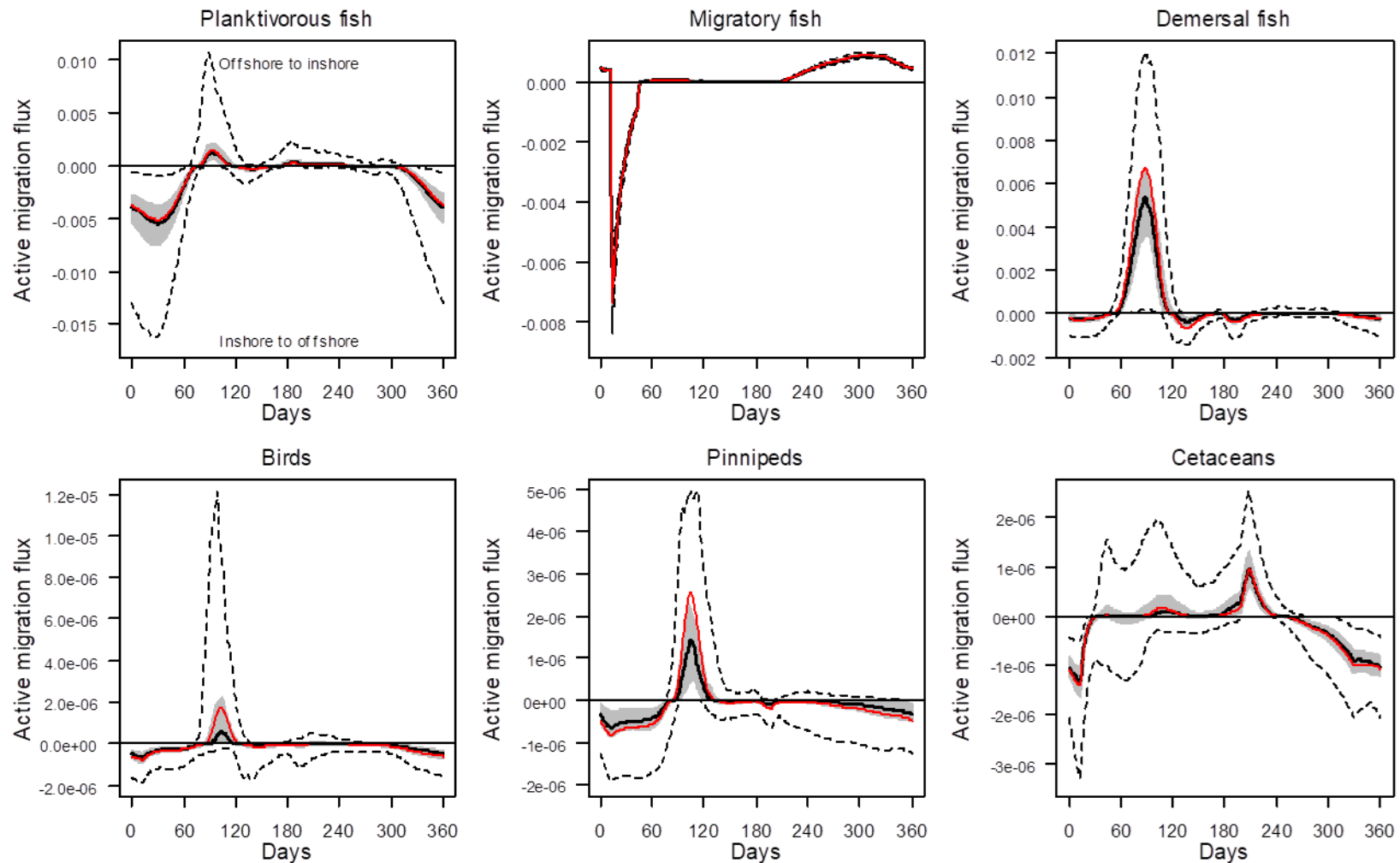


FIGURE 23 Stationary annual cycles of net active migration fluxes (mMN.m⁻².d⁻¹) of biomass between the offshore and inshore zones for the 1970-1999 fitted model. Positive values indicate net flux from offshore to inshore, and vice-versa for negative values. Dashed lines span the 99% likelihood interval of model outputs given the observed indices of the state of the North Sea during this period. Grey shading spans the 50% likelihood interval. Solid black line is the median of the likelihood distribution and the red line is the maximum likelihood model.

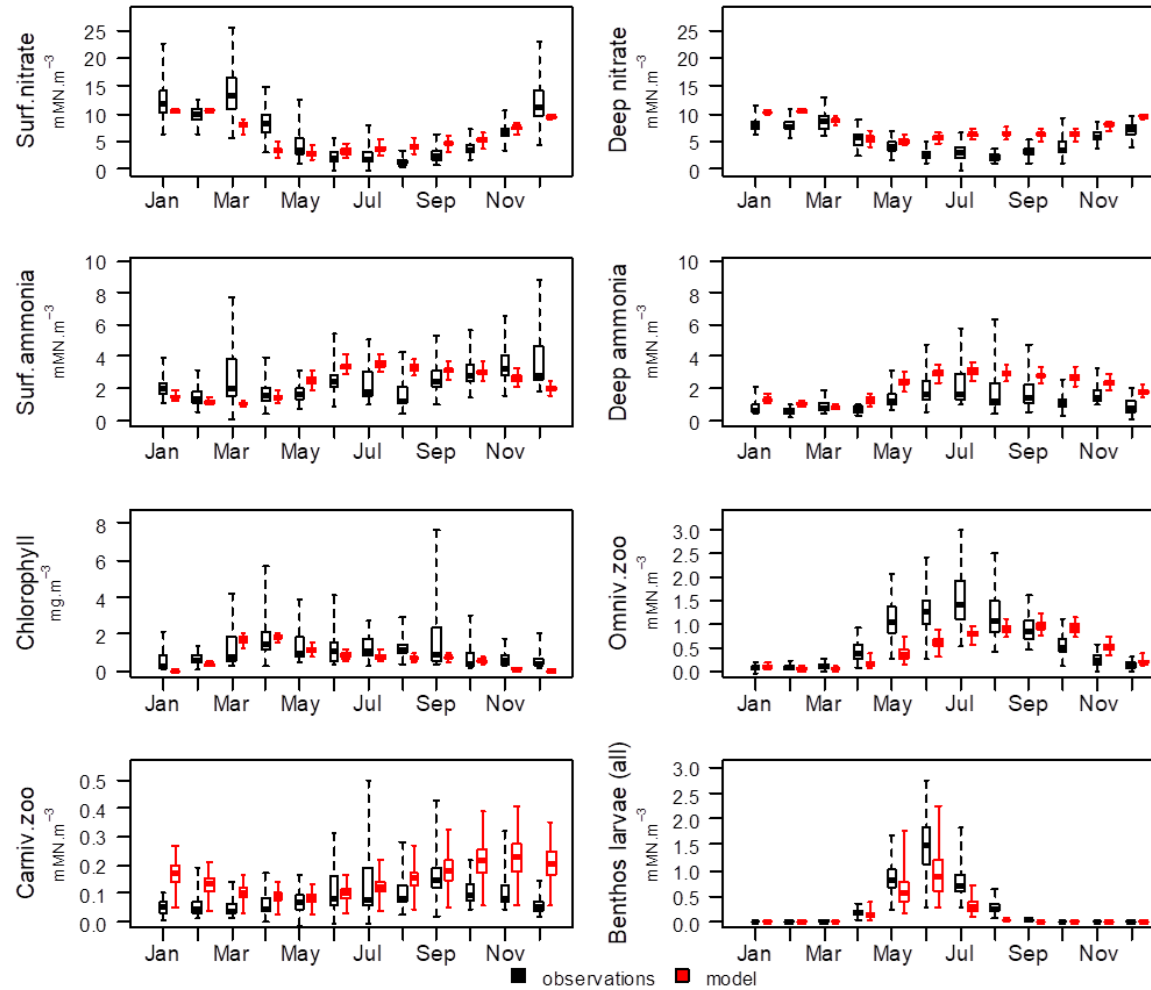


FIGURE 24 Monthly averages of the 1970-1999 stationary annual cycle daily resolution output for the whole model domain (red) and observed monthly averaged data from the North Sea (black). Box and whiskers for the model data show the 0.5, 25, 50, 75 and 99.5 centiles of the likelihood distribution of results given the uncertainty in fitted parameter values. For the observed data the box and whiskers show the equivalent variability in measurements from the North Sea aggregated over the period 1970-1999. Note that the model was not fitted to these observed data so the comparison represents a validation of the fitted model.

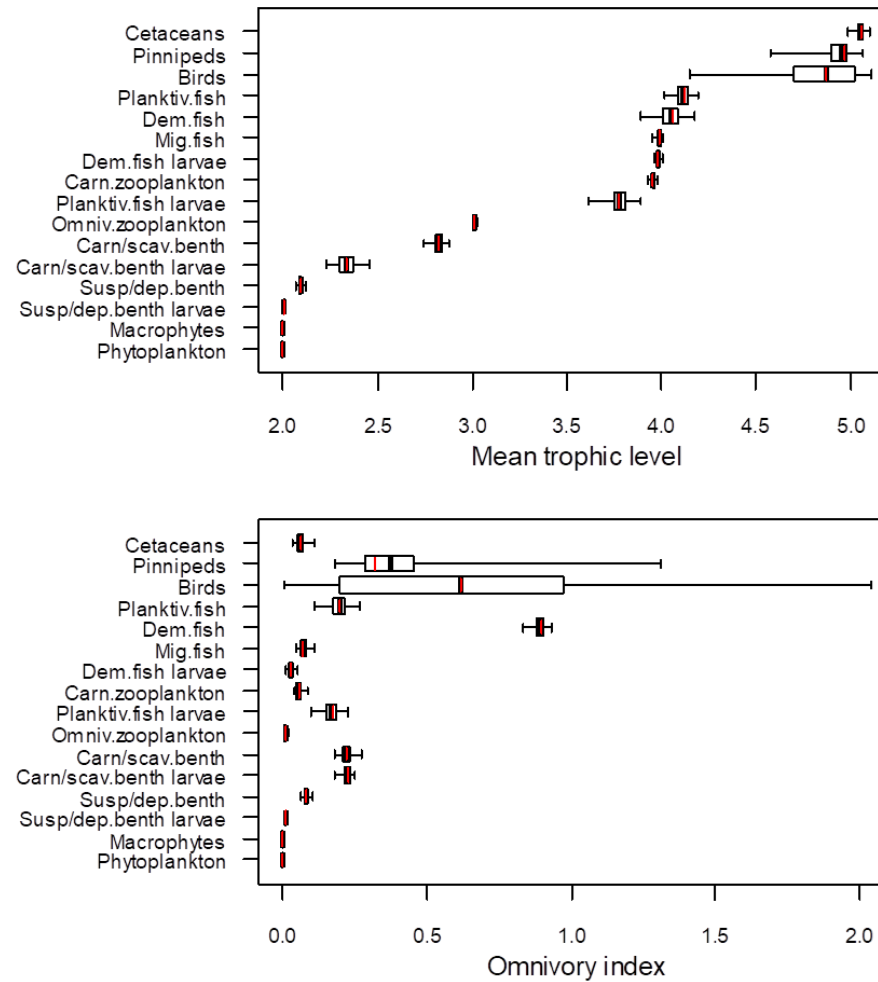


FIGURE 25 Credible intervals of the annual mean trophic level (upper panel) and omnivory index (lower panel) for the stationary 1970-1999 model. Black boxes span 50% of the likelihood interval, whiskers span 99%, thick black bar represents the median likelihood. The red bar in each case indicates the maximum likelihood model. Guilds (rows) in each panel are ranked by the mean trophic level in the maximum likelihood model.

Annual integrated mass fluxes in the maximum likelihood fitted model

At stationary state, 1970-1999 fishery landings represented 0.3 - 0.4% of annual gross primary production (GPP); sediment burial fluxes 15 - 17% of GPP, and denitrification was 130% of the combined atmospheric and riverine dissolved nutrient input (Figure 26). Overall, the model was a net importer of nitrogen from across the ocean boundaries, with exports due to advection and migrations equivalent to 91% of imports. A detailed breakdown of the mass balance fluxes is provided in Tables 40 and 41.

Burial of organic nitrogen in the seabed sediments emerges as a significant export flux from the model (17% of gross primary production in the inshore zone, 15% offshore) although the confidence intervals are wide (Tables 40 and 41). It is not at all clear whether this is a realistic figure or not. There seem to be few if any empirical estimates of nitrogen burial. Empirical estimates of carbon burial in the North Sea are also scarce and highly variable (de Haas *et al.*, 2002), but given the model assumption of constant Redfield stoichiometry (which may be particularly suspect in the context of sediment geochemistry) a 10% burial flux seems feasible though on the high-side of empirical evidence. Hence, the realism of the simulated burial fluxes remains an unresolved issue.

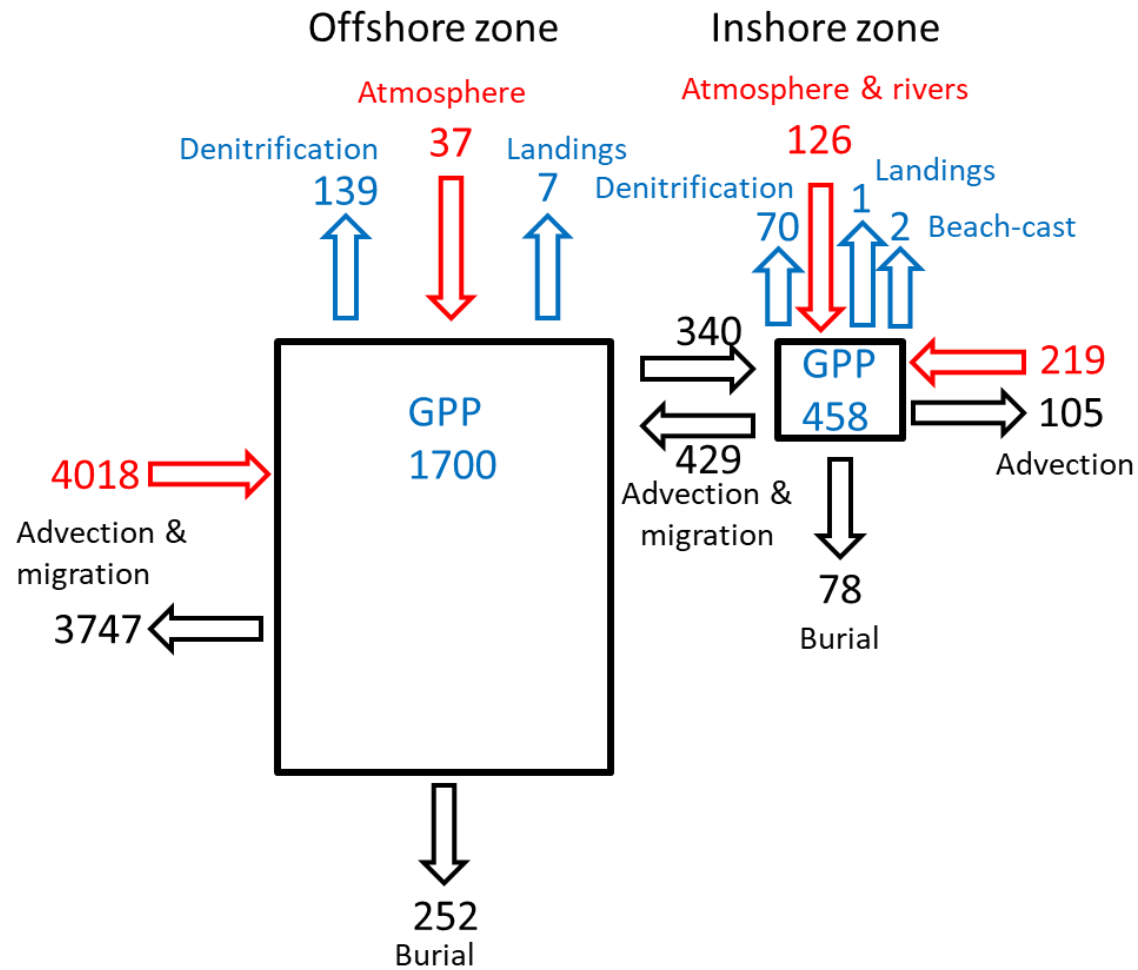


FIGURE 26 Stationary state nitrogen mass fluxes for the 1970-1999 model. Flux units: $\text{mMN}\cdot\text{y}^{-1}$ scaled to a model domain sea surface area of 1m^2 . Red arrows: fluxes defined by external driving data; blue arrows: modelled fluxes included in the target data set for fitting the model; black arrows: modelled fluxes not in the target data set. GPP indicates gross annual primary production – in the inshore zone this includes macrophytes.

TABLE 40 Stationary annual mass fluxes of nitrogen ($\text{mMN}\cdot\text{y}^{-1}$) into and out of the **offshore** zone of the 1970-1999 model domain (surface area 0.735 m^2). Figures in brackets are 99% credible intervals. DIN refers to dissolved inorganic nitrogen (nitrate + ammonia). In this case the fisheries landings are the processed weigh, not the live weight landed.

| | | Inputs | Outputs | | |
|---|-------------------------------|----------------------------------|----------------------------------|---------------------|-----------------------|
| | | Transport & migration | Transport & migration | Geochemistry | Fisheries |
| Ocean boundary | DIN | 3635.2 | 3350.0 (3220.9-3623.3) | | |
| | Plankton & detritus | 378.0 | 347.9 (223.7-504.3) | | |
| | Active migrations | 5.1 | 4.1 (3.9-4.3) | | |
| Inshore/offshore boundary (gross flux) | DIN | 356.0 (323.9-403.6) | 267.7 (247.8-292.8) | | |
| | Plankton & detritus | 72.6 (51.4-104.7) | 72.3 (52.9-92.8) | | |
| | Active migrations | 0.51 (0.15-1.64) | 0.35 (0.10-1.07) | | |
| Land and atmosphere | Atmospheric DIN deposition | 36.9 | | | |
| | River DIN discharges | 0 | | | |
| | Water column denitrification | | | 0.04 (0.04-0.05) | |
| | Sediment denitrification | | | 138.5 (117.1-164.3) | |
| | Macrophyte beach-cast | | 0 | | |
| Seabed sediments | Net burial | | | 251.9 (128.8-335.7) | |
| Human extraction | Planktivorous fish landings | | | | 4.84 (1.91-7.88) |
| | Demersal fish landings | | | | 1.16 (0.02-2.00) |
| | Migratory fish landings | | | | 0.77 (0.67-0.86) |
| | Susp/deposit benthos landings | | | | 0.060 (0.032-0.088) |
| | Carn/scav benthos landings | | | | 0.037 (0.008-0.082) |
| | Pelagic invert. Landings | | | | 0.003 (<0.001-0.017) |
| | Cetacean landings | | | | 4.04E-05 (0-14.6E-05) |
| TOTAL | | 4484.3 | 4086.9 | 390.5 | 6.87 |
| TOTAL GAS | | | | 138.6 | |
| TOTAL DIN IN/OUT | | 4028.1 | 3662.2 | | |
| TOTAL PON IN/OUT | | 456.2 | 424.7 | 251.9 | 6.87 |
| Phytoplankton gross production | | 1669.9.0 (1143.4-2167.1) | | | |

TABLE 41 Stationary annual mass fluxes of nitrogen ($\text{mMN}\cdot\text{y}^{-1}$) into and out of the **inshore** zone of the 1970-1999 model domain (surface area 0.265 m^2). Figures in brackets are 99% credible intervals. DIN refers to dissolved inorganic nitrogen (nitrate + ammonia). In this case the fisheries landings are the processed weigh, not the live weight landed.

| | | Inputs | Outputs | | |
|---|-------------------------------|----------------------------------|----------------------------------|---------------------|---------------------|
| | | Transport & migration | Transport & migration | Geochemistry | Fisheries |
| Ocean boundary | DIN | 147.2 | 89.7 (90.3-106.4) | | |
| | Plankton & detritus | 71.7 | 9.5 (5.7-14.9) | | |
| | Active migrations | 0 | 0 | | |
| Inshore/offshore boundary (gross flux) | DIN | 267.2 (242.1-294.5) | 356.0 (323.9-403.6) | | |
| | Plankton & detritus | 72.3 (52.9-92.8) | 72.6 (51.4-101.7) | | |
| | Active migrations | 0.35 (0.10-1.07) | 0.51 (0.15-1.64) | | |
| Land and atmosphere | Atmospheric DIN deposition | 18.2 | | | |
| | River DIN discharges | 107.5 | | | |
| | Water column denitrification | | | 0.005 (0.004-0.007) | |
| | Sediment denitrification | | | 70.3 (53.9-93.9) | |
| | Macrophyte beach-cast | | 1.59 (1.05-2.14) | | |
| Seabed sediments | Net burial | | | 77.6 (33.9-105.0) | |
| Human extraction | Planktivorous fish landings | | | | 0.80 (0.31-1.33) |
| | Demersal fish landings | | | | 0.52 (0.01-0.95) |
| | Migratory fish landings | | | | 0.009 (0.007-0.011) |
| | Susp/deposit benthos landings | | | | 0.034 (0.023-0.047) |
| | Carn/scav benthos landings | | | | 0.018 (0-0.076) |
| | Pelagic invert. Landings | | | | 0 (0.0-5.7E-5) |
| | Cetacean landings | | | | 0.0 |
| TOTAL | | 684.5 | 536.1 | 147.9 | 1.39 |
| TOTAL GAS | | | | 70.5 | |
| TOTAL DIN IN/OUT | | 540.1 | 451.9 | | |
| TOTAL PON IN/OUT | | 144.4 | 84.2 | 77.6 | 1.39 |
| Phytoplankton gross production | | 439.7 (320.6-532.0) | | | |
| Macrophyte gross production | | 18.6 (14.3-23.3) | | | |

Disaggregation of catch into landings and discards by each fishing gear

The raw output from the ecology model includes the total landings and discards of each guild from the inshore and offshore zones by the combined actions of all the fishing gears. Output from the fleet model is then used to disaggregate the annual integrated landings and discards between the gears (see model documentation). These results are illustrated in Figure 27 for a stationary year of the 1970-1999 model.

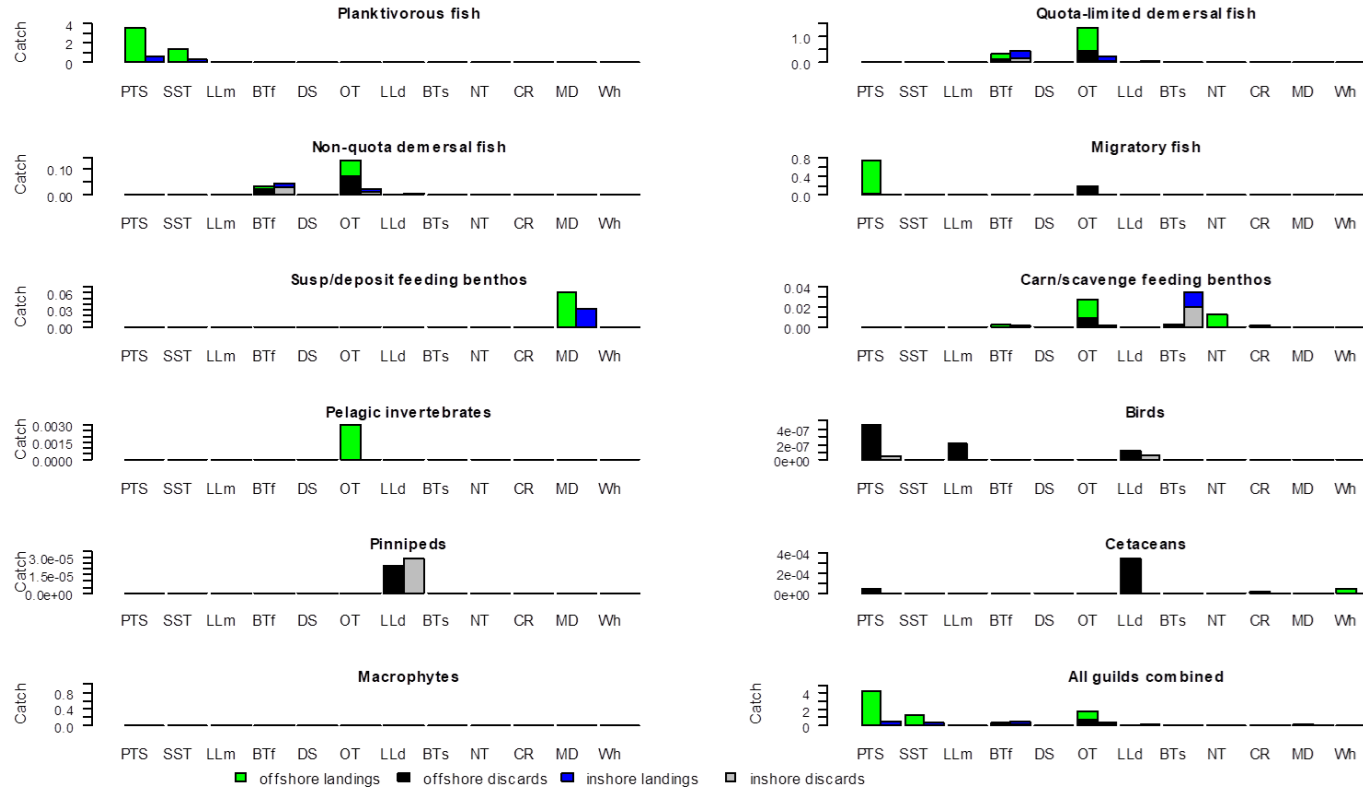


FIGURE 27 Distribution of 1970-1999 fishery catch across gears for each resource guild in the ecology model. Black and green bars represent discards and landings respectively from the offshore zone of the model. Grey and blue represent discards and landings from the inshore zone. The different fishing gear fleets are indicated along the x-axis by the codes: PT = Pelagic trawls and seines; SST = sandeel/sprat trawls; LLm = long-line for mackerel; BTf = fish beam trawl; DS = demersal seine; OT = demersal otter trawl; LLd = demersal long-lines and gill net; BTs = shrimp beam trawl; NT = Nephrops trawl; CR = creels; MD = mollusc dredges; Wh = Norwegian whaler.

Sensitivity to variations in pelagic and demersal harvesting at the scale of the whole model domain.

Sensitivity of the stationary state of the 1979-1999 model to bi-variate changes in pelagic and demersal fish harvest ratios was carried out by replicating the approach taken by Heath (2012). The model was run to a stationary state for each of 49 combinations of pelagic and demersal fish harvest ratios, forming a 7 x 7 matrix of values, with each harvest rate varying between 0 and 3-times the baseline 1970-1999 rate, in intervals of 0.5. Hence, the baseline model was represented by matrix cell coordinates 3,3 (pelagic and demersal harvest ratios 1.0 and 1.0-times the baseline respectively). The term 'pelagic' harvest ratio here refers to the harvest ratios of both planktivorous and migratory fish which were varied in synchrony.

Practically, the structured variations in harvest ratio were achieved by alterations to the harvest ratio multiplier values for the planktivorous, migratory and demersal fish guilds in the parameter file 'harvest_ratio_multiplier.csv'. This means that while the harvest ratios were varied, the activity rates of the fishing gears were not. So, other consequences of fishing such as seabed abrasion rates and harvest ratios on other resources guilds were unaffected. Essentially, the variations in pelagic and demersal harvesting were implicitly achieved by systematic changes in the selectivity patterns and catching power of each gear. Discard and processing at sea rates from all guilds except demersal fish were constant across all the model runs. In the case of demersal fish, discard rates were set to vary according the in-built density dependent relationship with demersal fish biomass within the model.

At the end of each run, the annual averaged biomasses of model components were calculated for the final year of the simulation, together with annual integrals of production rates, dietary fluxes, landings and discards. In addition, a range of network information indices were derived from the annual integrated flow matrix for the final year, using the R package NetIndices (Soetaert & Kones 2014).

On conclusion of all 49 model runs, the data on each individual model output (e.g. planktivorous fish landings) was assembled across all runs into a 7 x 7 matrix and visualised as contoured and colour-shaded maps (Figures 28-33)

The maps of planktivorous and demersal fish landings (Figure 28) both show a characteristic ridge of peak values running through the two-dimension parameter space defining pelagic and demersal harvest ratios. The 'height' of the crest of each ridge represents the maximum sustainable yield (MSY; in terms of landings) for each resource guild, which is a key criterion in fisheries management. The combination of harvest ratios defining the trajectory of MSY through the harvest ratio space (HR_{MSY}) represents the fishing conditions which deliver MSY – equivalent to the term F_{MSY} in the fisheries management context. These results are some of the most important outputs from the model – they show that MSY and HR_{MSY} for planktivorous and demersal fish are inter-independent. This interaction between the two fishing sectors arises as a result of the direct predator-prey relationship between the two guilds (demersal fish eat planktivorous fish and larvae; planktivorous fish eat demersal fish larvae), and also from indirect food web interactions via zooplankton and the predators on the fish guilds. The magnitude of MSY for planktivorous fish is highly sensitive to the demersal fish harvest rate (i.e. planktivorous MSY is strongly depressed at low demersal harvest ratio (high demersal fish biomass), and conversely high at high demersal harvest ratio (low demersal biomass)). On the other hand, demersal fish MSY is relatively insensitive to pelagic harvesting. The clear implication is that demersal fish exert a strong top-down effect on

planktivorous fish productivity, but planktivorous fish have only a weak bottom-up effect on demersal fish. These are key elements of guidance for the strategic management of fisheries.

For the demersal fish guild, the pattern of discard quantity across the pelagic and demersal harvest ratio space is not exactly proportional to the pattern of landings, because of the effects of the density-dependent discard rate which was implemented in the simulations. Demersal fish discard rate is indirectly related to demersal fish biomass in the model, mimicking the empirically observed response of average fish size to variations in stock biomass.

The landings map for migratory fish does not show the same distribution with respect to harvest ratios as the planktivorous fish despite the ratios for the two guilds being varied in concert, because the biomass of migratory fish in the system is sustained by external immigration. In the model, the immigration rate is independent of the intensity of harvesting within the model domain – the assumption is that harvesting within the North Sea is a minor component of the overall harvesting rate of the whole northeast Atlantic stock. Parameterising a feedback between the local harvest rate within the model domain and the magnitude of the external ocean stock of migratory fish, and potentially its migration pattern (i.e. the proportion entering the model each year) is a topic for future development.

The top-down effects of harvesting on the model food web are clearly visible in the maps of annual averaged masses of zooplankton, benthos, phytoplankton and nutrient within the two-dimensional harvest ratio space. Similarly, the bottom-up effects of harvesting on the annual average biomass of birds, pinnipeds and cetaceans (Figure 29). The response patterns are complex, but a clear feature is that the top-down effects generate only small variations in nutrient and plankton generate across the harvest-ratio space, whereas the bottom-up effects generate large variations on the top-predators with near-extinction in some parts of the space. Similarly, the top-down effects of harvesting caused only small variability in annual mean trophic level and omnivory indices of the zooplankton and benthos guilds, but the bottom-up effects on top-predator groups were considerably larger (Figure 34). The variations in omnivory indices were reflected in the emergent diet compositions of the top-predator groups which were mainly linked to the availability of planktivorous fish which are the main preferred food type of birds, pinnipeds and cetaceans (Figure 32).

The maps of network indices generated by the model simulations show clear patterns, but have not yet been subject to serious consideration (Figure 33). They are included here to illustrate the responses generated by the model. Possibly of significance are the maps of system ascendancy, and the ratio of ascendancy:capacity (AC ratio). Ascendancy is a measure of the degree of organisation of the network, or the extent to which the capacity is being utilised. The AC ratio has been proposed as a useful index of the 'maturity' of a network (Allesina & Ulanowicz, 2004). In these simulations, the AC ratio is clearly correlated with top-down driven variations in primary production, whilst the distribution of ascendancy appears to be loosely correlated with distribution of planktivorous fish landings.

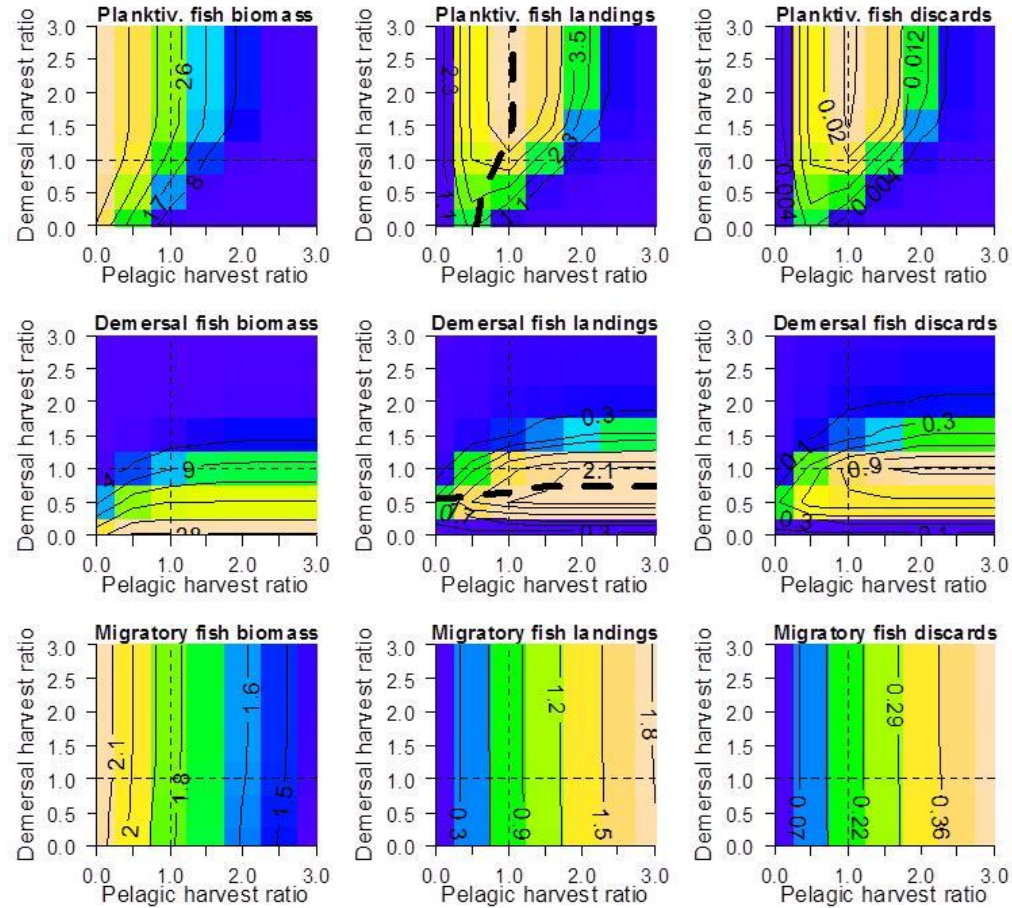


FIGURE 28 Variations in stationary annual mean biomass (mMN.m^{-2}), annual integrated landings ($\text{mMN.m}^{-2}.\text{y}^{-1}$), and annual integrated discards ($\text{mMN.m}^{-2}.\text{y}^{-1}$) for each guild of the three finfish in the 1970-1999 model, in relation to pelagic and demersal harvest ratios. The x and y-axis scales show the multipliers applied to the baseline 1970-1999 harvest ratios, so coordinates 1,1 (indicated by the thin dashed 'cross-wires') correspond to the baseline model. Harvest ratios for migratory and planktivorous fish were varied in simultaneously. The heavy black dashed line in the landings panel for planktivorous fish and demersal indicates the trajectory of harvest ratios generating maximum sustainable landings.

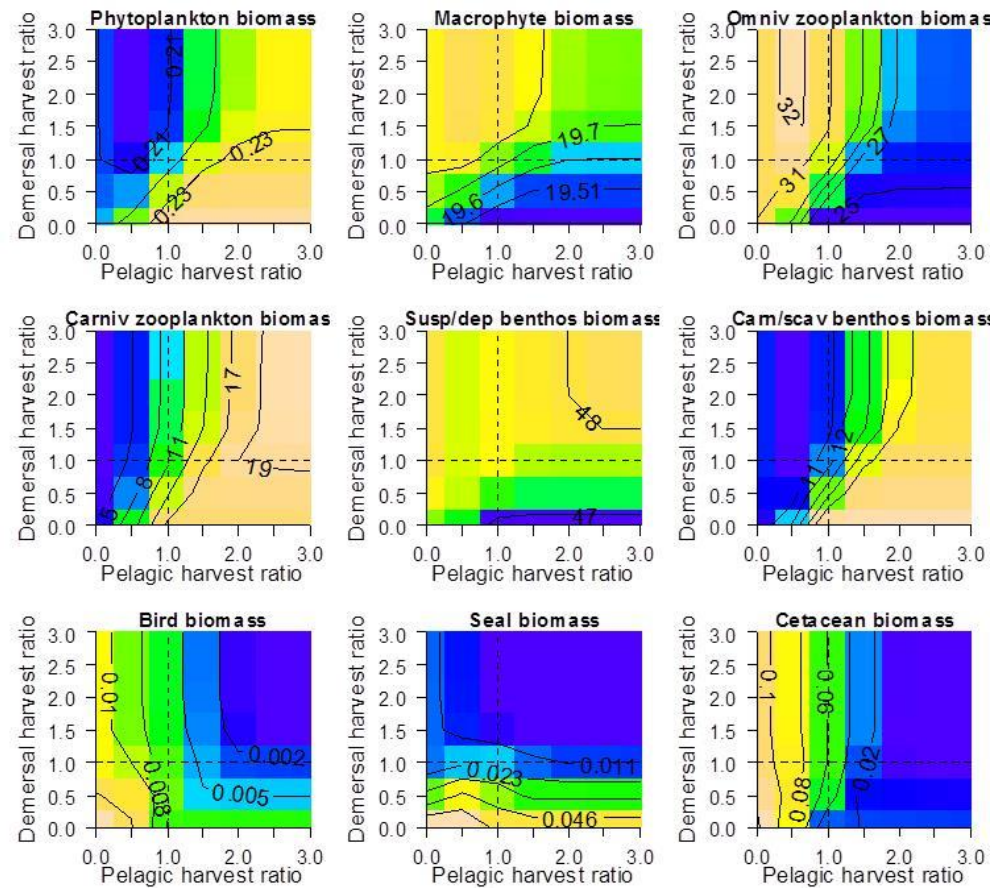


FIGURE 29 Variations in stationary annual mean biomass ($\text{mMN}\cdot\text{m}^{-2}$) of each living guild other than finfish in the 1970-1999 model, in relation to pelagic and demersal harvest ratios. The x and y-axis scales show the multipliers applied to the baseline 1970-1999 harvest ratios, so coordinates 1,1 (indicated by the thin dashed 'cross-wires') correspond to the baseline model. Harvest ratios for migratory and planktivorous fish were varied in simultaneously.

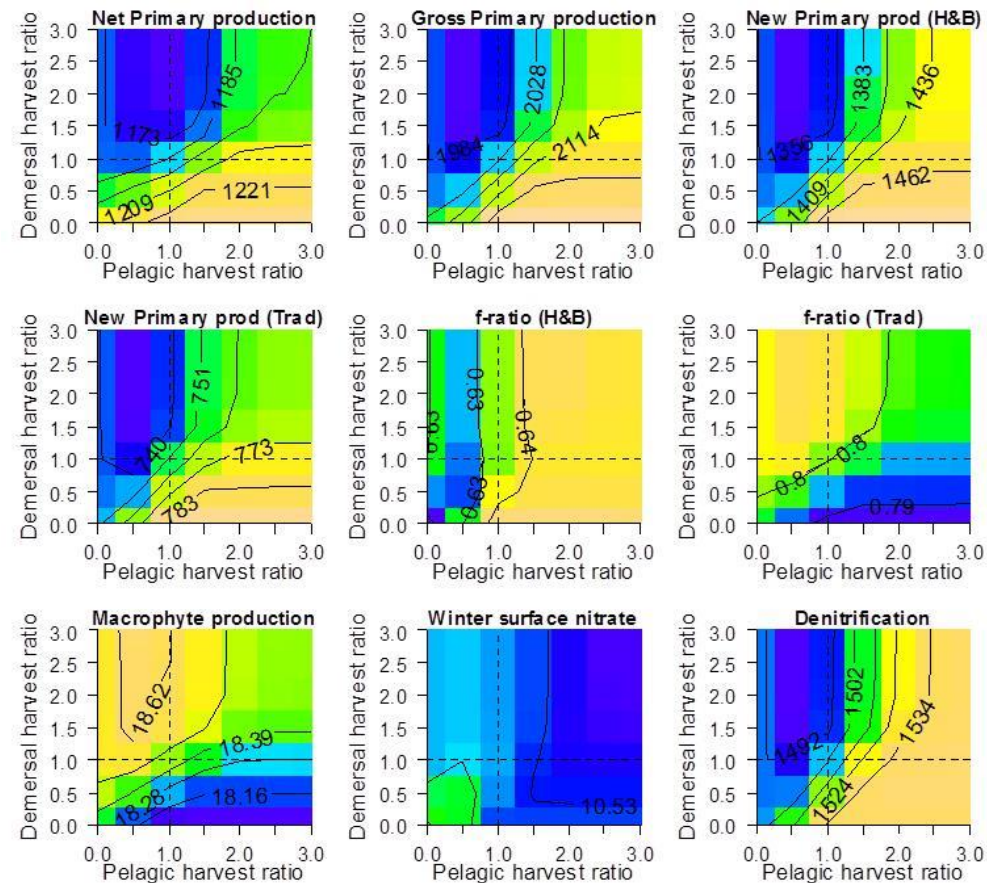


FIGURE 30 Variations in annual integrated primary production and denitrification fluxes ($\text{mMN}\cdot\text{m}^{-2}\cdot\text{y}^{-1}$), in the 1970-1999 model, in relation to pelagic and demersal harvest ratios. The x and y-axis scales show the multipliers applied to the baseline 1970-1999 harvest ratios, so coordinates 1,1 (indicated by the thin dashed 'cross-wires') correspond to the baseline model. Harvest ratios for migratory and planktivorous fish were varied in simultaneously. Net primary production is the net of gross production (annual integrated nutrient assimilation) and annual integrated non-grazing mortality and metabolic losses. Two versions of new primary production are shown (traditional: annual integrated nitrate assimilation, and H&B: annual draw-down of depth-integrated nitrate between winter and summer (Heath and Beare 2008)). The f-ratio is the ratio of new production to gross primary production.

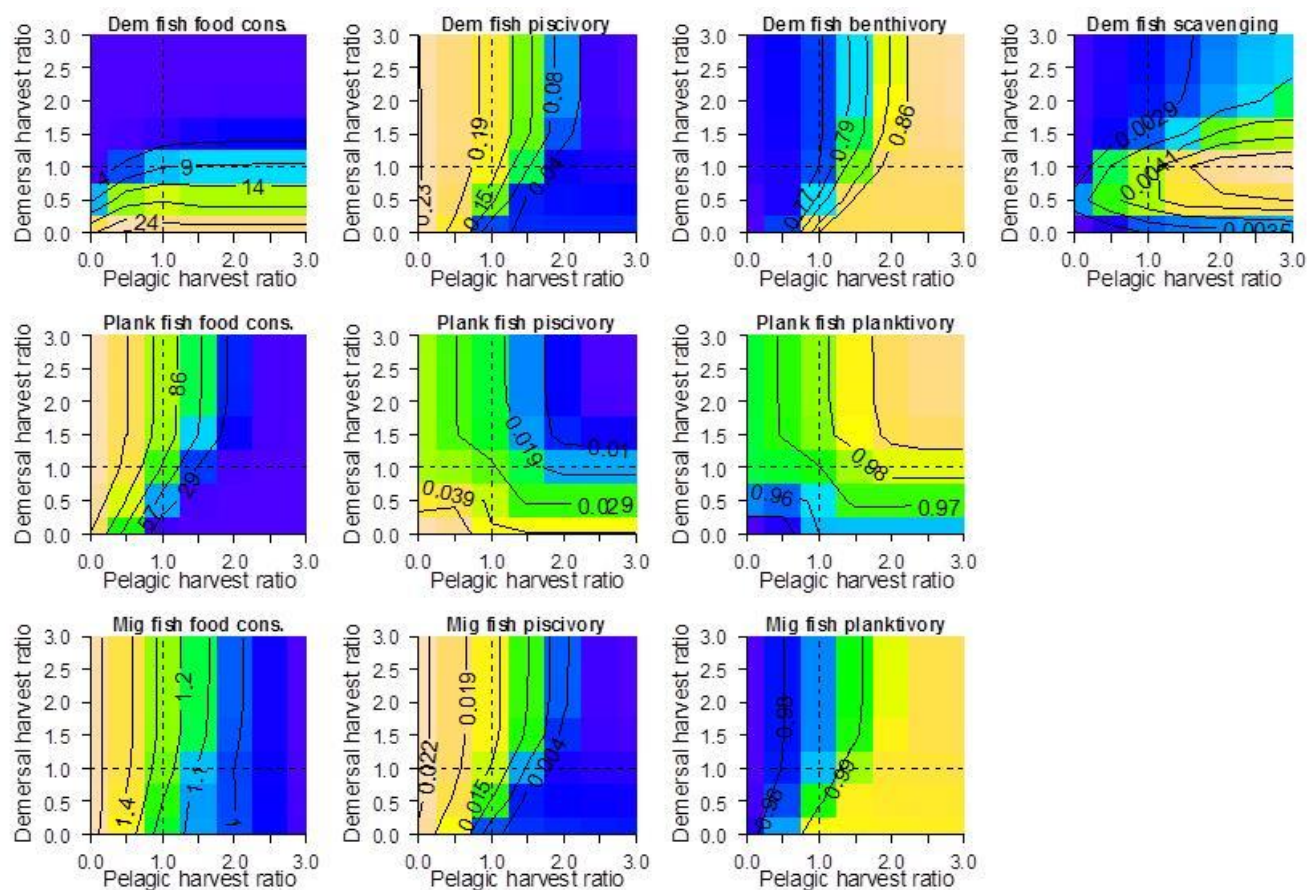


FIGURE 31 Variations in annual integrated food consumption and diet composition for each of the three finfish guilds in the 1970-1999 model, in relation to pelagic and demersal harvest ratios. The x and y-axis scales show the multipliers applied to the baseline 1970-1999 harvest ratios, so coordinates 1,1 (indicated by the thin dashed 'cross-wires') correspond to the baseline model. Harvest ratios for migratory and planktivorous fish were varied simultaneously. Left-hand column; food consumption ($\text{mMN}\cdot\text{m}^{-2}\cdot\text{y}^{-1}$), Second column: proportion of annual integrated food consumption made up of fish (piscivory). Third column: proportion of annual integrated food consumption made up of benthos or zooplankton (benthivory or planktivory). Final column (demersal fish only): proportion of annual integrated food consumption made up of corpses and discards (scavenging).

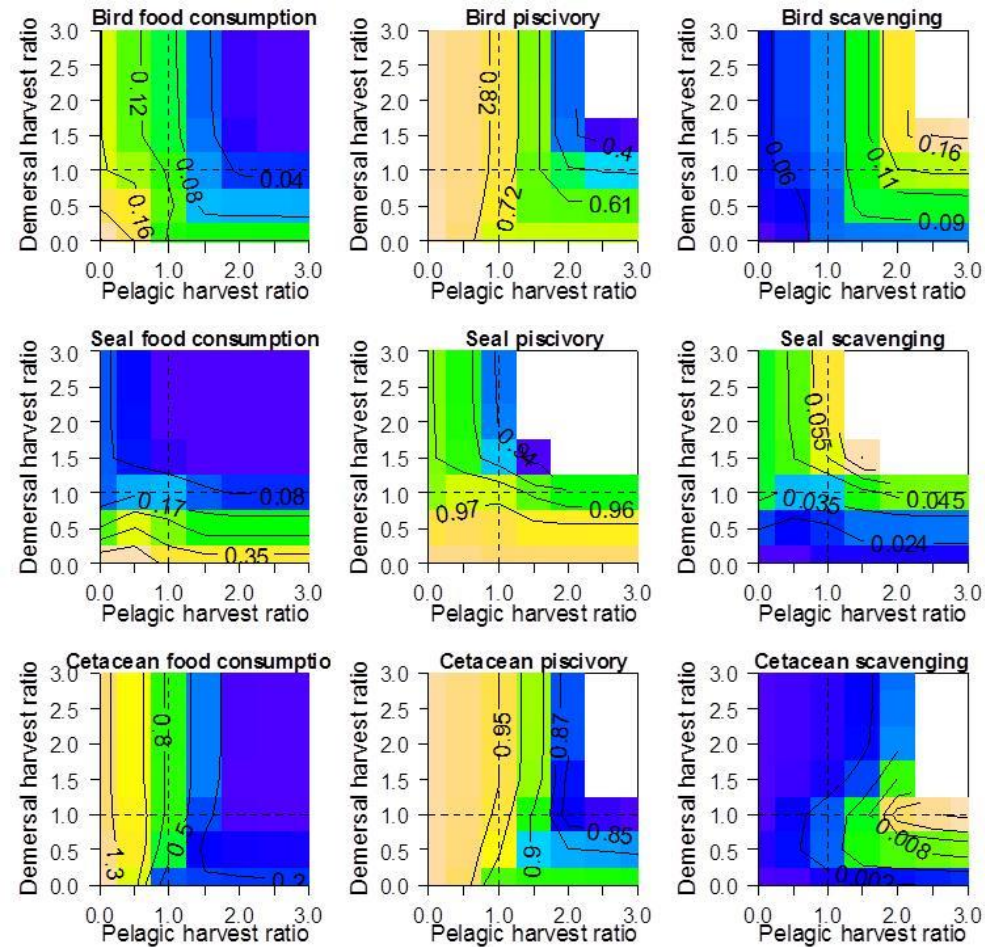


FIGURE 32 Variations in annual integrated food consumption and diet composition for each of the three top-predator guilds in the 1970-1999 model, in relation to pelagic and demersal harvest ratios. The x and y-axis scales show the multipliers applied to the baseline 1970-1999 harvest ratios, so coordinates 1,1 (indicated by the thin dashed 'cross-wires') correspond to the baseline model. Harvest ratios for migratory and planktivorous fish were varied in simultaneously. Left-hand column; food consumption ($\text{mMN}\cdot\text{m}^{-2}\cdot\text{y}^{-1}$), Second column: proportion of annual integrated food consumption made up of fish (piscivory). Third column: proportion of annual integrated food consumption made up of corpses and discards (scavenging). White areas indicate harvest ratio combinations where the diet proportion was undefined due the consumer mass being extremely small or zero.

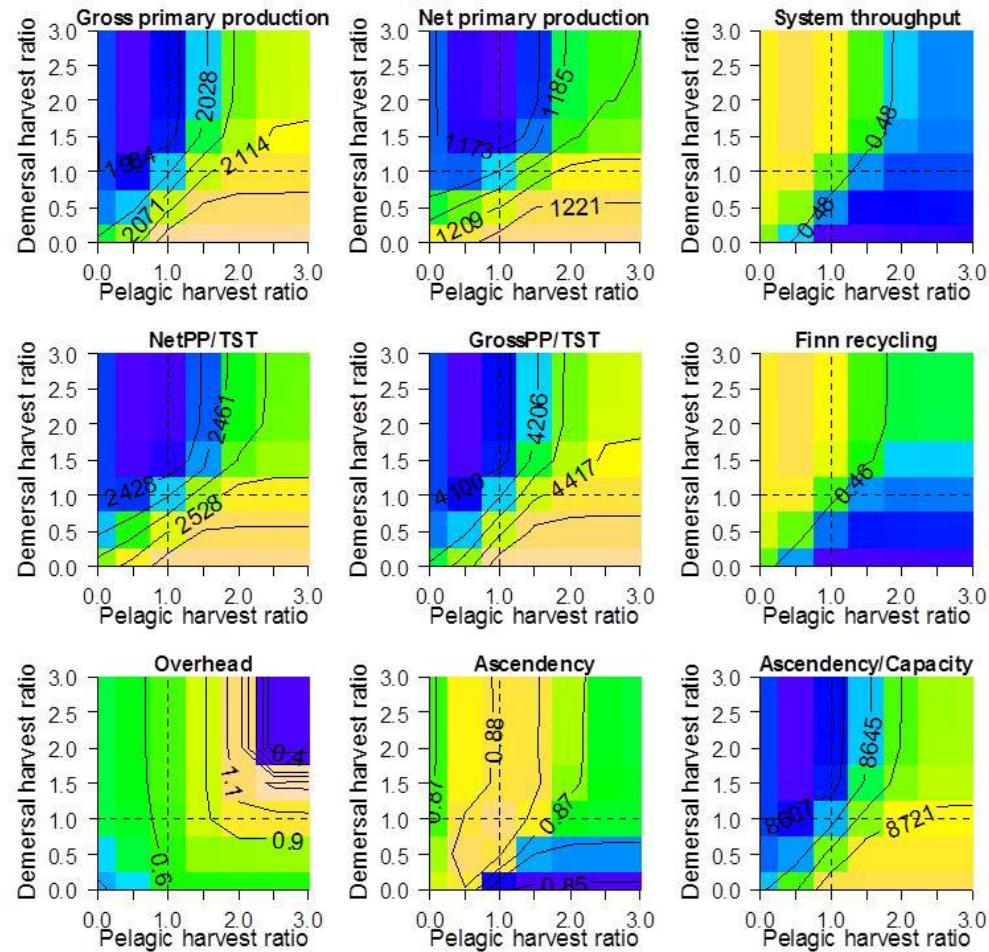


FIGURE 33 Variations in annual network information indices generated by the NetIndices package for the 1970-1999 model, in relation to pelagic and demersal harvest ratios. The x and y-axis scales show the multipliers applied to the baseline 1970-1999 harvest ratios, so coordinates 1,1 (indicated by the thin dashed 'cross-wires') correspond to the baseline model. Harvest ratios for migratory and planktivorous fish were varied in simultaneously.

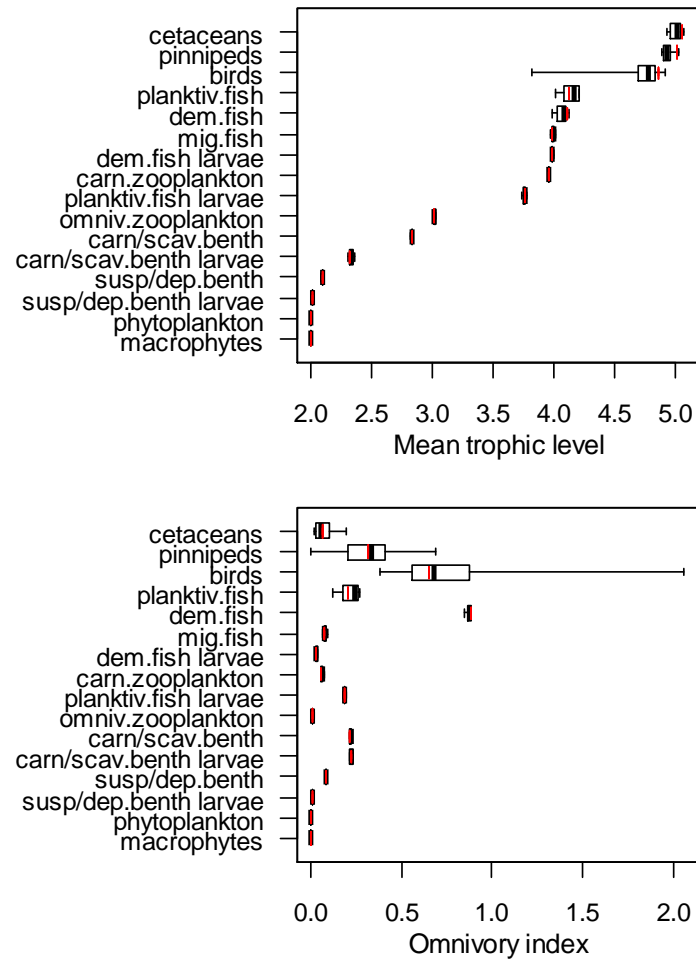


FIGURE 34 Variations in the stationary state annual mean trophic level (upper panel) and omnivory index (lower panel) of the living guilds within the bi-dimensional space of pelagic and demersal harvest ratio for the 1970-1999 model. Each harvest ratio was varied by factors of 0 to 3-times the 1970-1999 baseline value in intervals of 0.5-times. Black boxes span 50% of the distribution of values, whiskers the span full range, thick black bar represents the median value. The red bar in each case indicates the 1970-1999 baseline model. Guilds (rows) in each panel are ranked by the mean trophic level in the baseline model.

Acknowledgements

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